

Journal Pre-proof

Inter-Island Variability in Trace Elements and Trophic Ecology of Brown Booby (*Sula leucogaster*) in the South Atlantic

Padilha J.A.G., Almeida A.P., Souza-Kasprzyk J., Silva M., Cunha L.S.T., Soares T.A., Paiva T.C., Bighetti G.P., Torres J.P.M., Lepoint G., Michel L.N., Das K., Dorneles P.R.

PII: S0269-7491(24)02324-8

DOI: <https://doi.org/10.1016/j.envpol.2024.125607>

Reference: ENPO 125607

To appear in: *Environmental Pollution*

Received Date: 22 July 2024

Revised Date: 29 August 2024

Accepted Date: 27 December 2024

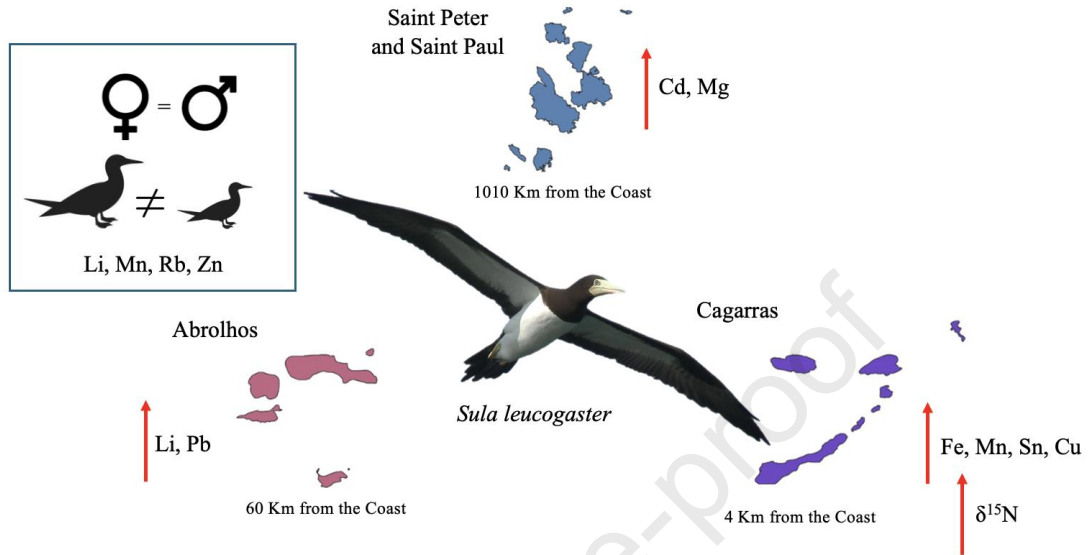
Please cite this article as: Padilha, J.A.G, Almeida, A.P, Souza-Kasprzyk, J, Silva, M, Cunha, L.S.T, Soares, T.A, Paiva, T.C, Bighetti, G.P, Torres, J.P.M, Lepoint, G, Michel, L.N, Das, K, Dorneles, P.R, Inter-Island Variability in Trace Elements and Trophic Ecology of Brown Booby (*Sula leucogaster*) in the South Atlantic , *Environmental Pollution*, <https://doi.org/10.1016/j.envpol.2024.125607> .

This is a PDF file of an article that has undergone enhancements after acceptance, such as the addition of a cover page and metadata, and formatting for readability, but it is not yet the definitive version of record. This version will undergo additional copyediting, typesetting and review before it is published in its final form, but we are providing this version to give early visibility of the article. Please note that, during the production process, errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2024 Published by Elsevier Ltd.



Graphical Abstract



34 **Highlights**

- 35 ● Inter-island differences in elemental levels in Brown Booby feathers
- 36 ● Higher contaminant levels in Cagarras, likely due to urban proximity.
- 37 ● Stable isotope analyses reveal distinct dietary patterns
- 38 ● No significant sex-based differences in metal concentrations.
- 39 ● Juveniles have different contaminant profiles than adults.

40

41

42 **Abstract**

43 This study investigates essential (Mg, Ca, Fe, Mn, Cu, Zn, Se, Ni) and non-essential (Li,
44 Be, Cr, Rb, Sr, Cs, Cd, Sn, Ba, and Pb) element concentrations and stable isotope ($\delta^{13}\text{C}$, $\delta^{15}\text{N}$,
45 $\delta^{34}\text{S}$) compositions in feathers of Brown Boobies (*Sula leucogaster*) from three distinct Atlantic
46 islands: the Archipelagos of Saint Peter and Saint Paul (SPSP), Abrolhos, and Cagarras. We aimed
47 to investigate the ecological and environmental factors influencing these seabird populations and
48 assess potential variations in contaminant exposure and dietary habits based on location, sex, and
49 maturity stages. Our finding revealed significant geographical differences in trace element
50 concentrations. The Brown Boobies from Cagarras had higher concentrations (mean \pm SD, $\mu\text{g g}^{-1}$)
51 of Fe (29 ± 20) and Mn (0.82 ± 0.82) than those from Abrolhos (Fe: 21 ± 20 ; Mn: 0.24 ± 0.09)
52 and SPSP (Fe: 15 ± 16 ; Mn: 0.21 ± 0.06). Tin concentrations were also higher in Cagarras (Sn:
53 0.02 ± 0.01) than in SPSP (Sn: 0.01 ± 0.01). Our analyses revealed significant differences in Li,
54 Mg, Rb, and Zn concentrations between adults and juveniles. However, there were no sex-related
55 differences in element concentrations within each locality. SIBER analyses revealed distinct
56 dietary differences among the three Brown Boobies populations, with the Cagarras seabirds
57 occupying a higher trophic position compared to the SPSP population. This study highlights the
58 importance of considering different populations to understand contaminant exposure and
59 ecological dynamics in Brown Boobies along the South Atlantic. The Cagarras population shows
60 significantly higher contaminant levels, likely due to proximity to anthropogenic activities. These

61 results highlight the necessity for ongoing monitoring to evaluate long-term effects on the more
62 impacted population and to ensure seabird health and sustainability in the Atlantic Ocean.

63 **Keywords:** Contaminant Exposure, Stable Isotopes, Bioaccumulation, Seabirds

64

65 1. Introduction

66

67 The biodiversity of coastal areas is often threatened by human activities, making it crucial
68 to understand the ecology of these regions (Fuentes et al., 2020). Brazil's coastal zone harbors vital
69 and dynamic ecosystems that support a range of critical ecological and socio-economic functions
70 (Bighetti et al., 2021; Herbst et al., 2020). However, this zone also faces many significant threats,
71 including habitat loss, pollution, overfishing, and climate change (Figueiredo et al., 2015). These
72 threats endanger the diverse marine species inhabiting this region and impact the millions of people
73 who rely on these resources for their livelihoods and well-being (Herbst et al., 2020).

74 To address these challenges, it is crucial to deepen our understanding on the ecology and
75 biochemistry of key species in this ecosystem (Van Weerelt et al., 2013; Bighetti et al., 2021; Signa
76 et al., 2021). Seabirds, in particular, are important bioindicators of environmental health and
77 ecological processes in coastal areas (Signa et al., 2021; Bighetti et al., 2022). As top predators,
78 they provide valuable means of monitoring the impacts of human activities on marine ecosystems
79 (Cunha et al., 2012; Dias et al., 2013).

80 The Brown Booby (*Sula leucogaster*) is a pantropical species commonly found along the
81 Brazilian coast. Studying these birds across various locations along the Brazilian coast enhances
82 our understanding on the ecological dynamics and challenges affecting coastal ecosystems,
83 thereby facilitating the development of targeted conservation strategies.

84 Trace elements (TEs) and stable isotopes serve as valuable tools for comprehending the
85 ecology of seabirds (Cipro et al., 2018; Bond & Lavers, 2020; Padilha et al., 2021). Trace elements

86 are crucial in various physiological processes and are often used to assess avian health (Zaman et
87 al., 2022). They can also be used to discriminate among populations and species based on
88 differences in exposure to environmental contaminants and dietary intake (Burger & Gochfeld,
89 2000, 2001; Moura et al., 2018). Additionally, stable isotope ratios are essential for understanding
90 seabird foraging habitats ($\delta^{13}\text{C}$), trophic level ($\delta^{15}\text{N}$), and habitat use ($\delta^{34}\text{S}$) (Connolly et al., 2004;
91 Cherel et al., 2014; Hobson et al., 2015; Polito et al., 2016; Pizzochero et al., 2018), helping to
92 elucidate the relationships between seabirds and their environments and offering a comprehensive
93 understanding of their ecological niches and responses to environmental changes.

94 Feathers are indicated as a suitable matrix for studying trace elements because birds
95 eliminate these contaminants through this integumentary system (Dolan et al., 2017). During the
96 molting period, there is a significant increase in the trace element concentrations in the blood of
97 these animals, indicating the remobilization of elements stored in other tissues into the bloodstream
98 (Burger, 1993; Dolan et al., 2017). As feathers develop, they have a blood supply, which will cease
99 after the complete development of these integumentary structures (Burger, 1993). Once fully
100 formed, feathers retain immobilized elements linked to sulfhydryl groups in their structure
101 (Burger, 1993; Yao, et al., 2021). In this way, the feathers will reflect the plasmatic concentrations
102 of these elements during the molting period, since there will be no further possibility of
103 remobilizing trace elements stored in this matrix (Burger, 1993).

104 Few studies have examined the differences among populations of the same species across
105 various geographic locations. In the case of the Brown Booby (*Sula leucogaster*), most of the
106 research that has been conducted primarily focuses on organic pollutants (Cunha et al., 2012;
107 Mello et al., 2012). Additionally, there is limited research on how proximity to the coast influences
108 exposure to pollutants and the ecological responses of these seabirds (da Silva et al., 2023). To the

109 authors' knowledge, no studies have compared different populations of this species across Saint
110 Peter and Saint Paul, Abrolhos, and Cagarras Archipelagos in the Atlantic Ocean in terms of
111 inorganic contaminant exposure. These gaps hinder our ability to fully understand the ecological
112 dynamics and conservation needs of coastal ecosystems.

113 Studies have shown that TEs like Pb, Cd, Cu, and Sn are often elevated in seabird
114 populations near urban and industrial areas due to pollution from sources such as urban runoff,
115 industrial discharges, and maritime activities (Burger & Gochfeld, 2000; Dorneles et al., 2020;
116 Moura et al., 2018). For example, seabirds in proximity to highly urbanized areas, such as Rio de
117 Janeiro, have shown higher concentrations of these elements compared to seabirds from a less
118 polluted area in the North Pacific (Burger & Gochfeld, 2000; de Assis Padilha et al., 2018).
119 However, not all TEs are influenced by urbanization; some, like Mg and Sr, are more affected by
120 natural geological factors or specific oceanographic conditions (Jerez et al., 2011; Gama et al.,
121 2022; da Costa et al., 2023).

122 In addition to geographical differences, variations in TE concentrations are also influenced
123 by the sex and age of seabirds. Previous research has indicated that juveniles may accumulate Zn
124 and Cu because of their higher nutritional requirements during their development (Lerma et al.,
125 2020). For instance, Zn is essential for growth and is typically found in higher concentrations in
126 younger individuals (Lerma et al., 2020; Zaman et al., 2022). Additionally, sex-based differences
127 have been reported in some studies, where females may exhibit different TE concentrations due to
128 factors such as egg-laying, which can influence the mobilization and excretion of certain elements
129 (Jerez et al., 2011; Bighetti et al., 2021).

130 However, these patterns are not universal across all species or environments. In some cases,
131 studies have found no significant sex-based differences in TE concentrations, suggesting that both

132 males and females are similarly exposed to environmental contaminants, particularly when both
133 sexes occupy similar ecological niches and share similar diets (Bighetti et al., 2021). Therefore, it
134 is crucial to consider the specific environmental context and species behavior when predicting how
135 sex and age might influence TE accumulation.

136 To address these existing gaps in our understanding of seabird ecology and exposure to
137 contaminants, our study investigates variations in trace element concentrations and stable isotope
138 compositions in the feathers of Brown Boobies from different islands in the South Atlantic Ocean.
139 By analyzing these biochemical markers, we aim to discern differences in diet, habitat use, and
140 exposure to contaminants among distinct Brown Booby populations. The varying distances of the
141 islands from the coast are essential in understanding how the proximity to coastal or oceanic
142 environments influences dietary sources and bioaccumulation of trace elements (Cunha et al.,
143 2012). This approach will enable us to evaluate the effects of environmental factors on these birds
144 and the ecosystems they inhabit.

145 Therefore, our objectives are:

- 146 - To compare trace element concentrations among Brown Booby populations from the three
147 archipelagos: Cagarras Islands, Abrolhos, and SPSP, focusing on the influence of the proximity
148 to urban and industrial areas.
- 149 - To use stable isotope data ($\delta^{13}\text{C}$, $\delta^{15}\text{N}$, $\delta^{34}\text{S}$) to assess trophic positions and dietary patterns of
150 Brown Booby populations from the three archipelagos, identifying variations in the elemental
151 concentrations influenced by local prey availability and nearby fishing and urban activities.
- 152 - To investigate the influence of maturity stage and sex on the trophic position and contaminant
153 exposure in Brown Boobies, examining differences between juveniles and adults, as well as
154 males and females, across the three populations.

155 We hypothesize that: 1) The Brown Booby population from the Cagarras Islands will
156 exhibit higher concentrations of trace elements and has a higher $\delta^{15}\text{N}$ value compared to
157 populations from Abrolhos and SPSP. This is due to their closer proximity to urban and industrial
158 areas with intense anthropogenic activity, influencing their diet and exposure to environmental
159 contaminants; 2) maturity stage and sex will influence the $\delta^{15}\text{N}$ values and exposure to
160 contaminants, with juveniles and adults, as well as males and females, showing different profiles.

161 2. Materials and Methods

162 2.1 Study area and sampling

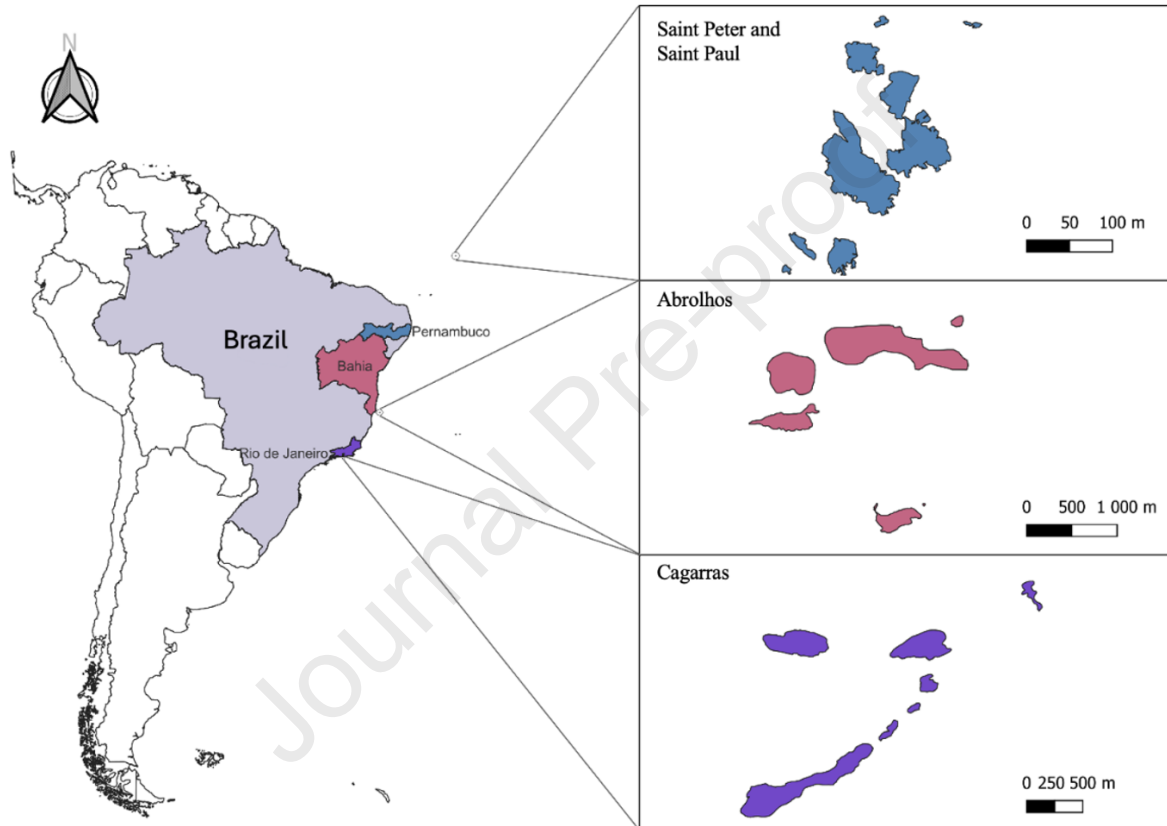
163
164 Samples were collected from Saint Peter and Saint Paul Archipelago (SPSP, $n = 32$,
165 $00^{\circ}55.01'\text{N}$, $029^{\circ}20.76'\text{W}$) in 2015, from Abrolhos Archipelago (Abrolhos, $n = 20$, $18^{\circ}00.00'\text{S}$,
166 $38^{\circ}40.00'\text{W}$) in 2015, as well as from the Cagarras Islands Natural Monument (Cagarras, $n = 35$,
167 $23^{\circ}01.58'\text{S}$, $43^{\circ}11.56'\text{W}$) between 2011 and 2015 (Figure 1). SPSP is the farthest archipelago from
168 the coast, 1010 km away from Pernambuco state, while Abrolhos is located 70 km from Bahia
169 state, and Cagarras is just 4 km from the coast of Rio de Janeiro city (Rio de Janeiro state). Adult
170 male and female Brown Boobies were captured at all three locations, while juveniles were only
171 sampled at the Cagarras Islands. As juveniles do not exhibit sexual dimorphism, the sex of the
172 juveniles was not determined.

173 It is important to note that on 5th November 2015, one of Brazil's worst environmental
174 disasters occurred when the Fundão dam collapsed, releasing approximately 50 million cubic
175 meters of metal-rich mud into the Doce River (Coimbra et al., 2020). This mud flowed into the
176 Atlantic Ocean, significantly impacting marine ecosystems, including the Abrolhos Archipelago,
177 located about 200 km north of the river's mouth. Although the samples in this study were collected

178 before the disaster, our data are essential as they provide a baseline and can be considered in future
179 studies to assess potential long-term effects on trace element levels in the region's wildlife.

180

181



182

183

184 Figure 1. Study Area, highlighting the islands considered in the current study for the sampling of
185 Brown Booby (*Sula leucogaster*) feathers. The map was created using QGIS 3.32 with the WGS
186 84 datum.

187 One feather was collected from the primary remiges (P8) for each individual, placed in
188 polyethylene bags, and kept at room temperature until the analyses. The seabirds were captured in
189 their nests with the help of a pull net. A digital caliper (accuracy: ± 0.2 mm) was used to measure
190 the beak and tarsus of the birds. Tail and wing measurements were taken with a 50 cm ruler. Body
191 mass was measured using a 2,500 g dynamometer-type precision scale (Accuracy: $\pm 0.3\%$). The
192 captured birds were identified with metal bands provided by the National Center for Research and
193 Conservation of Wild Birds (CEMAVE), from the Chico Mendes Institute for Biodiversity
194 Conservation (ICMbio).

195 **2.2 Sample preparation**

196 The feathers were washed three times with a sequence of Milli-Q ultrapure water (Merck
197 Millipore, USA), 0.01% EDTA (Spectrum, Tedia, USA), and finally Milli-Q ultrapure water
198 (Merck Millipore, USA) to eliminate external contamination. They were then oven-dried at 50 °C
199 for 24 h (Padilha et al., 2021) before being cut up using stainless steel scissors. An aliquot of 0.1
200 g of each sample was submitted to acid digestion (5 ml of HNO₃, 2 ml of H₂O₂, and 1 ml of
201 ultrapure deionized water (Milli-Q system). The samples were submitted to the microwave
202 digestion program for 15 min. After digestion, the samples were swelled to 50 mL with ultrapure
203 water (Milli-Q system).

204 **2.3 Elemental measurements**

205 The concentrations of Barium (Ba), Calcium (Ca), Cadmium (Cd), Chromium (Cr),
206 Cesium (Cs), Copper (Cu), Iron (Fe), Lithium (Li), Magnesium (Mg), Manganese (Mn), Nickel
207 (Ni), Lead (Pb), Rubidium (Rb), Selenium (Se), Tin (Sn), Strontium (Sr) and Zinc (Zn) were
208 determined with an inductively coupled plasma mass spectrometer (ICP MS, Perkin Elmer Elan -
209 9000) at the University of Liege, Belgium. As a Quality Assurance / Quality Control (QA/CQ)

210 procedure, the certified reference material was treated and analyzed in the same way as the
211 samples. The concentrations of these elements were only considered valid when the results of the
212 analyses of the NIES-1 (human hair; n = 8) were in agreement (between 90 and 110%) with the
213 value certified by the National Institute for Environmental Studies of Japan (NIES) (Table S1).
214 Additionally, blank solutions were analyzed using identical procedures as the samples. Such
215 QA/QC procedure allows any contamination in any analytical step to be observed when reading
216 such solutions. The limits of detection (LOD, in $\mu\text{g}\cdot\text{L}^{-1}$) and quantification (LOQ, in $\text{mg}\cdot\text{kg}^{-1}$)
217 depend on the amount of sample mineralized and the dilution volume. The LOQ in $\text{mg}\cdot\text{kg}^{-1}$ is
218 calculated as LOQ in $\mu\text{g}\cdot\text{L}^{-1}$ multiplied by the dilution volume divided by the weight of the
219 mineralized sample. The LOD values for each element were calculated as 0.3 times the
220 corresponding LOQ value, and the results are presented in Table S2.

221 **2.4 Measurement of C, N, and S isotopic ratios**

222 The elemental and isotopic composition of C, N, and S of the samples was determined by
223 continuous flow elemental analyses - Isotope Ratio Mass Spectrometry (CF - EA - IRMS) at the
224 University of Liège (Belgium), using a CNC elemental analyzer Vario MICRO cube (Elementary
225 Analysensysteme GmbH, Hanau, Germany) coupled to an IsoPrime100 isotope ratio mass
226 spectrometer (Isoprime, Cheadle, UK). Isotopic ratios were conventionally expressed as δ values
227 according to the standard established by (COPLEN et al., 2011) relative to international standards
228 C and N i.e. Vienna Pee Dee Belemnite (VPBD) for carbon, and atmospheric air for nitrogen.
229 International Atomic Energy Agency (IAEA, Vienna, Austria) certified reference materials IAEA-
230 C6 ($\delta^{13}\text{C} = -10.8 \pm 0.5 \text{ ‰}$, mean \pm SD), IAEA-S1 ($\delta^{34}\text{S} = -0.3\text{‰}$; mean), and IAEA-N2 ($\delta^{15}\text{N} =$
231 $20.3 \pm 0.2 \text{ ‰}$) as primary analytical standards, as well as sulphanilic acid ($\delta^{13}\text{C} = -25.9 \pm 0.3$; $\delta^{15}\text{N}$
232 $= -0.12 \pm 0.4$; $\delta^{34}\text{S} = 5.9 \pm 0.6$; mean \pm SD in each case) as secondary analytical standards. The

233 isotopic ratios of the samples were normalized using primary analytical standards. Standard
234 deviations in repeated measurements on multiple batches of secondary analyses (sulphanilic acid)
235 and laboratory standards (seabird feather) analyzed interspersed between samples (one repeat of
236 each standard every 15 samples) were 0.2 ‰ for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ and 0.4‰ for $\delta^{34}\text{S}$.

237 **2.5 Statistical analyses**

238 The statistical analyses were conducted using the R statistical environment (R Core Team,
239 2019). Data normalization was performed by log-transforming the concentrations to base 10.
240 Normality tests were then conducted to confirm that the log-transformed data followed a normal
241 distribution. Parametric tests were used to explore differences in trace element concentrations and
242 stable isotope compositions among the Brown Boobies from different localities and maturity
243 stages.

244 Since juveniles do not exhibit sexual dimorphism, we treated them as a distinct group
245 without distinguishing their sex. To assess differences in trace element concentrations and stable
246 isotope compositions between different localities, sex (females and males), and among maturity
247 stages (adults: females, males, and juveniles), we initially used ANOVA. If a significant difference
248 was found in the ANOVA, the Tukey post-hoc test was applied to determine the specific pairs of
249 groups that differed. Additionally, we applied the Benjamini-Hochberg (BH) method to adjust the
250 p-values for multiple comparisons, reducing the risk of false discoveries due to the number of tests
251 performed.

252 The SIBER (Stable Isotope Bayesian Ellipses in R) method (Jackson et al., 2011) was
253 utilized to explore ecological niches across different species. The ellipse areas were estimated
254 using the SEAc correction and Bayesian modeling for intergroup pairwise comparisons. The SEAb
255 (Bayesian estimate of the standard ellipse area) provides a comparative measure of niche widths

256 between groups based on the size of simulated ellipse areas and their estimated posterior
 257 distributions. Groups with similar SEAb values are inferred to have similar isotopic niche widths,
 258 indicating reliance on a similar diversity of prey items and feeding habitats. The SIBER 2.1.4
 259 method (Jackson et al., 2011) was employed for these analyses. Additionally, the Pearson
 260 correlation coefficient test was conducted to identify significant relationships between trace
 261 elements and stable isotopes.

262 3. Results

263 Among the 18 trace elements measured (Table 1, Figure 2) in the feathers of Brown Booby,
 264 the ANOVA test ($p < 0.05$) demonstrated that 8 element concentrations (Li, Fe, Mn, Rb, Cd, Sn,
 265 Cs, and Pb) differed between islands. Beryllium was the only element below the limits of detection.

266 Table 1. Mean \pm standard deviation ($\mu\text{g g}^{-1}$ dw) of the concentrations of essential and non-essential
 267 elements in Brown Booby (*Sula leucogaster*) feathers.

268

	SPSP		Abrolhos		Cagarras		
Sex	Male	Female	Male	Female	Male	Female	Juvenile
<i>n</i>	18	14	7	13	12	13	10
Li	0.06 \pm 0.02	0.07 \pm 0.02	0.07 \pm 0.01	0.07 \pm 0.03	0.07 \pm 0.03	0.06 \pm 0.02	0.04 \pm 0.02
Mg	577 \pm 131	645 \pm 147	651 \pm 194	533 \pm 87	623 \pm 391	542 \pm 151	316 \pm 230
Ca	761 \pm 140	896 \pm 253	689 \pm 91	789 \pm 214	616 \pm 210	690 \pm 154	653 \pm 117
Cr	0.68 \pm 0.44	0.65 \pm 0.49	0.62 \pm 0.45	0.40 \pm 0.19	0.77 \pm 0.84	0.49 \pm 0.28	0.97 \pm 0.83
Fe	15 \pm 16	11 \pm 3.5	21 \pm 20	15 \pm 5.1	29 \pm 20	21 \pm 9.1	35 \pm 24
Mn	0.21 \pm 0.06	0.28 \pm 0.21	0.24 \pm 0.09	0.20 \pm 0.05	0.82 \pm 0.82	0.62 \pm 0.41	0.84 \pm 0.53

Ni	0.26± 0.11	0.28± 0.20	0.16± 0.03	0.18± 0.05	0.24± 0.10	0.23± 0.08	0.27± 0.11
Cu	6.41± 0.94	6.73± 0.81	6.20± 0.63	7.48± 0.90	8.02± 2.50	8.15± 0.96	6.94± 0.88
Zn	113± 14	121± 18	114± 27	111± 12	102± 31	109± 22	151± 37
Se	2.48± 0.34	2.54± 0.45	2.70± 0.25	2.75± 0.35	1.78± 0.88	2.07± 0.99	1.59± 0.79
Rb	0.05± 0.02	0.05± 0.01	0.05± 0.01	0.05± 0.01	0.07± 0.04	0.06± 0.02	0.12± 0.08
Sr	7.27± 1.62	8.15± 2.61	8.19± 2.93	6.47± 0.81	5.59± 2.27	6.76± 7.52	3.76± 3.19
Cd	0.06± 0.04	0.05± 0.02	0.02± 0.01	0.01± 0.01	0.02± 0.01	0.03± 0.04	0.02± 0.01
Sn	0.01± 0.01	0.02± 0.01	0.02± 0.02	0.01± 0.002	0.02± 0.01	0.02± 0.01	0.02± 0.01
Cs	0.001± 0.00	0.001± 0.00	0.001± 0.00	0.001± 0.00	0.002± 0.00	0.001± 0.01	0.002± 0.00
Ba	0.09± 0.08	0.09± 0.03	0.07± 0.01	0.10± 0.06	0.13± 0.07	0.15± 0.05	0.12± 0.04
Pb	0.22± 0.15	0.35± 0.12	0.20± 0.01	0.23± 0.15	0.37± 0.12	0.38± 0.14	0.28± 0.18
$\delta^{15}\text{N}$	12.0± 0.65	12.0± 0.46	12.0± 1.42	13.0± 1.51	15.0± 0.92	15.0± 0.93	13.0± 1.37
$\delta^{13}\text{C}$	-17± 1.57	-17± 1.45	-16± 0.95	-16± 1.35	-15± 1.17	-15± 1.86	-15± 1.49
$\delta^{34}\text{S}$	16.0± 0.94	16.0± 0.97	14.0± 1.41	15.0± 1.75	13.0± 1.37	14.0± 1.35	14.0± 0.97

269

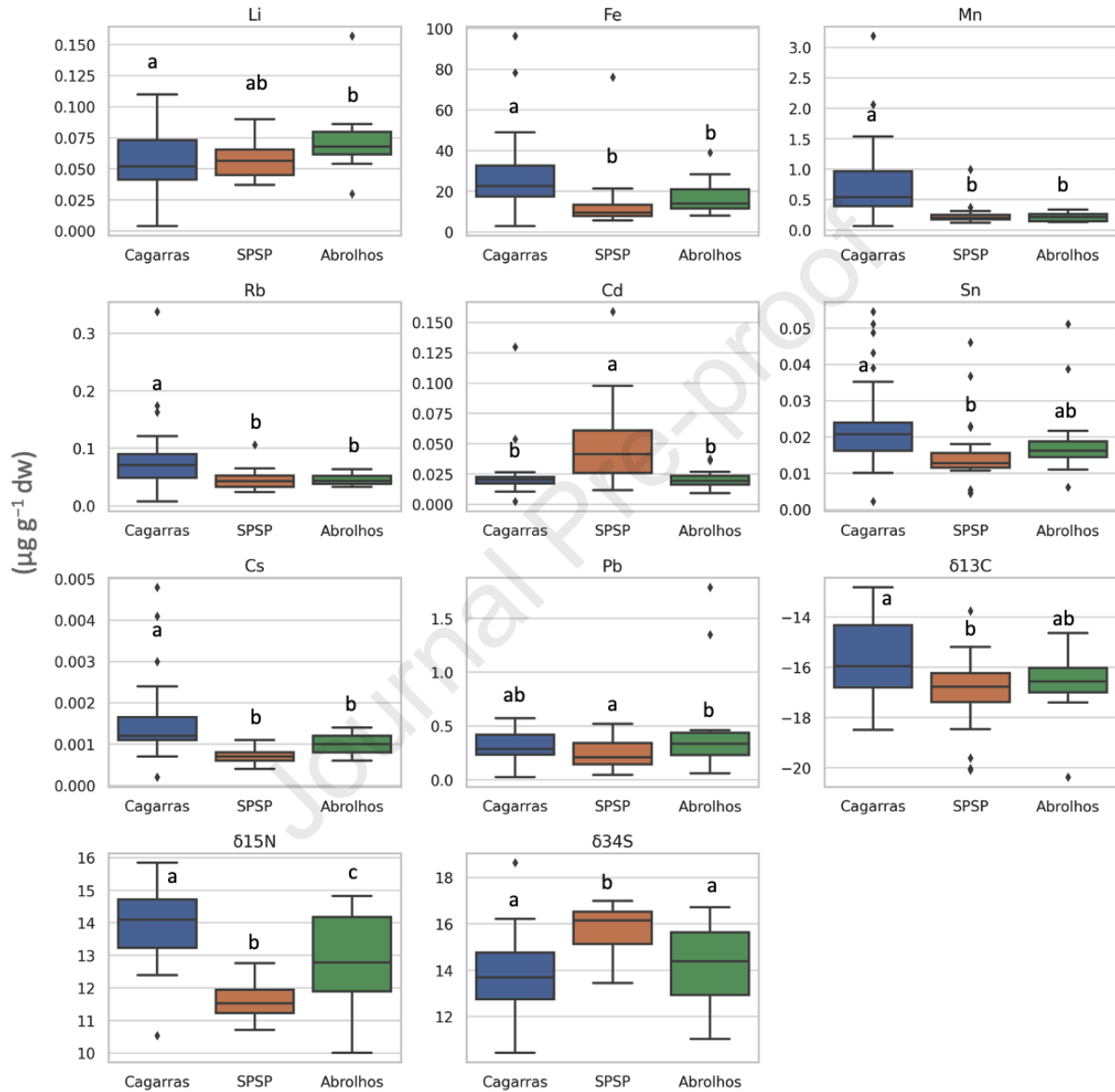
270 3.1 Geographical Comparison

271 The ANOVA test revealed significant differences in trace element concentrations among
 272 the Brown Booby populations from different islands ($F = 5.23$, $p < 0.001$). Brown Boobies from
 273 Cagarras had significantly higher concentrations of Fe ($p < 0.001$), Mn ($p < 0.001$), Rb ($p < 0.001$),
 274 and Cs ($p < 0.001$) compared to both Abrolhos and SPSP (Tukey post hoc Test). Tin concentrations
 275 were significantly higher in Cagarras ($p = 0.003$) compared to SPSP.

276 Cd concentrations were significantly higher in SPSP compared to both Cagarras ($p < 0.001$)
 277 and Abrolhos ($p < 0.001$). Brown Boobies from Abrolhos presented significantly higher Li

278 concentrations ($p = 0.039$) than those from Cagarras. Lead concentrations were significantly higher
 279 in Abrolhos ($p = 0.032$) than SPSP.

280



281

282 Figure 2. Boxplots of concentrations of essential and non-essential concentrations (Li, Fe, Mn, Rb,
 283 Cd, Sn, Cs, and Pb) and stable isotopes ($\delta^{13}\text{C}$, $\delta^{15}\text{N}$, and $\delta^{34}\text{S}$) in feathers of Brown Booby (*Sula*
 284 *leucogaster*) from Cagarras, Abrolhos, and SPSP archipelagos. Different letters (abc) indicate
 285 significant differences between groups.

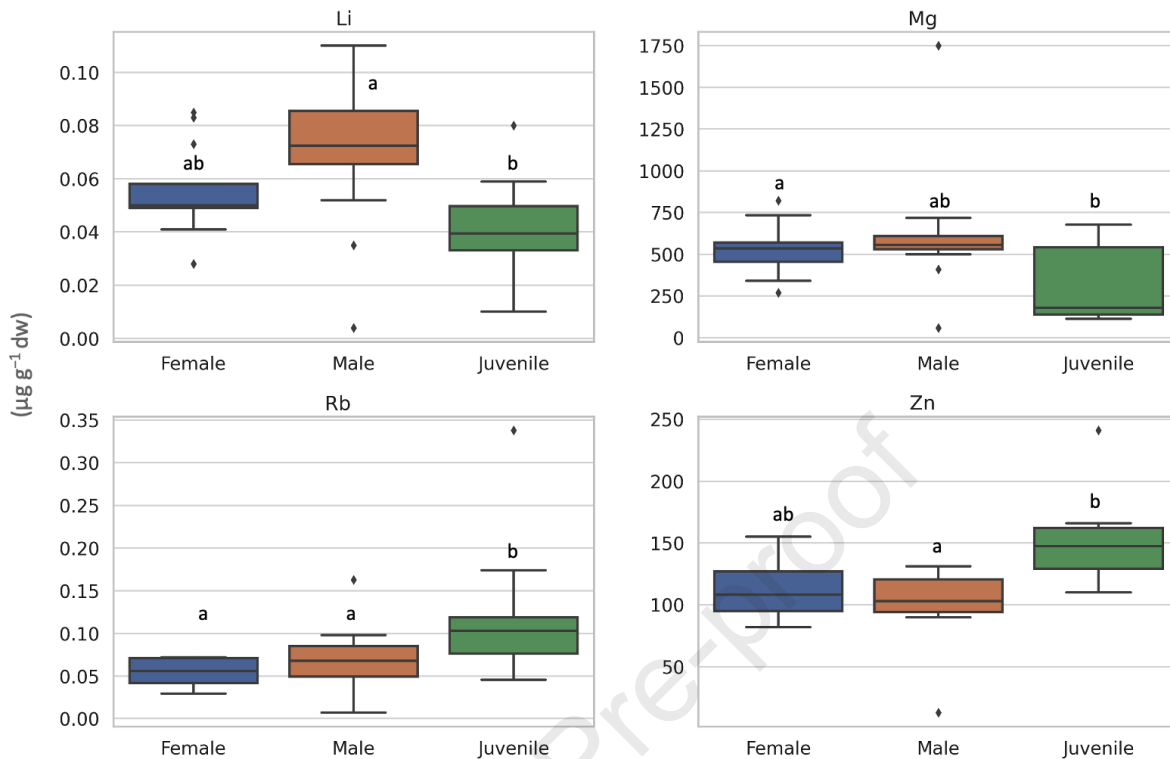
286 3.2 Sex-Based Comparison and Maturity Stage

287 The ANOVA analyses conducted separately for each location did not reveal any significant
288 differences in elemental concentrations between males and females for any of the elements studied
289 ($p > 0.05$).

290 Elemental concentrations differed significantly between juveniles and male and female
291 separately (Figure 3). Specifically, juveniles had significantly lower concentrations of Li ($F =$
292 4.595 , $p = 0.018$) and higher concentrations of Zn ($F = 8.395$, $p = 0.001$) compared to adult males
293 (post hoc test both $p = 0.01$). Magnesium levels ($F = 3.482$, $p = 0.043$) were significantly lower in
294 juveniles compared to adult females ($p = 0.02$), and also compared to adult males ($p = 0.03$).
295 Rubidium ($F = 5.019$, $p = 0.013$) was significantly higher in juveniles compared to adult females
296 ($p = 0.01$), as well as between males and juveniles ($p = 0.05$).

297 The ANOVA analyses revealed significant differences in body measurements (weight,
298 beak length, wing length, and tail length, Table S3) based on sex and maturity stage. Post hoc
299 Tukey's HSD tests indicated that females were significantly larger than both males and juveniles
300 across all measurements. Specifically, females had greater weight, beak length, wing length, and
301 tail length (all $p < 0.001$ for weight, beak length, and tail length; $p = 0.002$ for wing length)
302 compared to males and juveniles. Males were also significantly larger than juveniles in weight,
303 beak length, and tail length (all $p < 0.001$).

304



305

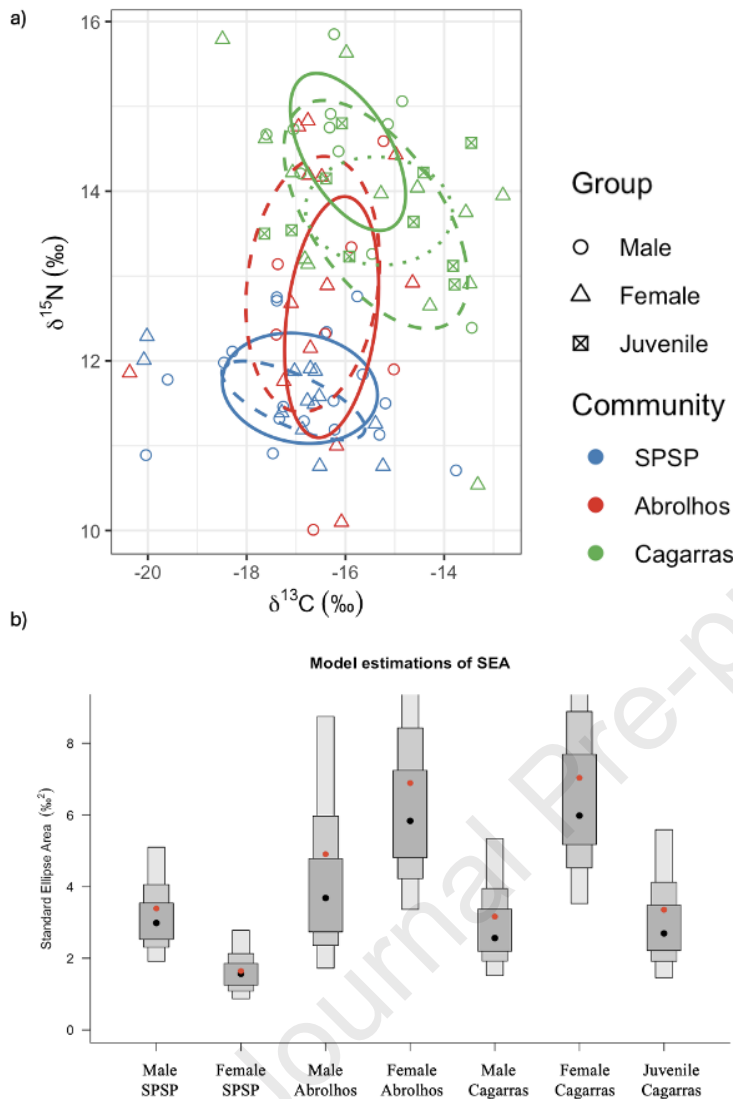
306 Figure 3. Boxplots of concentrations of metals (Li, Mg, Rb, and Zn) in Brown Boobies (*Sula*
 307 *leucogaster*) from the Cagarras location, categorized by adult female, juvenile (sex non-identified),
 308 and adult males. Different letters (abc) indicate significant differences between groups.

309 3.3 Trophic Niche Comparisons

310 Isotope values differed significantly among locations for $\delta^{13}\text{C}$ ($F = 5.733$, $p = 0.005$), $\delta^{15}\text{N}$
 311 ($F = 34.854$, $p < 0.001$), and $\delta^{34}\text{S}$ ($F = 15.421$, $p < 0.001$) (Figure 2). For $\delta^{13}\text{C}$, SPSP showed
 312 significantly lower values compared to Cagarras (diff = -1.333, $p = 0.003$). For $\delta^{15}\text{N}$, Cagarras had
 313 significantly higher values compared to Abrolhos (diff = 1.295, $p < 0.001$), and SPSP had
 314 significantly lower values compared to Abrolhos (diff = -1.149, $p = 0.001$) and Cagarras (diff = -
 315 2.444, $p < 0.001$). For $\delta^{34}\text{S}$, SPSP had significantly higher values compared to Abrolhos (diff =
 316 1.420, $p = 0.003$) and Cagarras (diff = 2.103, $p < 0.001$).

317 SIBER results (Figure 4) suggest that the core isotopic niches of the Brown Boobies from
318 SPSP were markedly separated from those of the Cagarras group. In terms of the overlap between
319 populations, Brown Boobies from SPSP and Abrolhos exhibited a 17% overlap in their isotopic
320 niche areas. The overlap between the Cagarras and Abrolhos populations was 13% of their
321 combined area.

322 In SPSP, the isotopic niche overlap between male and female Brown Boobies was 1.59‰²,
323 which constitutes 46% of their area. In Abrolhos, this overlap was 4.16‰², or 55% of their area,
324 and in Cagarras, it was 2.72‰², or 36% of their area. Additionally, in Cagarras, the overlap
325 between male and juvenile Brown Boobies was 1.36‰² (27% of their area), while the overlap
326 between female and juvenile Brown Boobies was 3.27‰² (46% of their area). The females from
327 Abrolhos and Cagarras showed a significantly wider trophic niche area than males ($p < 0.001$)
328 (Figure 4b).



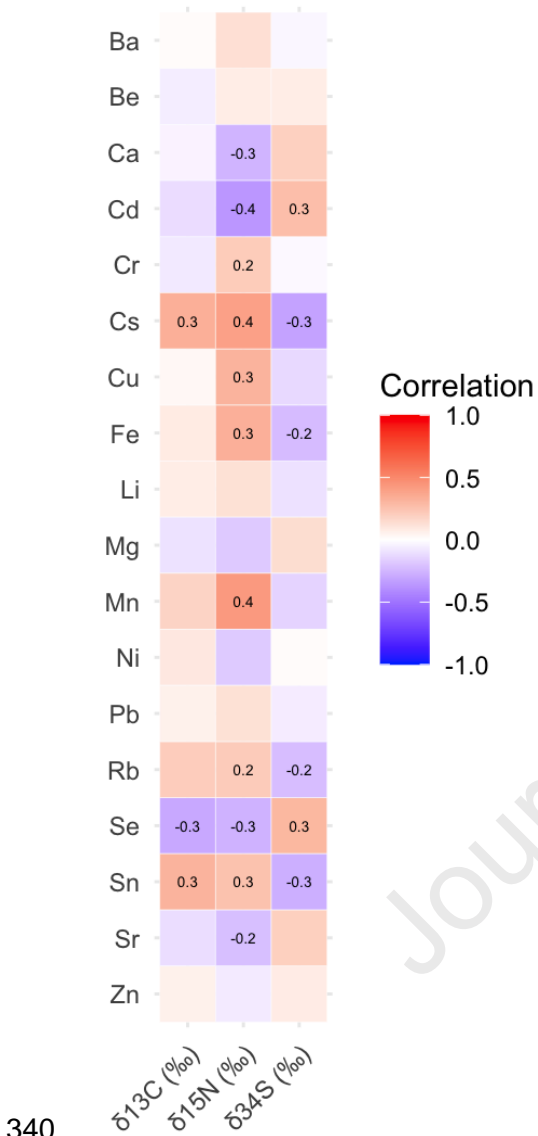
329

330 Figure 4. Isotopic niche sizes (a) for feathers of adult Brown Booby (*Sula leucogaster*), with their
 331 respective small sample-size corrected standard ellipses (b). Male (solid ellipse), Female (dashed
 332 ellipse), and Juvenile (dotted ellipse)

333 3.4 Stable isotope and element correlations

334 The Pearson correlation coefficient test (Figure 5) demonstrated several significant
 335 relationships. For $\delta^{13}\text{C}$, there is a moderate negative correlation with Se and a moderate positive
 336 correlation with Cs and Sn. For the $\delta^{15}\text{N}$, there is a strong negative correlation with Cd, a moderate
 337 negative correlation with Sr, Se, and Ca, a strong positive correlation with Cs and Mn, and a

338 moderate positive correlation with Rb, Cu, Fe, and Cr. For the $\delta^{34}\text{S}$, there is a moderate negative
 339 correlation with Cs, Sn, Rb, and Fe, and a moderate positive correlation with Cd and Se.



340

341

342 Figure 5. Results of the Pearson Correlation Test showing r values between essential and non-
 343 essential elements and stable isotopes ($\delta^{13}\text{C}$, $\delta^{15}\text{N}$, and $\delta^{34}\text{S}$) for Brown Booby (*Sula leucogaster*)
 344 populations from the archipelagos of SPSP, Cagarras, and Abrolhos.

345

346 4. Discussion

347 This study explored the concentrations of essential and non-essential elements and the
348 isotopic compositions of C, N, and S in the feathers of Brown Boobies from three distinct
349 archipelagos in the Atlantic Ocean. Our findings indicate that Brown Boobies from different areas
350 exhibit significant variations in elemental concentrations and stable isotope values, reflecting the
351 differences in their local diets and exposure to contaminants. These results provide crucial insights
352 into the environmental and ecological factors that influence these seabird populations.

353 4.1 Impact of Anthropogenic Proximity, Ecological Niches, and Dietary Influence on 354 Elemental Concentrations

355 Our results revealed significant geographical differences in elemental concentrations. As
356 expected, the Cagarras Archipelago population exhibited notably higher concentrations of Fe, Mn,
357 Rb, Cs, and Sn. These elevated levels support the hypothesis that proximity to urban areas
358 increases exposure to specific contaminants. However, this trend is not uniform for all elements,
359 suggesting a complex relationship between proximity to urban areas and contaminant sources.

360 Our findings are aligned with previous studies indicating heightened pollutant levels in
361 seabird populations near urban centers, due to increased anthropogenic activities (Gilmour et al.,
362 2019). For example, Finger et al. (2016) found that seabirds, such as little penguins (*Eudyptula*
363 *minor*), in areas with significant human influence have higher concentrations of elements like Al,
364 As, B, Ca, Cu, Fe, Hg, Pb, Se, and Zn in their blood. Similarly, Cagarras' proximity to a megacity
365 (Rio de Janeiro), with its substantial industrial and urban runoff, likely leads to the
366 bioaccumulation of pollutants in local wildlife (Dorneles et al., 2020).

367 Industrial activities and urban runoff likely influence the presence of elevated Rb and Cs
368 concentrations in Cagarras in the Rio de Janeiro area, where these elements are commonly

369 associated with electronics and petroleum extraction industries, with cesium specifically utilized
370 in drilling fluids and photoelectric cell production (Riley, 2019; Varala & Rao, 2015). Concerning
371 tin concentrations, it is important to highlight that previous studies from our research team have
372 demonstrated being possible to use total tin (ΣSn) concentrations for evaluating vertebrate
373 exposure to organotin compounds (Dorneles et al., 2008; Dorneles et al., 2020). The rationale for
374 this comes from the fact that inorganic tin forms are practically not absorbed through the digestive
375 systems of vertebrates (Berman, 1980; Guruge et al., 1996; Schilithz et al., 2013). Furthermore,
376 ΣSn biomagnification was found in the ichthyofauna of RJ waters at TMF (Trophic Magnification
377 Factor) values that were similar to those found for tributyltin (TBT) (Dorneles et al., 2020). These
378 aspects associated with the fact that the loss of integumentary structures has also been verified as
379 an excretion route for organotins (OTs) in seabirds (Guruge et al., 1996) strengthen the suitability
380 of the use of ΣSn concentrations for evaluating Brown Booby exposure to OTs. Additionally, the
381 significant maritime traffic around Guanabara Bay, a major port area, contributes to elevated tin
382 levels in Cagarras' Brown Booby population due to the extensive use of OTs in antifouling paints
383 on ships and boats (Padilha et al. 2015, de Assis Padilha et al., 2018; Castro et al., 2021).
384 Organotins are highly toxic compounds used to prevent biofouling, leach into the marine
385 environment, affecting local wildlife. Despite international bans by the International Maritime
386 Organization (IMO) in 2008, OT-based antifouling paints are still used in Brazil (Maciel et al.,
387 2018; Castro et al., 2021). Elevated tin levels in the Cagarras Brown Boobies are concerning due
388 to OT's toxicological effects, such as endocrine disruption, immunotoxicity, and reproductive
389 issues in birds, potentially leading to population declines (Maciel et al., 2018; Frouin et al., 2010).

390 Although higher levels of some elements in the Cagarras population could be partially
391 attributed to urban proximity, we cannot ignore the potential impact of differences in food webs

392 and local geology across the islands. Previous studies have shown that certain elements, like Cd,
393 can be elevated due to specific geological features, even in remote areas (Jerez et al., 2011).
394 Therefore, it is likely that a combination of environmental and ecological factors is influencing the
395 patterns observed. In this context, the higher concentrations of certain elements in Abrolhos and
396 SPSP are likely due to different environmental factors. In Abrolhos, Brown Boobies exhibited
397 elevated levels of Li and Pb. The Fundão dam collapse affected seabird foraging areas in the
398 Atlantic Ocean, including the Abrolhos Archipelago (Coimbra et al., 2020), significantly altering
399 the elemental levels in tropical seabirds from Abrolhos (Nunes et al., 2022). According to Nunes
400 et al. (2022), the collapse led to increased levels of non-essential elements like As and Cd, while
401 essential elements like Fe, Mn, and Zn decreased (Nunes et al., 2022). Furthermore, as shown in
402 Table 2, when compared with other studies worldwide, the post-collapse Cd levels in seabirds from
403 Abrolhos are notably higher, highlighting the severe impact of the disaster on contaminant
404 bioaccumulation in this region. In addition, a recent study by Bauer et al. (2024) reported
405 significant temporal changes in element concentrations in the blood and feathers of tropical
406 seabirds, including Brown Boobies, following the Fundão Dam disaster. The study reported acute
407 contamination events, notably observed in February 2021, which occurred after a major rainy
408 season in the Doce River basin (Bauer et al., 2024). This surge likely remobilized contaminated
409 sediments, which were transported into the ocean and led to elevated levels of non-essential
410 elements like As, Cd, Hg, and Pb in Brown Booby feathers. Additionally, the Abrolhos Bank
411 experiences frequent cold fronts and southern winds that resuspended sediments, potentially
412 leading to the remobilization of certain elements (Gama et al., 2022), which could include Li and
413 Pb. These findings underscore the role of tropical seabirds as indicators of marine pollution,
414 revealing strong temporal patterns in elemental concentrations driven by climatic and

415 environmental processes (Bauer et al., 2024). It is relevant to highlight that our study samples were
416 obtained before the Fundão dam collapse. In our study, the concentrations of non-essential
417 elements such as Cd ($0.02 \pm 0.01 \mu\text{g g}^{-1}$, mean \pm SD) and Pb ($0.23 \pm 0.01 \mu\text{g g}^{-1}$) in the feathers
418 of Brown Boobies from Abrolhos were lower compared to Nunes et al. (2022) and Bauer et al.,
419 2024 post-collapse (Cd: $0.436 \mu\text{g g}^{-1}$, Pb: $1.03 \pm 0.01 \mu\text{g g}^{-1}$ respectively). Essential elements like
420 Fe ($15 \pm 20 \mu\text{g g}^{-1}$), and Mn ($0.24 \pm 0.09 \mu\text{g g}^{-1}$) were also lower in our study compared to post-
421 collapse levels (Fe: $40 \mu\text{g g}^{-1}$, Mn: $19 \pm 0.09 \mu\text{g g}^{-1}$). This highlights the impact of severe
422 anthropogenic events on element concentrations and underscores the need for continuous
423 monitoring to assess long-term impacts on seabird populations.

424 SPSP, being the farthest from the mainland, showed higher concentrations of Cd, which
425 may be linked to unique oceanographic conditions and the presence of deep-sea hydrothermal
426 vents (da Costa et al., 2023). Cadmium concentrations in birds can vary based on age, diet,
427 ecosystem use, and physiological status (Burger et al., 1993). Cadmium's high solubility,
428 significant hydrophilicity, and low sedimentation rate facilitate its dissolution and dispersion in
429 seawater, leading to higher bioavailability and accumulation in marine organisms (Barcellos,
430 1995). Differences in cadmium levels between coastal (e.g., Cagarras Archipelago) and oceanic
431 (e.g., SPSP) environments were notable. Oceanic species tend to have higher Cd levels than coastal
432 species, likely due to the higher cadmium concentrations found in oceanic cephalopods compared
433 to those from coastal areas (Dorneles et al., 2007).

434 Concentrations of essential elements (Mg, Ca, Fe, Mn, Cu, Zn, Se) in the feathers of Brown
435 Boobies from our study were within the range previously reported for seabirds worldwide (Burger
436 & Gochfeld, 2000; Dolci et al., 2017; de Assis Padilha et al., 2018; Moura et al., 2018). This
437 consistency is expected, as essential elements are under homeostatic control with nutritional

438 requirements regulating their absorption (Walsh, 2018). A few studies have reported toxic levels
439 of certain elements in feathers. For instance, levels starting at $200 \mu\text{g g}^{-1}$ (dw) for Zn and at $26 \mu\text{g}$
440 g^{-1} (dw) for Se have been identified as harmful to bird growth and reproduction (Burger &
441 Gochfeld, 2000; Einoder et al., 2018).

442 Non-essential elements (Li, Be, Cr, Rb, Sr, Sn, Ba, and Pb) in the feathers of Brown
443 Boobies from our study were also within the range previously reported for seabirds worldwide
444 (Burger & Gochfeld, 2000; Dolci et al., 2017; de Assis Padilha et al., 2018; Moura et al., 2018).
445 Toxic effects of non-essential elements may be observed at lower concentrations, such as over 2
446 $\mu\text{g g}^{-1}$ for Cd and $4 \mu\text{g g}^{-1}$ for Pb (Burger & Gochfeld, 2000). The elemental levels reported in our
447 study are below these harmful thresholds.

448 Geographical differences in elemental concentrations may also be attributed to local dietary
449 habits and prey availability. The foraging behavior of Brown Boobies is an important factor
450 influencing their exposure to contaminants. Brown Boobies are known to forage in both coastal
451 and offshore areas, with variations in diet depending on local prey availability and oceanographic
452 conditions (Branco et al., 2005; Cunha et al., 2012). Birds from the Cagarras Archipelago, located
453 near the highly urbanized Rio de Janeiro, are likely foraging in coastal waters, where they may be
454 exposed to higher levels of pollution (Cunha et al., 2012). In contrast, birds from SPSP, located in
455 a more oceanic environment, encountering different prey, such as oceanic cephalopods, and their
456 elemental sources, may also be associated with hydrothermal vent activity (Nunes et al., 2018; da
457 Costa et al., 2023). These differences in foraging ranges and habitats could explain the observed
458 variations in elemental concentrations among the populations.

459 For instance, Becker et al. (2016) found that diet and trophic position are key factors
460 determining elemental concentrations in Southern Ocean seabirds. Our findings, using the SIBER

461 method, revealed significant isotopic niche separation among populations from SPSP and
462 Cagarras, indicating considerable dietary differences. This supports our hypothesis that the
463 Cagarras population occupies a higher trophic position compared to those from Abrolhos and
464 SPSP, influenced by the availability of higher trophic level prey and the impact of nearby fishing
465 and urban activities (Cunha et al., 2012). Boobies from more oceanic areas primarily feed on flying
466 fish and squid (Alves et al., 2004). Specifically, Brown Boobies in SPSP primarily consume flying
467 fish (*Exocoetus volitans*) and cephalopods (Nunes et al., 2018), while those from Abrolhos have a
468 diet that includes a mix of flying fish (*Exocoetus volitans*), reef-associated fish, and cephalopods
469 (Mello et al., 2012). However, Brown Boobies from Cagarras Archipelago inhabit a coastal area
470 heavily influenced by fisheries, which may lead them to consume a greater proportion of food
471 items provided by these activities (Branco et al., 2005; Cunha et al., 2012). Thus, Brown Boobies
472 from Cagarras feed more on sardines and fishing discards (Branco et al., 2005).

473 **4.2 Variations in Elemental Concentrations Based on Maturity Stage and Sex**

474 Our analyses revealed significant differences in elemental concentrations and trophic
475 positions between adults and juveniles for Li, Mg, Rb, and Zn in the Cagarras Archipelago. This
476 supports the hypothesis that the maturity stage influences the trophic position and contaminant
477 exposure. This finding supports the hypothesis that the maturity stage influences both trophic
478 position and contaminant exposure, consistent with studies by Bighetti et al. (2021, 2022) and
479 Padilha et al. (2018). Juveniles tend to accumulate certain elements differently than adults due to
480 variations in metabolic rates and dietary intake (Barbieri et al., 2010; Burger and Gochfeld, 2000;
481 Pacyna et al., 2019). Specifically, our study found higher Zn concentrations in juveniles compared
482 to adults, which is supported by similar findings in Blue-footed Booby nestlings (Lerma et al.,
483 2020). Additionally, the SIBER analyses showed that juveniles in Cagarras had a greater isotopic

484 niche overlap with females than with males, suggesting maternal influence on foraging behavior
485 and dietary preferences. Females, being larger and heavier, may be selected for a greater food
486 payload capacity and/or foraging range, providing more food for their offspring (Mellink et al.,
487 2001; Abdennadher et al. 2017).

488 Similar findings were described for Brown Boobies at Peña Blanca Island, where the $\delta^{13}\text{C}$
489 values of chicks were similar to the values of their mothers (Abdennadher et al. 2017). A broader
490 isotopic niche was observed in females, compared to males, during the post-breeding period,
491 suggesting that females explore a wider range of foraging habitats and prey types (Abdenadher et
492 al. 2017).

493 The role of body size in shaping the trophic structure of tropical seabird communities is
494 particularly relevant in this study. Mancini et al. (2014) discussed that larger seabirds often occupy
495 higher trophic positions due to their capacity to consume larger prey. This principle is evident in
496 our findings which demonstrate trophic segregation between adults and juveniles. This observed
497 body size-based trophic structure is consistent across different seabird communities and plays a
498 crucial role in niche differentiation and reducing interspecific competition during the breeding
499 season. Navarro et al. (2014) examined resource partitioning between incubating and chick-rearing
500 Brown Boobies on Christmas Island. Their $\delta^{13}\text{C}$ values revealed spatial segregation in foraging
501 grounds between species and different breeding stages of Brown Boobies. Specifically, Brown
502 Boobies shifted their foraging habitats from incubation to chick-rearing to minimize competition
503 for prey.

504 Contrary to our hypothesis, we did not observe significant sex-based differences in
505 elemental concentrations. This finding aligns with studies by de Assis Padilha et al. (2018) and

506 Bighetti et al. (2021), which also reported no significant sex-related differences in elemental
507 concentrations (Sn, Cd, Mn, Se, Cu, and Hg) in Brown Booby feathers.

508 In SPSP, there was substantial dietary overlap between Brown Boobies' males and females,
509 with an even higher overlap in Abrolhos, indicating minimal sex-based dietary differences. This
510 is consistent with findings by Lerma et al. (2020) on Masked Boobies, where overlapping isotopic
511 niches during breeding stages implied minimal diet differences between sexes. Another potential
512 factor contributing to sex-based differences in elemental concentrations is the egg-laying process
513 by females. However, it does not appear to result in significant excretion of essential and non-
514 essential elements in Brown Boobies females, likely explaining the lack of significant sex-based
515 differences in elemental concentrations (Bighetti et al., 2021). Additionally, the transfer of most
516 elements to the eggs is minimal, as the ovary restricts the passage of these elements. This selective
517 permeability of the ovarian and oviductal membranes ensures that mostly essential nutrients pass
518 through, while potentially harmful elements like non-essential elements are largely blocked, as
519 supported by studies on avian reproductive physiology (Klein et al., 2012; Sasanami et al., 2017).
520 Therefore, both sexes are equally exposed to and affected by local contamination sources,
521 highlighting the uniform environmental pressures on these populations.

522

523

524

525

526

527

528

529 Table 2. Concentrations of elements (mean in $\mu\text{g g}^{-1}$, dry weight) found in the feathers of adult seabirds of the genus *Sula* worldwide.

530

Species	Location	Reference	Cd	Cr	Cu	Fe	Mn	Zn	Pb	Sn	Li	Mg	Ca	Ni	Se	Rb	Sr	Cs	Ba
<i>Sula sula</i>	North Pacific	Burger et al. (1992)	0.13	-	-	-	-	-	2.08	-	-	-	-	-	3.68	-	-	-	-
<i>Sula sula</i>	Midway Atol	Burger & Gochfeld (2000)	0.05	2.53	2.9	-	1.46	-	0.97	2.28	-	-	-	-	2.34	-	-	-	-
<i>Sula leucogaster</i>	Brazil	Dolci et al. (2017)	0.05	-	15	47	-	94	-	-	-	815	-	0.29	-	-	-	-	-
<i>Sula leucogaster</i>	Brazil	Padilha et al. (2018)	0.03	-	8.2	-	1.6	-	-	0.2	-	-	-	-	-	-	-	-	-
<i>Sula leucogaster</i>	Brazil	Nunes et al. (2020)	0.45	0.97	0.66	40	19	25	0.09	-	-	-	-	-	-	-	-	-	-
<i>Sula leucogaster</i>	Australia	Lavers et al. (2020)	0.04	-	2.86	-	-	-	0.59	-	-	-	-	-	-	-	-	-	-
<i>Sula dactylatra</i>	Australia	Lavers et al. (2020)	0.03	-	2.92	-	-	-	0.48	-	-	-	-	-	-	-	-	-	-
<i>Sula leucogaster</i>	Brazil	Bauer et al. (2024)	0.33	-	-	-	-	-	1.03	-	-	-	-	-	-	-	-	-	-
<i>Sula leucogaster</i>	SPSP	Present study	0.06	0.62	6.41	15	0.21	113	0.22	0.01	0.06	577	761	0.26	2.48	0.05	7.27	0.001	0.09
<i>Sula leucogaster</i>	Abrolhos	Present study	0.02	0.4	7.48	15	0.2	111	0.23	0.01	0.07	533	789	0.18	2.75	0.05	6.47	0.001	0.1
<i>Sula leucogaster</i>	Cagarras	Present study	0.02	0.77	8.02	29	0.82	102	0.37	0.02	0.07	623	616	0.24	1.78	0.07	5.59	0.002	0.13

531

532

533 **5. Conclusion**

534 This study provides a comprehensive analysis of trace element concentrations and stable
535 isotope compositions in the feathers of Brown Boobies (*Sula leucogaster*) from three distinct
536 archipelagos in the South Atlantic Ocean: Cagarras, Abrolhos, and Saint Peter and Saint Paul
537 (SPSP). The results reveal significant geographical variations in elemental concentrations,
538 indicating that brown booby populations in these regions are influenced by different environmental
539 and ecological factors.

540 The Cagarras population, located near a major urban area, exhibited higher concentrations
541 of elements such as Fe, Mn, Cu, Rb, Cs, and Sn, supporting the hypothesis that proximity to urban
542 areas increases exposure to specific contaminants. This elevation in concentrations is likely related
543 to the intense anthropogenic activity in the area, including maritime traffic and industrial
544 pollutants. In contrast, the populations in Abrolhos and SPSP, which are farther from the coast and
545 direct urban influences, showed higher concentrations of elements such as Li, Pb, Mg, and Cd,
546 suggesting that local environmental factors, such as geological and oceanographic activities, also
547 play a significant role in contaminant bioaccumulation. Additionally, the study highlights the lack
548 of significant differences in elemental concentrations between sexes, suggesting that both males
549 and females are equally exposed to local contaminants despite potential differences in foraging
550 behavior. In contrast, juveniles were found to have different trace element profiles than adults,
551 reflecting changes in diet and metabolism during development.

552 Isotopic analyses revealed differences in diet and trophic niches among populations, with
553 the Cagarras population occupying a higher trophic position compared to those from Abrolhos and

554 SPSP. This could be attributed to the availability of higher trophic level prey and the influence of
555 nearby fishing and urban activities.

556 Overall, our study underscores the complex interplay between environmental
557 contamination, dietary habits, and habitat use in Brown Boobies. The findings highlight the need
558 for comprehensive conservation measures and continued monitoring to protect Brown Boobies
559 and other marine species from Atlantic Ocean.

560

561 **Acknowledgments**

562 This work was supported by the Brazilian National Council for Scientific and
563 Technological Development (CNPq) through CNPq / MCT 557049/2009–1, as well as through a
564 Universal Call CNPq - Project from PRD (proc. 432518/2016–9). This work was also supported
565 by scientific cooperation established between the Brazilian Foundation for the Coordination and
566 Improvement of Higher Level or Education Personnel (CAPES - process numbers
567 88881.154725/2017–01 88887.154724/2017–00) and Wallonie Bruxelles International (WBI,
568 from Belgium), coordinated by PRD and KD, as well as by the Rio de Janeiro State Government
569 Research Agency [FAPERJ - E-26/111.505/2010 and E - 26/210.464/2019 (249593)]. We would
570 like to thank the Brazilian Navy, which provided logistical support in Antarctica through the
571 “Secretariat of the Interministerial Commission for the Resources of the Sea” (SECIRM). GL is a
572 F.R.S.-FNRS research associate, and KD is a Senior F.R.S.- FNRS research associate. PRD has a
573 research grant from CNPq (PQ-2 proc. 08733/2019–3).

574 **6. References**

575
576 Abdennadher, A., Moreno Carillo, R., Hernandez, S., Aguilar, B., Gonzalez, D., Sanpera, C., &
577 Jover, L., 2017. Trophic segregation by sex and age in Brown Booby *Sula leucogaster*, a
578 reversed sexual size dimorphic species. *Bulletin de l’Institut National des Sciences et*

- 579 Technologies de la Mer (INSTM Salammbô), 20, 29-33.
580
- 581 Alves, V. S., Soares, A. B. A., & Couto, G. S., 2004. Aves marinhas e aquáticas das ilhas do
582 litoral do Estado do Rio de Janeiro. In: Branco, J.O. (Ed.), *Aves marinhas insulares*
583 *brasileiras: bioecologia e conservação*. Editora da UNIVALI, Itajaí, pp. 83-100.
584
- 585 Barbieri, E., Passos, E.D.A., Filippini, A., dos Santos, I.S., Garcia, C.A.B., 2010. Assessment of
586 trace metal concentration in feathers of seabird (*Larus dominicanus*) sampled in the
587 Florianópolis, SC, Brazilian coast. *Environmental Monitoring and Assessment*, 169, 631-
588 638.
589
- 590 Barcellos, C., 1995. *Geodinâmica de Cádmio e Zinco na Baía de Sepetiba*. MSc Thesis, UFRJ,
591 Rio de Janeiro.
592
- 593 Bauer, A. de B., Linhares, B. de A., Nunes, G.T., Costa, P.G., Zebral, Y.D., Bianchini, A.,
594 Bugoni, L., 2024. Temporal changes in metal and arsenic concentrations in blood and
595 feathers of tropical seabirds after one of the largest environmental disasters associated
596 with mining. *Environmental Research*, 248, 118240.
597 <https://doi.org/10.1016/j.envres.2024.118240>
598
- 599 Becker, P. H., Goutner, V., Ryan, P. G., & González-Solís, J., 2016. Feather mercury
600 concentrations in Southern Ocean seabirds: variation by species, site and time.
601 *Environmental Pollution*, 216, 253-263. <https://doi.org/10.1016/j.envpol.2016.05.061>
602
- 603 Berman, E., 1980. *Toxic Metals and Their Analysis*. Heyden and Son, London.
604
- 605 Bighetti, G.P., Padilha, J.A., Cunha, L.S.T., Kasper, D., Malm, O., Mancini, P.L., 2021.
606 Bioaccumulation of mercury is equal between sexes but different by age in seabird (*Sula*
607 *leucogaster*) population from southeast coast of Brazil. *Environmental Pollution*, 285,
608 117222. <https://doi.org/10.1016/j.envpol.2021.117222>
609
- 610 Bighetti, G. P., Padilha, J. A., Cunha, L. S. T., Malm, O., & Mancini, P. L., 2022. Ventral
611 feathers contained the highest mercury level in Brown Booby (*Sula leucogaster*), a
612 pantropical seabird species. *Chemosphere*, 298, 134305.
613
- 614 Branco, J.O., Fracasso, H.A.A., Machado, I.F., Bovendorp, M.S., Verani, J.R., 2005. *Rev. Bras.*
615 *Zool.* 22, 1044.
616
- 617 Burger, J., Schreiber, E.A.E., & Gochfeld, M., 1992. Lead, cadmium, selenium and mercury in
618 seabird feathers from the tropical mid-pacific. *Environmental Toxicology and Chemistry*,
619 11(6), 815-822.
620
- 621 Burger, J., 1993. Metals in avian feathers: bioindicators of environmental pollution. *Reviews of*
622 *Environmental Contamination and Toxicology*, 5, 203-311.
623
- 624 Burger, J., 1995. Heavy metal and selenium levels in feathers of herring gulls (*Larus argentatus*):

- 625 differences due to year, gender, and age at Captree, Long Island. *Environmental*
626 *Monitoring and Assessment*, 38, 37-50.
- 627
- 628 Burger, J., & Gochfeld, M., 2000. Metal levels in feathers of 12 species of seabirds from
629 Midway Atoll in the northern Pacific Ocean. *Science of the Total Environment*, 257(1),
630 37-52.
- 631
- 632 Burger, J., & Gochfeld, M., 2001. Effects of chemicals and pollution on seabirds. In: Schreiber,
633 E. A., Burger, J. (Eds.), *Biology of Marine Birds*. CRC Press, pp. 503-544.
- 634
- 635 Burger, J., & Gochfeld, M., 2000. Metals in Albatross Feathers from Midway Atoll: Influence of
636 Species, Age, and Nest Location. *Environmental Research*, 82, 207-221.
637 <https://doi.org/10.1006/enrs.1999.4015>
- 638
- 639 Castro, Í.B., Machado, F.B., de Sousa, G.T., Paz-Villarraga, C., & Fillmann, G., 2021. How
640 protected are marine protected areas: A case study of tributyltin in Latin America. *Journal*
641 *of Environmental Management*, 278, 111543.
- 642
- 643 Cherel, Y., Connan, M., Jaeger, A., & Richard, P., 2014. Seabird year-round and historical
644 feeding ecology: blood and feather $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ values document foraging plasticity
645 of small sympatric petrels. *Marine Ecology Progress Series*, 505, 267-280.
- 646
- 647 Cipro, C.V.Z., Bustamante, P., Petry, M.V., & Montone, R.C., 2018. Seabird colonies as relevant
648 sources of pollutants in Antarctic ecosystems: Part 1-Trace elements. *Chemosphere*, 204,
649 535-547.
- 650
- 651 Coimbra, K.T.O., Alcântara, E., & de Souza Filho, C.R., 2020. Possible contamination of the
652 Abrolhos reefs by Fundao dam tailings, Brazil–New constraints based on satellite data.
653 *Science of the Total Environment*, 733, 138101.
- 654
- 655 Connolly, R.M., Guest, M.A., Melville, A.J., & Oakes, J.M., 2004. Sulfur stable isotopes
656 separate producers in marine food-web analyses. *Oecologia*, 138, 161-167.
- 657
- 658 Cunha, L.S.T., Torres, J.P.M., Muñoz-Arnanz, J.A., & Jiménez, B., 2012. Evaluation of the
659 possible adverse effects of legacy persistent organic pollutants (POPs) on the Brown
660 Booby (*Sula leucogaster*) along the Brazilian coast. *Chemosphere*, 87(9), 1039-1044.
- 661
- 662 Cunha, L., Alves, V., Rajão, H., & Lanna, A., 2013. Aves do Cagarras Islands Natural
663 Monument. In: *História, pesquisa e biodiversidade do Cagarras Islands Natural*
664 *Monument. Série Livros*, 48, pp. 177-205.
- 665
- 666 de Assis Padilha, J., da Cunha, L.S.T., de Castro, R.M., Malm, O., Dorneles, P.R., 2018.
667 Exposure of Magnificent Frigatebird (*Fregata magnificens*) and Brown Booby (*Sula*
668 *leucogaster*) to Metals and Selenium in Rio de Janeiro State (Brazil) Coastal Waters.
669 *Orbital: The Electronic Journal of Chemistry*, 10(3), 254-261.
- 670

- 671 da Costa, T.F.C., Araujo, J.H., Sichel, S.E., da Silva Pastura, V.F., Motoki, K.F., Barão, L.M., ...
672 & Brunelli, D., 2023. Mapping of surface radiogenic heat production from in situ gamma
673 spectrometry and chemical data of exhumed mantle peridotites at the St. Peter and St.
674 Paul archipelago (equatorial Atlantic). *Applied Radiation and Isotopes*, 192, 110608.
675
- 676 da Silva, D.A.M., Colabuono, F.I., Taniguchi, S., Petry, M.V., & Montone, R.C., 2023.
677 Persistent organic pollutant patterns in seabirds from marine protected areas in the
678 tropical Atlantic Ocean. *Marine Pollution Bulletin*, 186, 114461.
679
- 680 de Seixas, J.T., Mello, S.C.R.P., Faria, A.S., Souza, L.L., & Melo, C., 2020. Análise
681 socioambiental da poluição por esgoto da Baía de Guanabara do Rio de Janeiro. *Revista*
682 *Valore*, 5, 5022.
683
- 684 Dehnhard, N., Achurch, H., Clarke, J., Michel, L.N., Southwell, C., Sumner, M.D., ... &
685 Emmerson, L., 2020. High inter-and intraspecific niche overlap among three
686 sympatrically breeding, closely related seabird species: generalist foraging as an
687 adaptation to a highly variable environment? *Journal of Animal Ecology*, 89(1), 104-119.
688
- 689 Dias, P.S., Cipro, C.V.Z., Taniguchi, S., Montone, R.C., 2013. Persistent organic pollutants in
690 marine biota of Saint Peter and Saint Paul Archipelago, Brazil. *Marine Pollution Bulletin*,
691 74, 435-440. <https://doi.org/10.1016/j.marpolbul.2013.06.025>
692
- 693 Dolan, K.J., Ciesielski, T.M., Lierhagen, S., Eulaers, I., Nygård, T., Johnsen, T.V., ... & Jaspers,
694 V.L.B., 2017. Trace element concentrations in feathers and blood of Northern goshawk
695 (*Accipiter gentilis*) nestlings from Norway and Spain. *Ecotoxicology and Environmental*
696 *Safety*, 144, 564-571.
697
- 698 Dolci, N.N., Sá, F., da Costa Machado, E., Krul, R., & Rodrigues Neto, R., 2017. Trace elements
699 in feathers and eggshells of Brown Booby *Sula leucogaster* in the Marine National Park
700 of Currais Islands, Brazil. *Environmental Monitoring and Assessment*, 189, 1-14.
701
- 702 Dorneles, P.R., Lailson-Brito, J., Dos Santos, R.A., da Costa, P.A.S., Malm, O., Azevedo, A.F.,
703 & Torres, J.P.M., 2007. Cephalopods and cetaceans as indicators of offshore
704 bioavailability of cadmium off Central South Brazil Bight. *Environmental Pollution*,
705 148(1), 352-359.
706
- 707 Dorneles, P.R., Lailson-Brito, J., Fernandez, M.A., Vidal, L.G., Barbosa, L.A., Azevedo, A.F., ...
708 & Malm, O., 2008. Evaluation of cetacean exposure to organotin compounds in Brazilian
709 waters through hepatic total tin concentrations. *Environmental Pollution*, 156(3), 1268-
710 1276.
711
- 712 Dorneles, P.R., Schilithz, P.F., Paiva, T.D.C., Flach, L., Barbosa, L.A., Domit, C., ... & Lailson-
713 Brito, J., 2020. Total tin (TSn) biomagnification: Evaluating organotin trophic flow and
714 dispersion using hepatic TSn concentrations and stable isotope (C, N) data of nektonic
715 organisms from Brazil. *Marine Environmental Research*, 161, 105063.
716

- 717 Einoder, L.D., MacLeod, C.K., & Coughanowr, C., 2018. Metal and isotope analyses of bird
718 feathers in a contaminated estuary reveals bioaccumulation, biomagnification, and
719 potential toxic effects. *Archives of Environmental Contamination and Toxicology*, 75,
720 96-110.
721
- 722 El Hanafi, K., Pedrero, Z., Ouerdane, L., Marchán Moreno, C., Queipo-Abad, S., Bueno, M.,
723 Pannier, F., Corns, W.T., Cherel, Y., Bustamante, P., Amouroux, D., 2022. First Time
724 Identification of Selenoneine in Seabirds and Its Potential Role in Mercury
725 Detoxification. *Environmental Science & Technology*, 56, 3288-3298.
726 <https://doi.org/10.1021/acs.est.1c04966>
727
- 728 Fan, C., Yu, B., & Chen, D., 2009. Effects of different sources and levels of selenium on
729 performance, thyroid function and antioxidant status in stressed broiler chickens.
730 *International Journal of Poultry Science*, 8(6), 583-587.
731
- 732 Finger, A., Lavers, J.L., Orbell, J.D., Dann, P., Nuggeoda, D., & Scarpaci, C., 2016. Seasonal
733 variation and annual trends of metals and metalloids in the blood of the Little Penguin
734 (*Eudyptula minor*). *Marine Pollution Bulletin*, 110, 261-273.
735 <https://doi.org/10.1016/j.marpolbul.2016.06.055>
736
- 737 Frouin, H., Pelletier, E., Lebeuf, M., Saint-Louis, R., & Fournier, M., 2010. Toxicology of
738 organotins in marine organisms: a review. *Organometallic Compounds: Preparation*.
739
- 740 Fuentes, M.M.P.B., Wildermann, N., Gandra, T.B.R., Domit, C., 2020. Cumulative threats to
741 juvenile green turtles in the coastal waters of southern and southeastern Brazil.
742 *Biodiversity and Conservation*, 29, 1783-1803. [https://doi.org/10.1007/s10531-020-](https://doi.org/10.1007/s10531-020-01964-0)
743 [01964-0](https://doi.org/10.1007/s10531-020-01964-0)
744
- 745 Gama, I.H., de Almeida, M.G., Rangel, T.P., Marques, J.S., de Oliveira, B.C., Araújo, B.F., ... &
746 de Rezende, C.E., 2022. Metals and organic matter baselines in sediments in a cross-shelf
747 gradient at Abrolhos Bank, SW Atlantic. *Science of the Total Environment*, 802, 149867.
748
- 749 Gilmour, M.E., Hudson, S.T., Lamborg, C., Fleishman, A.B., Young, H.S., & Shaffer, S.A.,
750 2019. Tropical seabirds sample broadscale patterns of marine contaminants. *Science of*
751 *the Total Environment*, 691, 631-643.
752
- 753 Guruge, K.S., Tanabe, S., Iwata, H., Taksukawa, R., & Yamagishi, S., 1996. Distribution,
754 biomagnification, and elimination of butyltin compound residues in common cormorants
755 (*Phalacrocorax carbo*) from Lake Biwa, Japan. *Archives of Environmental*
756 *Contamination and Toxicology*, 31, 210-217.
757
- 758 Herbst, D.F., Gerhardinger, L.C., Vila-Nova, D.A., de Carvalho, F.G., Hanazaki, N., 2020.
759 Integrated and deliberative multidimensional assessment of a subtropical coastal-marine
760 ecosystem (Babitonga bay, Brazil). *Ocean and Coastal Management*, 196, 105279.
761 <https://doi.org/10.1016/j.ocecoaman.2020.105279>
762

- 763 Jerez, S., Motas, M., Palacios, M.J., Valera, F., Cuervo, J.J., & Barbosa, A., 2011. Concentration
764 of trace elements in feathers of three Antarctic penguins: geographical and interspecific
765 differences. *Environmental Pollution*, 159(10), 2412-2419.
766
- 767 Klein, R., Bartel-Steinbach, M., Koschorreck, J., Paulus, M., Tarricone, K., Teubner, D.,
768 Wagner, G., Weimann, T., & Veith, M., 2012. Standardization of egg collection from
769 aquatic birds for biomonitoring-a critical review. *Environmental Science & Technology*,
770 46(10), 5273-5284.
- 771 Kopp, D., Lefebvre, S., Cachera, M., Villanueva, M.C., & Ernande, B., 2015. Reorganization of
772 a marine trophic network along an inshore–offshore gradient due to stronger pelagic–
773 benthic coupling in coastal areas. *Progress in Oceanography*, 130, 157-171.
774
- 775 Lailson-Brito, J., Dorneles, P.R., Azevedo-Silva, C.E., Azevedo, A.F., Vidal, L.G., Zanelatto,
776 R.C., ... & Malm, O., 2010. High organochlorine accumulation in blubber of Guiana
777 dolphin, *Sotalia guianensis*, from Brazilian coast and its use to establish geographical
778 differences among populations. *Environmental Pollution*, 158(5), 1800-1808.
779
- 780 Lavers, J.L., Humphreys-Williams, E., Cramer, N.J., & Bond, A.L., 2020. Trace element
781 concentrations in feathers from three seabird species breeding in the Timor Sea. *Marine
782 Pollution Bulletin*, 151, 110876.
783
- 784 Lerma, M., Castillo-Guerrero, J.A., García-Hernández, J., & Fernández, G., 2020. Zinc
785 concentrations in Blue-footed booby (*Sula nebouxii*) eggs, nestlings, and adults. *Journal
786 of Sea Research*, 165, 101952.
787
- 788 Lerma, M., Dehnhard, N., Luna-Jorquera, G., Voigt, C.C., & Garthe, S., 2020. Breeding stage,
789 not sex, affects foraging characteristics in masked boobies at Rapa Nui. *Behavioral
790 Ecology and Sociobiology*, 74, 1-16.
791
- 792 Maciel, D.C., Castro, Í.B., de Souza, J.R.B., Yogui, G.T., Fillmann, G., & Zanardi-Lamardo, E.,
793 2018. Assessment of organotins and imposex in two estuaries of the northeastern
794 Brazilian coast. *Marine Pollution Bulletin*, 126, 473-478.
795
- 796 Mancini, P.L., Hobson, K.A., & Bugoni, L., 2014. Role of body size in shaping the trophic
797 structure of tropical seabird communities. *Marine Ecology Progress Series*, 497, 243-257.
798 trophic variation in Brown Booby (*Sula leucogaster*) from the Southwestern Atlantic. *Marine
799 Biology*, 170(1), 1.
800
- 801 Mancini, P. L., Valim, E. E. M., de Barros Bauer, A., & Fischer, L. G., 2023. Intraspecific
802 trophic variation in Brown Booby (*Sula leucogaster*) from the Southwestern Atlantic.
803 *Marine Biology*, 170(1), 1.
804
- 805 Mellink, E., Domínguez, J., Luévano, J., & Domínguez, J., 2001. Diet of Eastern Pacific Brown
806 Boobies *Sula leucogaster brewsteri* on Isla San Jorge, north-eastern Gulf of California,
807 and an April comparison with diets in the Middle. *Marine Ornithology*, 28, 23-28.
808

- 809 Mello, F. V., Cunha, L. S. T., & Torres, J. P. M., 2012. Persistent organic pollutants in the diet of
810 Brown Boobies (*Sula leucogaster*) from three different archipelagos of Brazil.
811 Organohalogen Compounds, 74, 524-527.
812
- 813 Moura, J. F., Tavares, D. C., Lemos, L. S., Acevedo-Trejos, E., Saint’Pierre, T. D., Siciliano, S.,
814 & Merico, A., 2018. Interspecific variation of essential and non-essential trace elements
815 in sympatric seabirds. Environmental Pollution, 242, 470-479.
816
- 817 Navarro, J., Moreno, R., Braun, L., Sanpera, C., & Hennicke, J. C., 2014. Resource partitioning
818 between incubating and chick-rearing Brown Boobies and red-tailed tropicbirds on
819 Christmas Island. Zoological Studies, 53, 1-6.
820
- 821 Neto, J. A. B., Gingele, F. X., Leipe, T., & Brehme, I., 2006. Spatial distribution of heavy metals
822 in surficial sediments from Guanabara Bay: Rio de Janeiro, Brazil. Environmental
823 Geology, 49, 1051-1063.
824
- 825 Nunes, G. T., Bertrand, S., & Bugoni, L., 2018. Seabirds fighting for land: phenotypic
826 consequences of breeding area constraints at a small remote archipelago. Scientific
827 Reports, 8(1), 665.
828
- 829 Nunes, G. T., Efe, M. A., Barreto, C. T., Gaiotto, J. V., Silva, A. B., Vilela, F., ... & Bugoni, L.,
830 2022. Ecological trap for seabirds due to the contamination caused by the Fundão dam
831 collapse, Brazil. Science of the Total Environment, 807, 151486.
832
- 833 Ohlendorf, H. M., 2002. Ecotoxicology of selenium. In: Hoffman, D. J., Rattner, B. A., Burton
834 Jr., G. A., & Cairns Jr., J. (Eds.), Handbook of Ecotoxicology. CRC Press, Boca Raton,
835 pp. 489-524.
836
- 837 Pacyna, A.D., Jakubas, D., Ausems, A.N.M.A., Frankowski, M., Polkowska, Ż., & Wojczulanis-
838 Jakubas, K., 2019. Storm petrels as indicators of pelagic seabird exposure to chemical
839 elements in the Antarctic marine ecosystem. Science of the Total Environment, 692, 382–
840 392. <https://doi.org/10.1016/j.scitotenv.2019.07.137>
841
- 842 Padilha, J. A., Castro, R.M., Cunha, L.S.T., Paiva, T. C., Malm, O., & Dorneles, P.R., 2013.
843 Concentrações de cádmio, estanho e cobre em penas de atobás (*Sula leucogaster*) e
844 fragatas (*Fregata magnificens*) do Cagarras Islands Natural Monument, Rio de Janeiro,
845 Brasil. Natural Resources, 3, 14-14. [https://doi.org/10.6008/ESS2237-](https://doi.org/10.6008/ESS2237-9290.2013.002.0009)
846 [9290.2013.002.0009](https://doi.org/10.6008/ESS2237-9290.2013.002.0009)
847
- 848 Padilha, J. A., Carvalho, G. O., Espejo, W., Souza, J. S., Pizzochero, A. C., Cunha, L. S. T., ... &
849 Dorneles, P. R., 2021. Factors that influence trace element levels in blood and feathers of
850 *Pygoscelis penguins* from South Shetland Islands, Antarctica. Environmental Pollution,
851 284, 117209.
852
- 853 Pereira, E., Baptista-Neto, J. A., Smith, B. J., & Mcallister, J. J., 2007. The contribution of heavy
854 metal pollution derived from highway runoff to Guanabara Bay sediments: Rio de

- 855 Janeiro/Brazil. *Anais da Academia Brasileira de Ciências*, 79, 739-750.
856
- 857 Pizzochero, A. C., Michel, L. N., Chenery, S. R., McCarthy, I. D., Vianna, M., Malm, O., ... &
858 Dorneles, P. R., 2018. Use of multielement stable isotope ratios to investigate ontogenetic
859 movements of *Micropogonias furnieri* in a tropical Brazilian estuary. *Canadian Journal of*
860 *Fisheries and Aquatic Sciences*, 75(6), 977-986.
861
- 862 Polito, M. J., Brasso, R. L., Trivelpiece, W. Z., Karnovsky, N., Patterson, W. P., & Emslie, S. D.,
863 2016. Differing foraging strategies influence mercury (Hg) exposure in an Antarctic
864 penguin community. *Environmental Pollution*, 218, 196-206.
865
- 866 Riley, W. J., 2019. A history of the rubidium frequency standard. *IEEE UFFC-S History*, 2, 1-10.
867
- 868 Hobson, K. A., Blight, L. K., & Arcese, P., 2015. Human-induced long-term shifts in gull diet
869 from marine to terrestrial sources in North America's coastal Pacific: more evidence from
870 more isotopes ($\delta^2\text{H}$, $\delta^{34}\text{S}$). *Environmental Science & Technology*, 49(18), 10834-10840.
871
- 872 Roman, L., Kastury, F., Petit, S., Aleman, R., Hardesty, B. D., & Wilcox, C., 2023. Nutrients and
873 seabird biogeography: Feather elements differ among oceanic basins in the Southern
874 Hemisphere, reflecting bird size, foraging range and nutrient availability in seawater.
875 *Global Ecology and Biogeography*, 32(4), 495-510.
876
- 877 Sasanami, T. (Ed.), 2017. *Avian Reproduction: From Behavior to Molecules*. Springer, Tokyo.
878
- 879 Schilithz, P. F., Dorneles, P. R., & Brito, J. L., 2013. Exposição de cetáceos a compostos
880 butilêstânicos: uma revisão. *Oecologia Australis*, 17(3), 429-448.
881
- 882 Signa, G., Mazzola, A., & Vizzini, S., 2021. Seabird influence on ecological processes in coastal
883 marine ecosystems: An overlooked role? A critical review. *Estuarine, Coastal and Shelf*
884 *Science*, 250, 107164.
885
- 886 Van Weerelt, M., Cunha, L., Dorneles, P. R., Padilha, J., Ormond, J., Torres, F., Torres, J. P.,
887 Batista, D., Nudi, A., & Wagener, A., 2013. Monitoramento da qualidade das águas e dos
888 poluentes no MoNa das Ilhas Cagarras e entorno. *História, pesquisa e biodiversidade do*
889 *Monumento Natural das Ilhas Cagarras, Rio de Janeiro. Museu Nacional*, pp. 229-243.
890
- 891 Varala, R., & Rao, K. S., 2015. Cesium salts in organic synthesis: A review. *Current Organic*
892 *Chemistry*, 19(13), 1242-1274.
893
- 894 Walsh, P. M., 2018. The use of seabirds as monitors of heavy metals in the marine environment.
895 In: Furness, R. W., & Rainbow, P. S. (Eds.), *Heavy Metals in the Marine Environment*.
896 CRC Press, pp. 183-204.
897
- 898 Wróblewski, M., Wróblewska, J., Nuskiewicz, J., Pawłowska, M., Wesołowski, R., & Woźniak,
899 A., 2023. The role of selected trace elements in oxidoreductive homeostasis in patients
900 with thyroid diseases.

Journal Pre-proof

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Journal Pre-proof