



Impact of chronic cadmium exposure at environmental dose on escape behaviour in sea bass (*Dicentrarchus labrax* L.; Teleostei, Moronidae)

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Cadmium exposure involved a significant bioaccumulation in fish scales, slight damage to the lateral line system and a significant decrease in fish escape behaviour.

Abstract

The effect of chronic exposure to a low concentration ($0.5 \mu\text{g l}^{-1}$) of cadmium ions was investigated on escape behaviour of sea bass, *Dicentrarchus labrax*, using video analysis. Observations were also performed on the microanatomy of lateral system neuromasts. When fish were exposed for 4 h per day over 8 days to the cadmium ions, most of both types of neuromasts observed remained intact. However, some of them presented damaged sensory maculae. Whereas before cadmium exposure, fish responded positively to nearly all the lateral system stimulations, after exposure they decreased by about 10% their positive responses to stimulations. From the 15th day after the beginning of cadmium exposure, neuromasts presented progressively less damage, cadmium accumulation in gills and scales decreased significantly and fish escape behaviour had recovered. This study presents a new concept in ecotoxicology: using behavioural change to reveal the effects of pollution levels, scarcely detectable by currently used techniques (physiological responses).

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Keywords: Fish lateral system; Neuromast; Chronic cadmium exposure; Escape behaviour; Bioaccumulation

1. Introduction

In French coastal ecosystems, a significant number of contaminants are present, particularly metal ions (Chiffolleau et al., 1999; Boutier et al., 2000). Yet, many such affected coastal areas constitute nurseries, essential habitats for juvenile fish. Originating from the discharge of waste, specialized industries or mining activities, these pollutants are thought to be responsible for the reduction of fish resources in estuaries and coastal waters (Cole, 1979; Waldichuk, 1979). In order to protect aquatic wildlife, it is therefore necessary to determine contamination levels. In this way, water quality criteria have been established, based primarily on bioaccumulation

in mollusc and crustacean organs (Cole, 1979; Waldichuk, 1979 for review) or on acute lethality tests (Eisler and Hennekey, 1977; Voyer et al., 1979; Hollis et al., 1999, 2000). However, one aim of aquatic toxicology is to reveal the subtler and more insidious changes induced by pollutants on aquatic organisms and their environment (Larsson et al., 1985). According to Atchinson et al. (1987), tests based on standard acute-toxicity assays (LC₅₀: concentration of the toxic substance that is lethal to 50% of individuals after a specific exposure time) and chronic-toxicity tests based on either full or partial life cycles, or on early life stages (LOEC: lower observed effect concentration), are less sensitive than behavioural studies. Behaviour is obviously a very important organism-level response that is the result of molecular, physiological and ecological processes (Scott and Sloman, 2004; Weis, 2004). According to Scott and Sloman (2004), behaviour may hence be useful for studying environmental-pollutant

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115 effects because it can provide a bioassay to determine an “ecological death” that may occur after much lower exposures to
116 the toxicant. These authors also argued that fish are an excellent
117 model in this regard, since many ecologically relevant fish
118 behaviours are easily observed and quantified in a controlled
119 setting. Even if fish are not overtly harmed by a contaminant,
120 they may be unable to function in an ecological context if their
121 normal behaviour is altered (Bruslé and Quignard, 2004).
122 Indeed, behavioural reactions may occur at concentrations sig-
123 nificantly less than those producing gross physiological effects
124 or death (Jensen and Bro-Rasmussen, 1992; Baker and Mont-
125 gomery, 2001). Behavioural consequences may include: (1)
126 impaired predatory behaviour resulting in poor diet, which
127 can cause reduced growth and longevity; (2) altered pred-
128 ator-avoidance behaviour; or (3) impaired schooling leading
129 to increased mortality and/or altered reproductive function
130 (Weis, 2004). All these behavioural events form an important
131 part of a successful adaptive life history strategy. Altered
132 behaviours caused by exposure to pollutants may hence cause
133 serious risks to the success of fish populations and disrupt
134 aquatic communities (Scott and Sloman, 2004). These authors
135 underline that more research is required concerning the impact
136 of chronic exposure to low toxicant concentrations on fish
137 behaviours.

138 In fish, the lateral line system is involved in many behav-
139 ioural events such as predator and prey detection (Hoekstra
140 and Janssen, 1986; Montgomery, 1989; Janssen et al., 1999),
141 rheotaxis (Montgomery et al., 1997; Northcutt, 1997; Baker
142 and Montgomery, 1999a,b; Coombs et al., 2001), obstacle
143 avoidance (Blaxter and Batty, 1985) and intraspecific interac-
144 tions (Partridge and Pitcher, 1980). The functional units of this
145 lateral line system are mechanoreceptors, the neuromasts, dis-
146 tributed on the head, trunk and tail of the fish (Coombs et al.,
147 1989). The morphological and functional integrity of these
148 mechanoreceptors thus appears indispensable for the existence
149 and the survival of a fish species in an ecosystem. Among
150 metal ions, cadmium is considered as the most toxic ion after
151 mercury because concentrations leading to death are much
152 lower than for other metal ions (Eisler and Hennekey, 1977).
153 Also, in contrast to several metal ions (cobalt, copper, iron,
154 zinc, etc.), the cadmium ion has no known metabolic role
155 and does not seem to be biologically essential or beneficial
156 to metabolism (Friberg et al., 1974; Bryan, 1979). Given
157 that cadmium is a calcium antagonist at the level of the gills
158 (Verbost et al., 1987, 1988), and that calcium ions play a pre-
159 ponderant role in signal transduction mechanisms in neuro-
160 mast hair cells in the fish lateral line system (Sand, 1975;
161 Hudspeth and Corey, 1977; Jørgensen, 1984), cadmium ions
162 might affect mechanoreception and thereby alter the behaviour
163 of fish exposed to them. Several studies carried out mainly in
164 freshwater, reported the impact of metal ions on the fish
165 sensory system and the consequences for behaviour (see
166 Atchinson et al., 1987 for review). For example, Baker and
167 Montgomery (2001) showed that cadmium ions were respon-
168 sible for impaired olfactory function and altered rheotaxis be-
169 haviour associated with damage to the lateral line system in
170 freshwater fish. Very few studies, to date, have been performed

171 on the effect of cadmium exposure on marine fish behaviour.
172 A previous study (Faucher et al., 2006) showed that when
173 sea bass were submitted to acute cadmium exposure at low
174 concentration (4 h at 0.5 $\mu\text{g l}^{-1}$, which represents the maximal
175 cadmium concentration encountered in contaminated French
176 estuaries), neither alteration in neuromast tissue, nor any be-
177 havioural modification could be detected. In contrast, after
178 an acute cadmium exposure at 10-fold higher concentration,
179 severe neuromast tissue damage was observed, contributing
180 to a decrease in their escape behaviour by about 56%. This es-
181 cape behaviour is induced by the detection of hydrodynamic
182 stimuli from predator displacements that act on fish lateral
183 line system in association with their inner ear (Coombs
184 et al., 1989).

185 The aim of this study has been to determine the impact of
186 chronic low-concentration cadmium exposure on the fish es-
187 cape response. A major innovation in the present study is
188 the determination of cadmium effects over a long time span
189 (chronic exposure) at a concentration close to that measured
190 in the fish's more polluted habitats (0.5 $\mu\text{g l}^{-1}$) on the lateral
191 line system of sea bass *Dicentrarchus labrax*. In addition,
192 this work combines for the first time data concerning accumu-
193 lation of cadmium in tissues, sensory tissue damage on both
194 types of neuromasts of the lateral line and consequences on
195 fish-escape behaviour.

2. Materials and methods

2.1. Experimental fishes

200 Experimental sea bass were obtained from the Ferme des Baleines, Ile de Ré,
201 France. They were placed in 400-l seawater tanks at constant temperature (18 °C)
202 for three months with a natural photoperiod. They were fed twice a week with
203 commercial pellets. The experiments took place between June and September
204 2005. They were carried out in two identical sets realized simultaneously, each
205 one consisting of twelve fish (about 6 g and 7 cm standard length).

2.2. Experimental set up

206 Experiments took place in two 40 l-tanks (100 × 40 × 10 cm) of seawater
207 at constant temperature (18 °C). The photoperiod was controlled (14-L: 10-D)
208 and an automatic feeder delivered food each day, about thirty minutes after the
209 beginning of the light phase. Fish were placed for one week in the tanks before
210 the beginning of the experiment.

211 In order to test the function or the dysfunction of their lateral line system,
212 the same set up was used as that previously employed (Faucher et al., 2006) to
213 study the impact of acute cadmium exposure on the trunk lateral line neuro-
214 masts and consequences on the fish behaviour. A pipette connected to
215 a hand-operated syringe was used to inject a water jet between the water surface
216 and the base of the tank when fish swam in the vicinity (about 5 cm) of the pi-
217 pette. Each day, three stimulations (injection of a water jet) were performed and
218 the fish responses were recorded with an analog video camera (SONY CCD-
219 VX1E Handicam Pro, 25-frames s^{-1}) positioned at a height of ~1 m above
220 the water surface. The lateral line system of the fish was considered as func-
221 tional when the water jet stimulation provoked a sudden escape reaction, char-
222 acterised by the bending of the fish's body into a C-like shape, followed by an
223 abrupt swimming acceleration away from the initial location (see Faucher et al.,
224 2006). Such a response was counted as a positive response and noted as 1. Im-
225 mobility or a constant swimming velocity was noted as null response and noted
226 as 0 (see Faucher et al., 2006 for illustrations). Each day, the number of positive
227 responses out of the three expected was calculated. In this way, sea bass were
228 recorded each day under control conditions for three weeks.

229

2.3. Cadmium exposure

To reveal the impact of chronic cadmium exposure on the lateral line system through its consequences on escape responses, two sets of experiments with two separate groups of fish were performed simultaneously. First, fish response was recorded under control conditions every day for 3 weeks. Then, each day for 8 days, sea bass were collected and placed for 4 h (the time needed for cadmium adsorption onto a particle, Chiffolleau et al., 1999) in a 10 l-tank of seawater to which $0.5 \mu\text{g l}^{-1}$ cadmium ($\text{Cd}(\text{NO}_3)_2$, Merck, cadmium standard solution 1000 mg l^{-1} in nitric acid 0.5 M) had been added. This intermittent exposure for 4 h per day during 8 consecutive days in another tank was chosen instead of 8 continuous days in the experimental tank to avoid the risk that cadmium ions might destroy the biological filtration system in the experimental tank. The concentration tested represents the maximal cadmium concentration encountered in highly polluted estuaries such as the Gironde, Scheldt and Hudson estuaries (Klinkhammer and Bender, 1981; Elbaz-Poulichet et al., 1987; Jouanneau et al., 1990). Such an intermittent cadmium exposure may occur in natural estuaries when fish cross through maximum turbidity zone where sediments are known to adsorb contaminated particles as metal ions (Jouanneau et al., 1990; Chiffolleau et al., 1999). Fish were then placed back in their experimental tank. Their swimming behaviour was normal. After allowing several hours for recovery, the time required for the entire disappearance of the stress caused by cadmium exposure, the sea bass responses to the three daily stimulations by the water jet were recorded, every day until the restoration of a normal behaviour.

To evaluate the stress caused by daily manipulations, two placebo treatments (4-h baths in seawater without cadmium) were performed one week apart during the three weeks of recording under control conditions.

230

2.4. Water contamination analyses

To determine the cadmium concentration to which the fish were really exposed each day during the 4 h of exposure, sample water was collected at 0, 2 and 4 h after the addition of $0.5 \mu\text{g l}^{-1}$ cadmium, on the 6th, 7th and 8th day of exposure. Samples collected at each time (0, 2 and 4 h) from the 6th, 7th and 8th day were taken to constitute a “0 h” sample, a “2 h” sample and a “4 h” sample. Analyses of cadmium concentration in these three water samples were performed in the Institut Pasteur of Lille. Water samples were filtered through $0.45 \mu\text{m}$, and cadmium concentration measurements were realized using ion-adsorption onto resin.

231

2.5. Metal analyses

To evaluate metal tissue contamination, gills and scales covered by mucus were collected from three fish sampled simultaneously: in control conditions, and on the 3rd day, the 8th day, the 15th day (8 days of exposure followed by 7 days of depuration) and the 21st day (8 days of exposure followed by 13 days of depuration) after the beginning of cadmium exposure. Fish gills and scales were chosen because they are in direct contact with pollutants and thus might represent short-term biomarkers of contamination compared to the long-term contamination biomarkers (liver, kidney) usually used. Tissue samples were dried at 50°C for 2 days. Dry samples were weighed and digested for 2 days in 5 ml concentrated (14 N) nitric acid at 80°C until the digestion was completed, then heated to dryness. A total of 2 ml 0.3 N nitric acid was then added. Three analytical blanks were prepared in a similar manner without samples to check for possible contamination. Corrections were applied wherever necessary. The digestion procedure was also applied to standards (200 mg of DORM-1: dogfish muscle powder) of Cd concentrations to check the spectrophotometer calibration. Metal concentrations were then measured with an atomic absorption spectrophotometer (HITACHI, Polarized Zeeman Atomic Absorption Spectrophotometer Z-5000). The results for the standard reference materials were in good agreement with the certified values reported (2.56% deviation). All glassware was carefully decontaminated with acid (3.5% nitric acid, Merck +5% fuming hydrochloric acid, Merck) for 24 h and was then copiously rinsed with distilled water before any new use. The experiment was conducted in triplicate and the reported values are an average of the three values measured in tissue of the three fish collected in control

conditions, on the 3rd day, the 8th day, the 15th day and the 21st day after the beginning of cadmium exposure. Results are expressed as the metal concentration in μg reported on a dry weight basis.

2.6. Observation of lateral line system tissue status

For each set of experiments, to verify the tissue status of both superficial and canal lateral line system neuromasts in sea bass after cadmium exposure, four batches of twelve fish (three taken on each date) were collected. The first batch, sampled on the 3rd day after cadmium exposure, the second at the end of the period of chronic cadmium exposure (the 8th day), the third batch on the 15th day, and the last batch sampled on the 21st day after the beginning of chronic cadmium exposure. The neuromast tissue status of cadmium-exposed fish was compared with that of two control fish collected from each experimental tank after three weeks of recording under control conditions. Prior to sacrifice, collected fish were anaesthetised with 75 mg l^{-1} MS-222 (3-amino-benzoic acid ethyl, Sigma) for about 15 min. The whole of trunk lateral line mechanoreceptors was then sampled. Tissue samples were prepared for scanning electron microscope observations following the same set up as in Faucher et al. (2003).

2.7. Statistical analyses

To estimate damage caused by cadmium to both types of neuromasts, the average number of superficial and canal neuromasts damaged per scale was counted. Data obtained were then compared between fish exposed to cadmium and control fish using χ^2 -tests.

Cadmium accumulation data in fish gills and scales were examined with a two-factor analysis of variance (ANOVA) with organ (gills and scales) and exposure time (control, 3, 8, 15 and 21 days) as dependent variables, after finding homogeneity of variances ($p > 0.05$). Significant main effects were followed up with Student-Newman–Keuls post-hoc tests.

Behavioural responses to water jet stimulations were analysed following the same data treatment as in Faucher et al. (2006). Data obtained are expressed as the average percentage of positive responses $P_k \pm \text{SD}$ (standard deviation of the mean). The number of data points obtained each day with the two sets of experiments is indicated between brackets. The percentages obtained before and after cadmium exposure were compared using χ^2 -test.

All statistical analyses were performed with the statistical softwares XISTAT-Pro 6.0 and Minitab 13.0. The level of significance was set at $p < 0.05$.

3. Results

Water analyses showed that sea bass were exposed to an average concentration of dissolved cadmium ions in seawater of $0.48 \pm 0.18 \mu\text{g l}^{-1}$ ($n = 3$) during the four hours of exposure. For all the duration of the experiment, fish mortality was null.

3.1. Effects of cadmium exposure on sea bass lateral line system

Compared to control fish (Fig. 1A,B), the majority of superficial (Fig. 1C) and canal (Fig. 1D) neuromasts of sea bass exposed to $0.5 \mu\text{g l}^{-1}$ for three days, did not present any apparent tissue damage. The majority of them possessed intact sensory maculae: hair bundles of subjacent sensory hair cells were well developed. However, some seemed to be slightly damaged (Fig. 1E,F): their hair cell bundles were sparse, shortened, and sometimes not visible. Although the percentage of superficial neuromasts damaged (8.6%, $n = 35$) was not significantly different from that in control fish for which all neuromasts were intact (0% destroyed, $n = 19$, $\chi^2 = 2.515$,

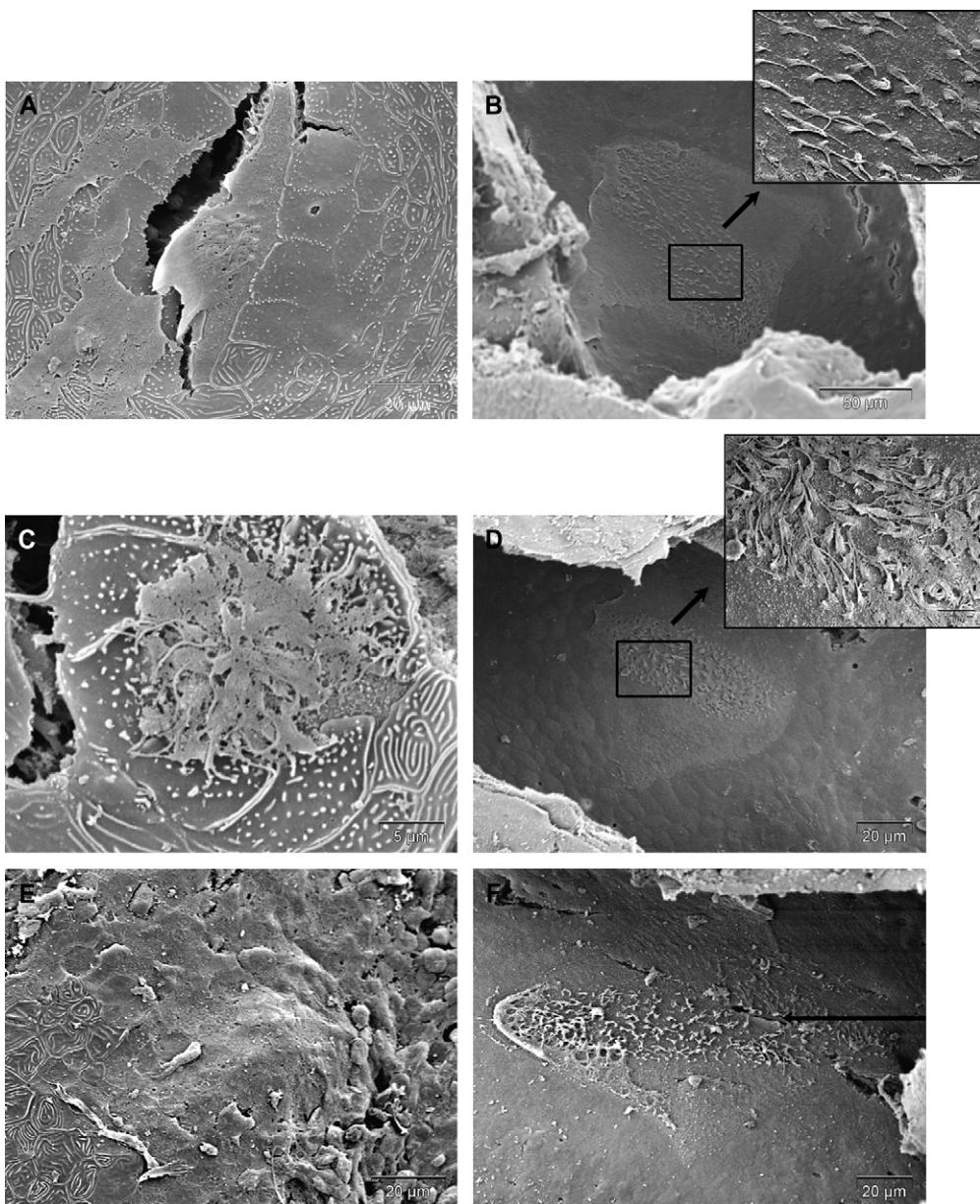


Fig. 1. Scanning electron micrographs showing the effect of low-concentration cadmium exposure (4 h per day at $0.5 \mu\text{g l}^{-1}$) for 3 days on tissue status of both types of neuromasts from the trunk lateral line of sea bass, *Dicentrarchus labrax* (L.). Intact superficial (A) and canal (B) neuromasts observed in a control fish. Superficial neuromast is still covered by its cupula (A) whereas its absence on canal neuromast reveals hair bundles (insert in B). The crushed appearance of superficial neuromast cupula is due to a manipulation artefact. C, D. Three days after chronic cadmium exposure, the majority of superficial (C) and canal (D) neuromasts appeared still intact. Hair bundles present within sensory maculae were normal (insert in D). However, some superficial (E) and canal (F) neuromasts were damaged: their sensory maculae presented hair bundles shortened, sparse (black arrow in F) or not visible (E).

$p = 0.113$), the percentage of canal neuromasts damaged (27.8%, $n = 18$) after three days of cadmium exposure was significantly greater than in control fish (0%, $n = 12$, $\chi^2 = 4.623$, $p = 0.032$). At the end of the period of chronic cadmium exposure (8 days), fish presented mainly intact superficial (Fig. 2A) and canal neuromasts (Fig. 2B). However, some of them were nevertheless damaged as illustrated by the Fig. 2C,D. The percentage of superficial neuromasts altered (6.3%, $n = 24$) was not significantly different from that in control fish (0%, $n = 19$, $\chi^2 = 1.709$, $p = 0.191$). In contrast, the percentage of canal neuromasts damaged (30.8%, $n = 13$) remained higher than that observed in control fish

(0%, $n = 12$, $\chi^2 = 4.719$, $p = 0.030$). Then, 15 days after the beginning of cadmium exposure, superficial (Fig. 3A) and canal (Fig. 3B) neuromasts were still mainly intact. A small percentage of superficial (4.8%, $n = 21$) and canal (22.2%, $n = 9$) neuromasts were once again damaged (Fig. 3C,D) but less markedly than previously. The percentage of neuromasts altered by cadmium exposure was not significant compared to that observed in control fish, whether in the case of superficial ($\chi^2 = 1.267$, $p = 0.260$) or canal neuromasts ($\chi^2 = 3.159$, $p = 0.076$). As before, at the end of the experiment (21 days after the beginning of cadmium exposure), nearly all the neuromasts of each type were intact (Fig. 4A,B).

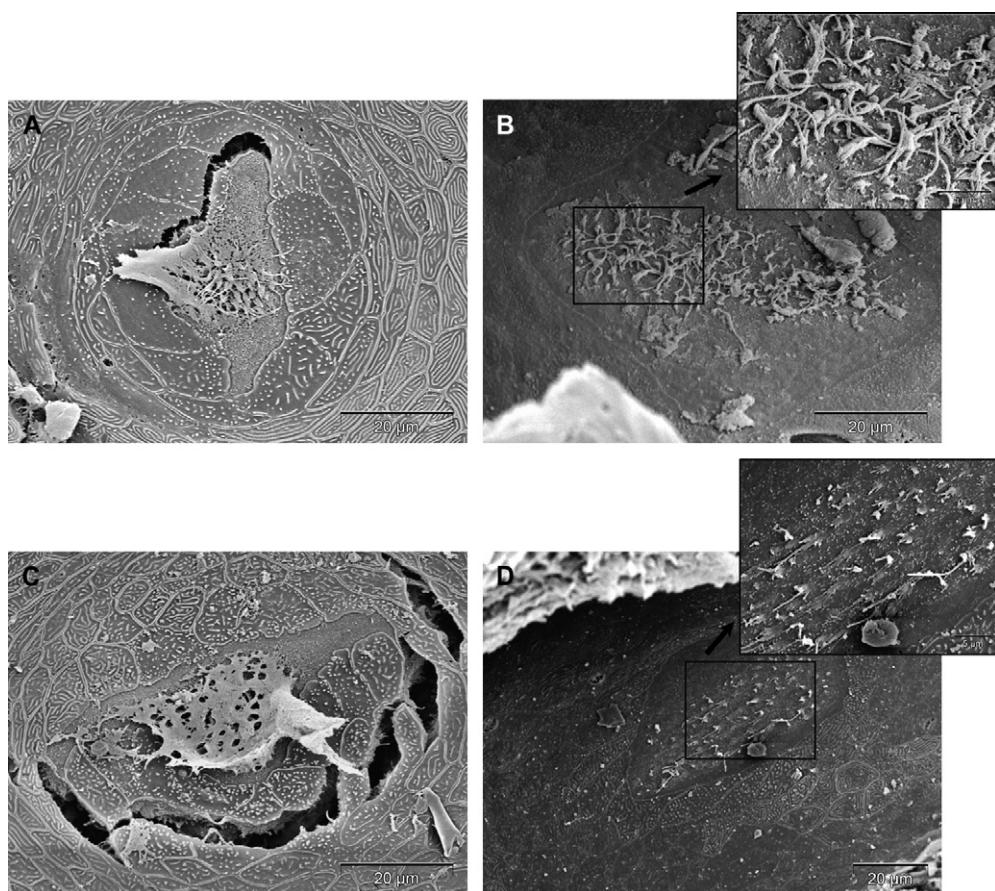


Fig. 2. Scanning electron micrographs showing the effect of low-concentration cadmium exposure ($0.5 \mu\text{g l}^{-1}$ for 4 h per day) for 8 days on tissue status of both types of neuromasts of sea bass trunk lateral line at the end of exposure. The majority of superficial (A) and canal (B) neuromasts presented normal morphology (insert in B): their sensory maculae were similar to those observed in control fish. Nevertheless, some superficial (C) and canal (D) neuromasts were damaged: their sensory maculae presented hair bundles shortened, sparse (insert in D) or even not visible (C).

Nevertheless, some superficial (5.6%, $n = 18$) and canal (16.7%, $n = 6$) neuromasts remained slightly altered (Fig. 4C,D): their hair bundles seemed to be a little sparse or shortened. However, the percentage of damaged neuromasts was not significantly greater than in control fish ($\chi^2 = 1.413$, $p = 0.235$ for superficial and $\chi^2 = 2.094$, $p = 0.148$ for canal neuromasts). It is relevant to note that the percentages of damaged neuromasts and their tissue alteration (hair bundles almost non-visible) were maximal between 3 and 8 days after exposure. Indeed, no significant difference was observed in the case of the percentage of damaged superficial ($\chi^2 = 0.096$, $p = 0.757$, $n = 59$) and canal neuromasts ($\chi^2 = 0.018$, $p = 0.894$, $n = 31$) between 3 and 8 days after the beginning of cadmium exposure.

3.2. Cadmium bioaccumulation

Fig. 5 shows the Cd accumulation as a function of exposure time. For gills (Fig. 5A) as for scales (Fig. 5B), average cadmium concentrations measured in control fish were relatively high: $0.054 \pm 0.032 \mu\text{g g}^{-1}$ of dry weight ($n = 3$) in gills and $0.052 \pm 0.005 \mu\text{g g}^{-1}$ of dry weight ($n = 3$) for scales. In fish gills, cadmium ions did not seem to accumulate significantly

($F_{4,15} = 0.760$, $p = 0.574$, $n = 15$). In contrast, for the whole duration of experiment, the average concentration of cadmium in fish scales was significantly greater than that in gills ($F_{2,30} = 13.811$, $p = 0.001$, $n = 30$). The maximal concentration in cadmium was observed in scales after 3 days of exposure and was $0.147 \pm 0.015 \mu\text{g g}^{-1}$ dry weight ($n = 3$). This concentration of cadmium in scales after 3 days of exposure was significantly higher than that measured in control fish ($t = 5.856$, $p < 0.0001$, $n = 6$), and after 8 days ($t = 4.874$, $p = 0.001$, $n = 6$), 15 days ($t = 4.434$, $p = 0.001$, $n = 6$) and 21 days ($t = 6.329$, $p < 0.0001$, $n = 6$). Then, cadmium concentration in scales tended to decrease to be not significantly different from that measured in control fish ($F_{3,12} = 1.273$, $p = 0.348$, $n = 12$).

3.3. Consequences of cadmium exposure on fish responses to the water jet

During the three weeks of recording under control conditions, sea bass responded positively at $94.05 \pm 8.88\%$ ($n = 42$): they swam away after stimulation by the water jet (Fig. 6). The two placebo treatments realised did not generate any significant behavioural modification in fish. The day of

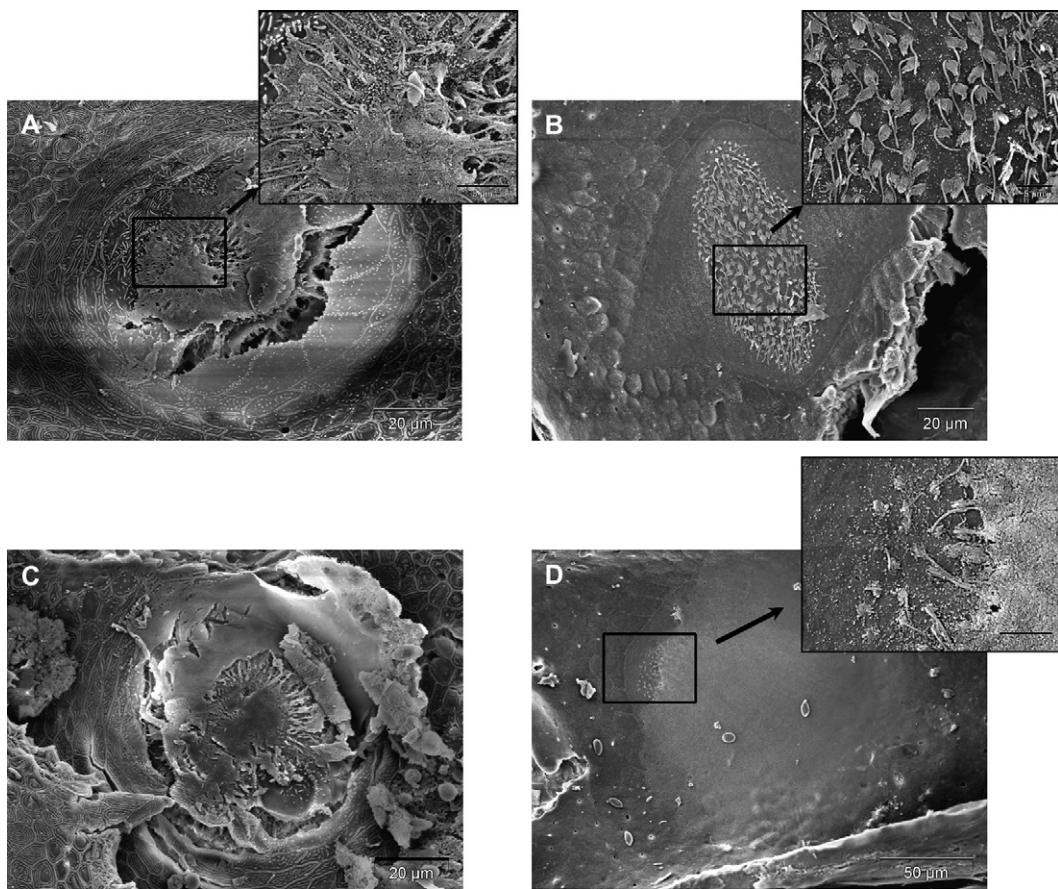


Fig. 3. Scanning electron micrographs showing the effect of chronic low-concentration cadmium exposure ($0.5 \mu\text{g l}^{-1}$ for 4 h per day), 15 days after the beginning of the exposure, on tissue status of both types of neuromasts of sea bass trunk lateral line. Superficial (A) and canal (B) neuromasts were usually intact. Inserts in A and B illustrate details of sensory maculae with normal hair bundles. However, some superficial (C) and canal (D) neuromasts did present altered morphology. Their hair cell bundles seemed to be damaged: they were shortened, sparse (insert in D) or even not visible.

cadmium exposure, the average positive response percentage fell significantly ($\chi^2 = 6.290, p = 0.012$): sea bass responded positively in only $66.67 \pm 0.00\%$ ($n = 2$) of stimulations. From day 1 and during all the cadmium exposure period (8 days), this average positive response percentage progressively decreased from $100.00 \pm 0.00\%$ ($n = 2$) to $66.67 \pm 0.00\%$ ($n = 2$). Until the 15th day (8 days of exposure followed by 8 days of depuration), cadmium-treated fish went on being significantly less reactive to stimulation by the water jet than control fish: they presented an average response percentage of $84.17 \pm 15.22\%$ ($n = 30, \chi^2 = 5.284, p = 0.022$). Then, from the 15th day, fish started to positively respond again to stimulations in $95.56 \pm 7.63\%$ ($n = 15$) of cases. From this day, the average percentage positive response was no longer significantly different from that recorded in control conditions ($94.05 \pm 8.88\%, n = 42, \chi^2 = 0.168, p = 0.682$).

All results obtained in this study are summed up in the Fig. 7. In summary, cadmium exposure involved: (1) a significant cadmium bioaccumulation in scales; (2) slight damages to both types of neuromasts, canal neuromasts being the more altered; and (3) a significant decrease in fish escape behaviour during the time of exposure. After this time, fish tended to restore their escape behaviour in association with a regeneration of neuromasts tissue and a cadmium depuration in gills and scales.

4. Discussion

Many researchers have proposed using behavioural indicators in fish for ecologically relevant monitoring of environmental contamination (reviewed by Atchinson et al., 1987). We have now done this for chronic cadmium exposure in sea bass, *Dicentrarchus labrax*, at the concentration occurring *in situ* in polluted French estuaries.

In a previous study, Faucher et al. (2006) showed that 48 h after 4 h $0.5 \mu\text{g l}^{-1}$ acute cadmium exposure, neither type of neuromasts presented any apparent tissue damage. In the present study, after 3 days of similar exposure, some neuromasts of both types (at most 8.6% for superficial and 30.8% for canal neuromasts) were damaged by cadmium ions. Three days of intermittent $0.5 \mu\text{g l}^{-1}$ cadmium exposure might thus be the threshold period needed to affect sea bass lateral line system tissues. This hypothesis is reinforced by our other results: the maximum cadmium bioaccumulation in fish scales and major behavioural consequences were also measured after three days of exposure.

Before chronic cadmium exposure, gills and scales from control fish already presented a relatively high amount of cadmium, as previously found in the black goby *Gobius niger* (Migliarini et al., 2005) and in the juvenile olive flounder

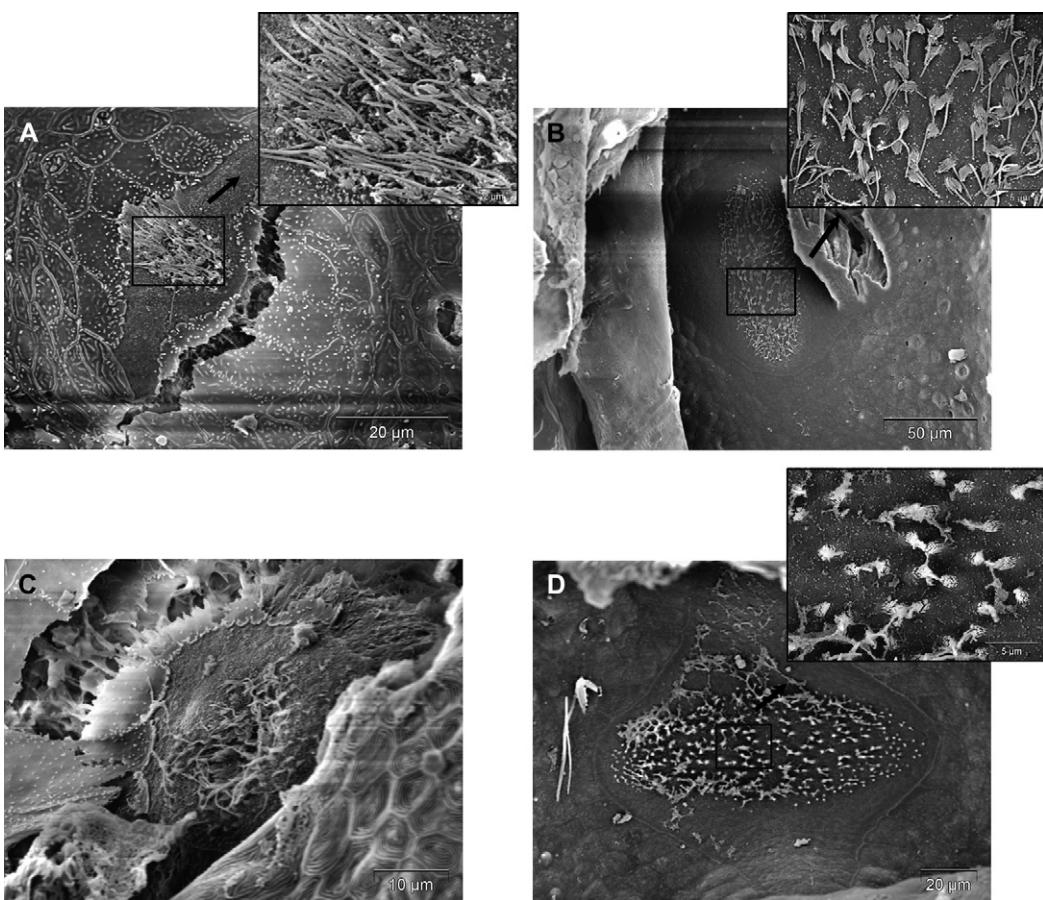


Fig. 4. Scanning electron micrographs showing the effect of chronic low-concentration cadmium exposure ($0.5 \mu\text{g l}^{-1}$ for 4 h per day) on tissue status of both types of neuromasts of sea bass trunk lateral line at the end of the experiment, 21 days after the beginning of exposure. Superficial (A) and canal (B) neuromasts were generally intact. Inserts in A and B illustrate details of sensory maculae with normal hair bundles. Nevertheless, some superficial (C) and canal (D) neuromasts did appear slightly altered.

Paralichthys olivaceus (Kim et al., 2004). This may be due to previous cadmium exposures of the fish during the course of their life. Sea bass used in this study were obtained from a commercial source located in Ile de Ré, an island located on the west coast of France and known for its relatively high concentration of dissolved cadmium ions in seawater (Boutier et al., 2000). Cadmium concentrations measured in gills and scales of control fish corresponded thus to the background. Nevertheless, bioaccumulation data obtained showed that after exposure, cadmium accumulated much more in fish scales than in gills. It is hence relevant to note that fish scales might be a pollutant marker more sensitive than gills which have been commonly used until now. This suggestion is supported by the observation of the adverse effect of cadmium on the morphology (Yoshitomi et al., 1998) and the structural aspect (Rishi and Jain, 1998) of freshwater fish scales. Rishi and Jain (1998) argued that fish scales could thus be used as a biomarker of pollution, particularly as these can be used without sacrificing the animal. In vitro, it has been also demonstrated that cadmium influenced osteoclastic activities after acute exposure and inhibited osteoblastic activities under long-term exposure (Suzuki et al., 2004). Moreover, our results showed that after a maximum at three days, cadmium accumulation tended to

decrease. The fact that fish accumulate less cadmium in their scales after 8 days of exposure could be explained by the action mechanism of cadmium ions on cells. Cadmium and calcium ions are known to be mutually antagonistic in their fixation on sites located at the gills (Verbost et al., 1987, 1988). Given that the lateral line system functions through calcium ion flux, we suggest that cadmium may block the Ca^{2+} -ATPase pump of the baso-lateral membrane of neuromast hair cells. The result would be a blocking of calcium transport in cells associated with their clearly observed degeneration. One could hypothesise that, as the cells degenerate, fixation sites become less available to cadmium ions and the consequence could be less measured cadmium accumulation in the tissue. This hypothesis corroborates Migliarini et al.'s (2005) speculation that cadmium ions fix to binding sites on the gills until they are totally occupied. Yet, fish gills and scales are thought to be usually covered by mucus that protects for the skin and the sensitive gill epithelium against xenobiotics such as metal ions (Pawert et al., 1998; Bruslé and Quignard, 2004). However, at least in the present work, mucus appeared not to protect much against metal ions such as cadmium, since at least a few of both types of neuromasts were damaged by cadmium. Cadmium ions must have passed across

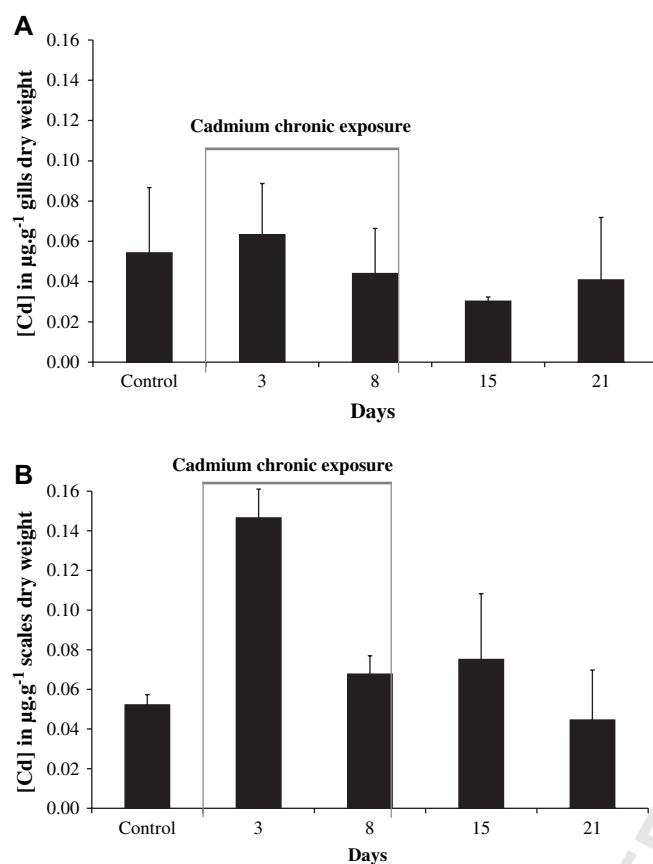


Fig. 5. Average cadmium concentrations (in µg g⁻¹ of dry weight) in fish gills (A) and scales (B) of sea bass, *Dicentrarchus labrax*, exposed chronically to cadmium ions (4 h per day at 0.5 µg l⁻¹) for 8 days. Vertical bars represent the standard deviation.

the mucus layer and damaged sensory hair cell bundles of the fish lateral line system, in spite of the low-concentration of cadmium applied. This result, combined with those obtained with fish lateral line systems exposed to acute high concentration of cadmium (Faucher et al., 2006), refutes the proposal of Døving (1991) that the lateral line organs are shielded from direct pollutant exposure by a set of supporting cells and by their gelatinous cupulae. Our results show that, in contrast to Hudspeth's (1983) and Døving's (1991) hypotheses, mechanoreceptors of the lateral line system are not only accessible by pollutants *via* the internal path (blood) but also by direct external exposure.

In our experiments, the behavioural consequence of alteration of the lateral line system by cadmium ions was a decrease in fish escape responses by about 10%. This result is supported by some studies realized on freshwater fish showing that cadmium ions can disrupt reproductive behaviour (Jones and Reynolds, 1997), agonistic behaviour (Sloman et al., 2003), spawning site selection and natal homing (Baker and Montgomery, 2001), predator avoidance and prey capture (Scott et al., 2003) and also electroreception (Neuman et al., 1991). Our previous study had shown that when fish were exposed to acute high-concentration cadmium, the sea bass lateral line system regenerated itself after about twenty-one days (Faucher et al., 2006). In this study, about fifteen days were

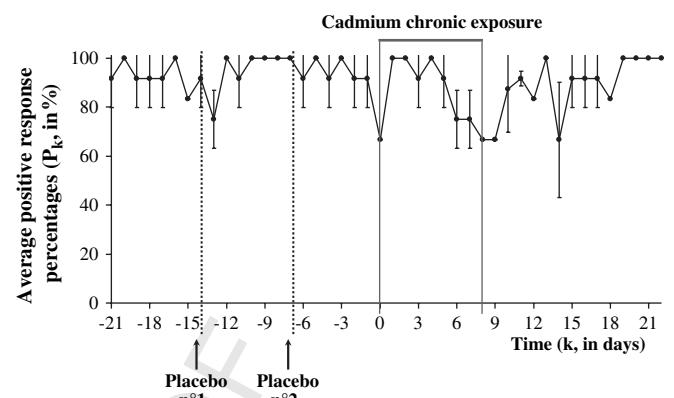


Fig. 6. Average percentages of positive C-start escape responses caused by lateral line system stimulations over consecutive days. Day zero on the x-axis corresponds to the day from when fish were exposed to 0.5 µg l⁻¹ cadmium. Before cadmium exposure, the majority of sea bass positively reacted to water jet. In contrast, as soon as their lateral line system was exposed to low-concentration cadmium, the average positive response percentage fell significantly. This average percentage positive response declined during the period of cadmium exposure (8 days). Then, a recovery to baseline escape behaviour percentages in response to jet stimulation was observed from the 15th day after the beginning of cadmium exposure. Vertical bars represent the standard deviation.

necessary after the beginning of the chronic low-concentration cadmium exposure for fish to show a progressive restoration of their escape behaviour. In parallel, at this time, the tissues of both types of neuromasts presented progressively less damage compared to observations realized during high-cadmium exposure. We can thus conclude that, after such a chronic

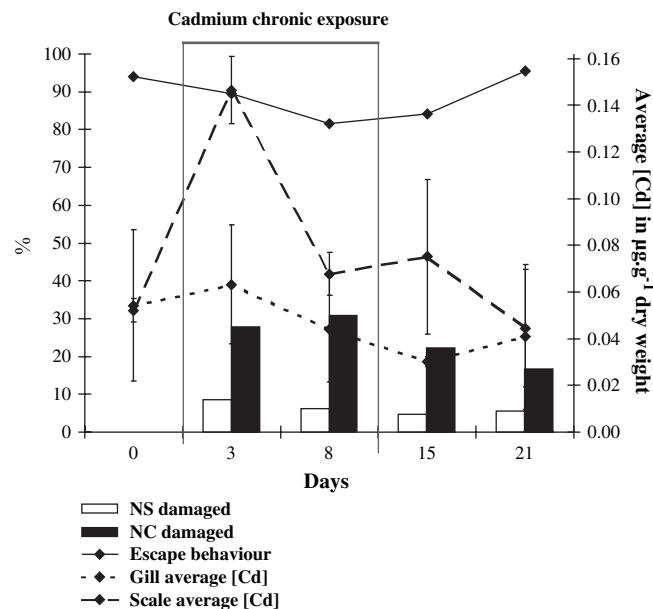


Fig. 7. Summary of all results obtained in this study: cumulated percentage positive escape responses of fish (continuous black line) obtained before, at 3, 8, 15 and 21 days after the beginning of cadmium chronic exposure (grey box), cumulated percentages of superficial (NS, white histogram) and canal neuromasts (NC, black histogram) damaged and average concentrations of cadmium in fish gills (wide black dashed line) and scales (narrow black dashed line). Vertical bars represent the standard deviation of average cadmium concentration.

low-concentration cadmium exposure, the lateral line system needs about 15 days to regenerate itself at a sufficient rate to allow full detection of hydrodynamic stimuli. After chronic exposure (for 8 days) to cadmium, regeneration of both types of neuromasts proved to be quicker than Faucher et al. (2006) found after acute exposure (for only 1 day) to 10-fold higher concentration cadmium. This shorter time needed for neuromasts to regenerate is likely because, in the present study, few neuromasts of both types were entirely destroyed whereas after the acute exposure to high-concentration cadmium, all superficial canal neuromasts were totally destroyed.

To test the sea bass lateral line system function, the stimuli applied in this study were relatively strong, probably more intense than signals received by fish when a predator or a prey approaches. In this way, if sea bass exposed to this low concentration of cadmium responded less frequently (reduction by 10%) to strong stimuli, it is likely that reaction to the weaker stimuli characteristic of their natural environment would be reduced relatively even more markedly. Even if the tissues of their neuromasts remained normal in aspect, lower efficiency is likely in the detection of potential predators or prey in natural environments after exposure to cadmium. To confirm or refute this hypothesis, the lateral line system response to variable intensity stimuli would have to be tested by varying the current velocity of stimulations, attempting to approach as far as possible the range of stimuli generated by moving prey. In addition, in French estuaries, fish are permanently exposed to a mixture of metal ions (Ag, Cd, Co, Cu, Hg, Ni, Pb, Zn, etc.), it would hence be relevant to examine whether there exists any synergy or antagonism among these different metal ions on fish lateral line systems inducing consequences for fish behaviour and for the survival of different fish species *in situ*.

5. Conclusions

This study has produced new data and understanding about the vulnerability of the sea bass lateral line system to cadmium, and it also illustrates a new concept in ecotoxicology. Although after chronic low-concentration cadmium exposure, accumulation and sensory tissue damage were both relatively slight, we have clearly demonstrated that such exposure leads to behavioural consequences. More behavioural studies in ecotoxicology are now needed. Behaviour is an important organism trait response that may represent a pollution marker more sensitive and more relevant than observations of changes in physiology or microanatomy alone (Døving, 1991; Scott and Sloman, 2004; Weis, 2004). Behaviour is furthermore integrated with other levels of biological organization (Scott and Sloman, 2004), so it needs to be considered as a predictor and a result of other internal and external biological processes such as ecological and physiological indicators of toxicity.

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References

Atchinson, G.J., Henry, M.G., Sandheinrich, M.B., 1987. Effects of metals on fish behavior: a review. *Environ. Biol. Fish.* 18 (1), 11–25.

Baker, C.F., Montgomery, J.C., 1999a. The sensory basis of rheotaxis in the blind Mexican cave fish, *Astyanax fasciatus*. *J. Comp. Physiol. A.* 184, 519–527.

Baker, C.F., Montgomery, J.C., 1999b. Lateral line mediated rheotaxis in the antarctic fish *Pagothenia borchgrevinkii*. *Polar Biol.* 21, 305–309.

Baker, C.F., Montgomery, J.C., 2001. Sensory deficit induced by cadmium in banded kokopu, *Galaxias fasciatus*, juveniles. *Environ. Biol. Fish.* 62, 455–464.

Blaxter, J.H.S., Batty, R.S., 1985. Herring behaviour in the dark: responses to stationary and continuously vibrating obstacles. *J. Mar. Biol. Ass. U.K.* 65, 1031–1049.

Boutier, B., Chiffolleau, J.-F., Gonzalez, J.-L., Lazure, P., Auger, D., Truquet, I., 2000. Influence of the Gironde estuary outputs on cadmium concentrations in the coastal waters: consequences on the Marennes-Oléron bay (France). *Oceanol. Acta* 23 (7), 745–757.

Bruslé, J., Quignard, J.P., 2004. Les poissons et leur environnement – éco-physiologie et comportements adaptatifs. TEC & DOC, Paris Lavoisier, 1522 pp.

Bryan, G.W., 1979. Bioaccumulation of marine pollutants. *Phil. Trans. R. Soc. Lond. B.* 286, 483–505.

Chiffolleau, J.-F., Gonzalez, J.-L., Miramand, P., Thouvenin, B., Guyot, T., 1999. Le cadmium: comportement d'un contaminant métallique en estuaire. Programme scientifique Seine-Aval. 10, 39 pp.

Cole, H.A., 1979. Pollution of the sea and its effects. *Proc. R. Soc. Lond. B.* 205, 17–30.

Coombs, S., Janssen, J., Webb, J.F., 1989. Diversity of lateral systems: evolutionary and functional considerations. In: Atema, J., Fay, R.R., Popper, A.N., Tavolga, W.N. (Eds.), *Sensory Biology of Aquatic Animals*. Springer-Verlag, New York, pp. 553–593.

Coombs, S., Braun, C.B., Donovan, B., 2001. The orienting response of lake Michigan mottled sculpin id mediated by canal neuromasts. *J. Exp. Biol.* 204, 337–348.

Døving, K.B., 1991. Assessment of animal behaviour as a method to indicate environmental toxicity. *Comp. Biochem. Physiol.* 100C (1–2), 247–252.

Eisler, R., Hennekey, R.J., 1977. Acute toxicities of Cd²⁺, Cr⁶⁺, Hg²⁺, Ni²⁺ and Zn²⁺ to estuarine macrofauna. *Arch. Environ. Contam. Toxicol.* 6, 315–323.

Elbaz-Poulichet, F., Martin, J.M., Huang, W.W., Zhu, J.X., 1987. Dissolved Cd behaviour in some selected French and Chinese estuaries. Consequences on Cd supply to the ocean. *Mar. Chem.* 22 (2–4), 125–136.

Faucher, K., Aubert, A., Lagardère, J.P., 2003. Spatial distribution and morphological characteristics of the trunk lateral line neuromasts of the sea bass (*Dicentrarchus labrax*, L.; Teleostei, Serranidae). *Brain Behav. Evol.* 62, 223–232.

Faucher, K., Fichet, D., Miramand, P., Lagardère, J.P., 2006. Impact of cadmium exposures on the trunk lateral line neuromasts and consequences on the “C-start” response behaviour of the sea bass (*Dicentrarchus labrax* L.; Teleostei, Moronidae). *Aquat. Toxicol.* 76 (3–4), 278–294.

Friberg, L., Piscator, M., Nordberg, G.F., Kjellstrom, T., 1974. Cadmium in the Environment, second ed. CRC Press, New-York.

Hoekstra, D., Janssen, J., 1986. Lateral line receptivity in the mottled sculpin (*Cottus bairdi*). *Copeia* 1, 91–96.

Hollis, L., McGeer, J.C., McDonald, D.G., Wood, C.M., 1999. Cadmium accumulation, gill Cd binding, acclimation, and physiological effects during long term sublethal Cd exposure in rainbow trout. *Aquat. Toxicol.* 46, 101–119.

Hollis, L., McGeer, J.C., McDonald, D.G., Wood, C.M., 2000. Effects of long term sublethal Cd exposure in rainbow trout during soft water exposure: implication for biotic ligand modelling. *Aquat. Toxicol.* 51, 93–105.

1027 Hudspeth, A.J., 1983. Mechanoelectrical transduction by hair cells in the
1028 acousticolateralis sensory system. *Annu. Rev. Neurosci.* 6, 187–215.

1029 Hudspeth, A.J., Corey, D.P., 1977. Sensitivity, polarity, and conductance
1030 change in the response of vertebrate hair cells to controlled mechanical
1031 stimuli. *Proc. Natl. Acad. Sci. U.S.A.* 74 (6), 2407–2411.

1032 Janssen, J.V., Sideleva, V., Biga, H., 1999. Use of the lateral line for feeding in
1033 two Lake Baikal sculpins. *J. Fish Biol.* 54, 404–416.

1034 Jensen, A., Bro-Rasmussen, F., 1992. Environmental cadmium in Europe. *Rev.*
1035 *Environ. Contam. Toxicol.* 125, 101–181.

1036 Jones, J.C., Reynolds, J.D., 1997. Effects of pollution on reproductive behav-
1037 iour of fishes. *Rev. Fish Biol. Fish* 7, 463–491.

1038 Jørgensen, F., 1984. Influence of Ca^{2+} on the voltage-dependent mechanosen-
1039 sitivity of the hair cells in the lateral line organs of *Xenopus laevis*. *Acta*
1040 *Physiol. Scand.* 120, 481–488.

1041 Jouanneau, J.M., Boutier, B., Chiffolleau, J.F., Latouche, C., Phillips, I., 1990.
1042 Cadmium in the Gironde fluvioestuarine system: behaviour and flow. *Sci.*
1043 *Total Environ.* 97/98, 465–479.

1044 Kim, S.G., Jee, J.H., Kang, J.C., 2004. Cadmium accumulation and elimina-
1045 tion in tissues of juvenile olive flounder, *Paralichthys olivaceus* after
1046 sub-chronic cadmium exposure. *Environ. Pollut.* 127, 117–123.

1047 Klinkhammer, G.P., Bender, M.L., 1981. Trace metal distributions in the
1048 Hudson River estuary. *Est. Coast. Shelf Sci.* 12, 629–643.

1049 Larsson, Å., Haux, C., Sjöbeck, M.L., 1985. Fish physiology and metal pollution:
1050 results and experiences from laboratory and field studies. *Ecotoxicol.*
1051 *Environ. Saf.* 9, 250–281.

1052 Migliarini, B., Campisi, A.M., Maradonna, F., Truzzi, C., Annibaldi, A.,
1053 Scarpioni, G., Carnevali, O., 2005. Effects of cadmium exposure on testis
1054 apoptosis in the marine teleost *Gobius niger*. *Gen. Comp. Endocrinol.*
1055 142, 241–247.

1056 Montgomery, J.C., 1989. Lateral detection of planktonic prey. In: Coombs, S.,
1057 Görner, P., Münz, H. (Eds.), *The Mechanosensory Lateral Line, Neuro-
1058 biology and Evolution*. Springer Verlag, New York, pp. 561–573.

1059 Montgomery, J.C., Baker, C.F., Carton, A.G., 1997. The lateral line can mediate
1060 rheotaxis in fish. *Nature* 389, 960–963.

1061 Neuman, I.S.A., van Rossum, C., Bretschneider, F., Teunis, P.F.M., Peters, R.C.,
1062 1991. Biomonitoring: cadmium deteriorates electro-orientation perfor-
1063 mance in catfish. *Comp. Biochem. Physiol.* 100C (1–2), 259–262.

1064 Northcutt, R.G., 1997. Swimming against the current. *Nature* 389, 915–916.

1065 Partridge, B.L., Pitcher, T.J., 1980. The sensory basis of fish schools: relative
1066 roles of lateral line and vision. *J. Comp. Physiol. A* 135, 315–325.

1067 Pawert, M., Müller, E., Triebskorn, R., 1998. Ultrastructure changes in fish
1068 gills as biomarker to assess small stream pollution. *Tissue Cell* 30 (6),
1069 617–626.

1070 Rishi, K.K., Jain, M., 1998. Effect of toxicity of cadmium on scale morphol-
1071 ogy in *Cyprinus carpio* (Cyprinidae). *Bull. Environ. Contam. Toxicol.* 60,
1072 323–328.

1073 Sand, O., 1975. Effects of different ionic environments on the mechano-
1074 sensitivity of lateral line organs in the mudpuppy. *J. Comp. Physiol.*
1075 102, 27–42.

1076 Scott, G.R., Sloman, K.A., 2004. The effects of environmental pollutants on
1077 complex fish behaviour: integrating behavioural and physiological indi-
1078 cators of toxicity. A review. *Aquat. Toxicol.* 68, 369–392.

1079 Scott, G.R., Sloman, K.A., Rouleau, C., Wood, C.M., 2003. Cadmium disrupts
1080 behavioural and physiological responses to alarm substance in juvenile
1081 rainbow trout (*Oncorhynchus mykiss*). *J. Exp. Biol.* 206, 1779–1790.

1082 Sloman, K.A., Baker, D.W., Ho, C.G., McDonald, D.G., Wood, C.M., 2003.
1083 The effects of trace metal exposure on agonistic encounters in juvenile
1084 rainbow trout, *Oncorhynchus mykiss*. *Aquat. Toxicol.* 63, 187–196.

1085 Suzuki, N., Yamamoto, M., Watanabe, K., Kamagawa, A., Hattori, A., 2004.
1086 Both mercury and cadmium directly influence calcium homeostasis result-
1087 ing from the suppression of scale bone cells: the scale is a good model for
1088 the evaluation of heavy metals in bone metabolism. *J. Bone Miner. Metab.*
1089 22, 439–446.

1090 Verbost, P.M., Flik, G., Lock, R.A.C., Wendelaar Bonga, S.E., 1987. Cadmium
1091 inhibition of Ca^{2+} uptake in rainbow trout gills. *Am. J. Physiol.* 253,
1092 R216–R221.

1093 Verbost, P.M., Flik, G., Lock, R.A.C., Wendelaar Bonga, S.E., 1988. Cad-
1094 mium inhibits plasma membrane calcium transport. *J. Membr. Biol.*
1095 102, 97–104.

1096 Voyer, R.A., Heltsche, J.F., Kraus, R.A., 1979. Hatching success and larval
1097 mortality in an estuarine teleost, *Menidia menidia* (Linnaeus), exposed to
1098 cadmium in constant and fluctuating salinity regimes. *Bull. Environ. Contam.*
1099 Toxicol.

1100 Waldichuk, M., 1979. The assessment of sublethal effects of pollutants in the
1101 sea. Review of the problems. *Phil. Trans. R. Soc. Lond. B* 286, 399–424.

1102 Weis, J.S., 2004. Does pollution affect fisheries? Book critique. *Environ. Biol.*
1103 *Fish* 25, 1–3.

1104 Yoshitomi, T., Koyama, J., Iida, A., Okamoto, N., Ikeda, Y., 1998. Cadmium-
1105 induced scale deformation in carp (*Cyprinus carpio*). *Bull. Environ. Contam.*
1106 Toxicol.