# ORIGINAL RESEARCH

Revised: 1 November 2018



# Biogeography of functional trait diversity in the Taiwanese reef fish fauna

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#### **Funding information**

Ministry of Science and Technology, Taiwan, Grant/Award Number: 104-2611-M-002-020-MY2, 106-2611-M-002-008 and 107-2611-M-002-011; National Taiwan University, Grant/Award Number: 106R4000

#### Abstract

The richness of Taiwanese reef fish species is inversely correlated to latitude as a direct consequence of the abiotic environment and its effects on benthic habitats. However, to date, no studies have investigated the variations in the diversity of traits (FD) linked with the role of these fishes in the ecosystem. FD is usually considered more sensitive than species richness in detecting early changes in response to disturbances, and therefore could serve as an indicator of ecological resilience to environmental changes. Here, we aim to characterize FD in the Taiwanese reef fish fauna and to document its regional variations. Six traits were used to categorize the 1.484 reef fish species occurring in four environmentally contrasted regions around Taiwan. The number of unique trait combinations (FEs), their richness (FRic), their redundancy (FR), their over-redundancy (FOR), and their vulnerability (FV) were compared among these regions. Overall, 416 FEs were identified. Their number decreased from south to north in step with regional species richness but FRic remained similar among regions. FR and FOR were higher to the south. At the local scale, variations in FEs and FRic are in concordance with the worldwide pattern of FD. High-latitude, impoverished fish assemblages, offer a range of trait combinations similar to diversified tropical assemblages. Increasing diversity in the latter mainly contributes to raising FR and supports already over-redundant entities. High vulnerability makes many combinations highly sensitive to species loss, and was higher at intermediate latitudes when using a fine resolution in trait categories. It suggests that the loss of FEs may first be characterized by an increase in their vulnerability, a pattern that could have been overlooked in previous global scale analyses. Overall, this study provides new insights into reef fish trait biogeography with potential ramifications for ecosystem functioning.

#### KEYWORDS

functional diversity, high-latitude, ichthyofauna, Kuroshio, marginality, traits

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# 1 | INTRODUCTION

Rising seawater temperatures from climate change are pushing tropical marine organisms close to their upper thermal limits (Walther et al., 2002) and transforming coral reef ecosystems (Hughes et al., 2018: Stuart-Smith, Brown, Ceccarelli, & Edgar, 2018). At a global scale, this is commonly associated with a flattening of coral reefs (Alvarez-Filip, Dulvy, Gill, Côté, & Watkinson, 2009; Bozec, Alvarez-Filip, & Mumby, 2015; Wilson et al., 2010), which further reduces the availability of niche spaces upon which a large number of living organisms depend (Graham & Nash, 2013). These modifications notably alter the structure of reef fish assemblages as well as their trophic interactions (Darling et al., 2017; Graham et al., 2007; Pratchett, Hoey, Wilson, Messmer, & Graham, 2011; Richardson, Graham, Pratchett, Eurich, & Hoey, 2018; Wilson et al., 2010). The susceptibility of fish species to extirpation is usually nonrandom (Graham et al., 2011) and benefits generalists, which today seem able to proliferate on reefs worldwide (Munday, 2004; Richardson et al., 2018). This occurs to the detriment of a small proportion of species with often prominent roles in maintaining ecological processes (Mouillot, Bellwood, et al., 2013a), although generalists can play influential roles too. The loss of functionally important fishes can eventually decrease the capacity of the ecosystems to recover from, or resist transitions to, alternative states, which may have only a limited capacity to sustain original ecosystem services (Chong-Seng, Nash, Bellwood, & Graham, 2014; Fung, Seymour, & Johnson, 2011; Hoey & Bellwood, 2011; Hughes et al., 2017; Pratchett, Hoey, & Wilson, 2014). Consequently, a better understanding of the diversity in functional groups, as well as the functional relationships between coral habitats and associated fish assemblages are of critical importance in managing reefs and prioritizing conservation efforts (Richardson et al., 2018).

Reef fishes represent a dominant structuring force on coral reefs (Raymundo, Halford, Maypa, & Kerr, 2009), controlling the distribution and abundance of many taxa such as mollusks and algae, as well as hard corals (Ceccarelli, Jones, & McCook, 2011; Jompa & McCook, 2002; McClanahan, 1994). Indeed, several critical functional groups of fishes have now been identified that enhance the prospects of recovery toward initial conditions after disturbances (e.g., Bellwood, Hughes, & Hoey, 2006a; Cheal et al., 2010; Graham, Jennings, MacNeil, Mouillot, & Wilson, 2015). For example, herbivorous fishes can mediate the growth, and sediment trapping abilities of algal turf (Goatley, Bonaldo, Fox, & Bellwood, 2016; Tebbett, Goatley, & Bellwood, 2017), which is likely to benefit coral recruitment (Birrell, McCook, & Willis, 2005). In some other cases, predatory fishes may structure reef communities through top-down control, potentially preventing trophic cascades (Dulvy, Freckleton, & Polunin, 2004; but see Casey et al., 2017 and discussion in Strong, 1992).

Recently, the unique combination of traits (characteristics) describing the role of fish species in the ecosystem has tentatively been proposed as synthetic representation of the functions they performed (Mouillot et al., 2014). This representation is commonly considered a pragmatic description of their potential niche and corresponds to any feature at the individual level that can affect and determine how species use resources and how they interact with each other (Cadotte, Carscadden, & Mirotchnick, 2011). To date, traits responding to environmental variations and focusing on individual performance ("functioning traits"; see Denis, Ribas-deulofeu, Sturaro, Kuo, & Chen, 2017) have been largely overlooked in this definition. Instead, six categorical traits related to food acquisition and locomotion, considered pertinent to representing key facets of fish ecology (size, diet, mobility, gregariousness, period of activity, and vertical position in the water column) have been widely employed (Mouillot, Bellwood, et al., 2013a; Mouillot et al., 2014). The depiction of the diversity of those traits through discrete values has gained in popularity, and sometimes adjusted, for providing a better insight into ecosystem functions beyond that given by species diversity measures (Richardson et al., 2018). Although, several important caveats undermine this relationship (Brandl, Emslie, & Ceccarelli, 2016) and this approach is perfectible (Villéger, Brosse, Mouchet, Mouillot, & Vanni, 2017). However, functional trait-based approach and, more precisely, the characteristics of the unique trait combinations, represents today a benchmark for assessing the diversity of functional traits in the communities (Mouillot, Graham, et al., 2013b).

Diversity of traits (FD), the range of things that organisms do in communities and ecosystems (Petchey & Gaston, 2006), is directly connected to ecosystem processes and stability (Chapin et al., 2000; McCann, 2000; Purvis & Hector, 2000). FD is measured in several ways (Cadotte et al., 2011), and its monotonic response to disturbance usually makes it a better indicator than the traditional taxonomic measurements of early changes affecting reef communities (D'Agata et al., 2014; Mouillot, Graham, et al., 2013b). As an example, focusing on trophic trait, Bellwood, Hughes, Folke, and Nyström (2004) examined fish species richness in 14 pre-determined categories comparing Caribbean and Australian reefs. The Caribbean assemblage was characterized by a lower redundancy (*i.e.*, average number of species among categories) than the Australian, which depicted the higher vulnerability of Caribbean reef ecosystems for the trait of interest.

By including morphological information into a trait morphospace, Bellwood, Wainwright, Fulton, and Hoey (2006b) further confirmed high versatility in trophic categories characterizing coral reef fish diversity. In most feeding guilds, this was supported by a considerable overlap in the occupation of the morphospace by different morphologies. Further extension to behavioral traits reinforced that a small number of key combination of traits could provide the basic ecological structure of reef fish assemblages (Guillemot, Kulbicki, Chabanet, & Vigliola, 2011). This is illustrated on a global scale by a high redundancy that is disproportionately packed into few trait combinations (over-redundancy). In a nonrandom assembly, it leaves a large number of entities supported by only one species, thus increasing the overall vulnerability of trait combinations in the reef fish fauna and the potential sensitivity of FD to the loss of a few species (Mouillot et al., 2014). Surprisingly, FD is revealed to be relatively stable across biogeographical provinces despite their high variability in species richness (Mouillot et al., 2014). Hence, impoverished fish faunas such as those from the Tropical Eastern Pacific and

Biogeographical provinces primarily correspond to geographical areas with distinctive fish faunas (Briggs & Bowen, 2012; Kulbicki et al., 2013). Actually, their definitions confound environmental parameter operating at a more local scale that control species distribution. Abjotic parameters such as temperature. light, and hydrodynamic regime are important variables defining ecoregions (Spalding et al., 2007) and strongly determine the regional composition of coral communities (Kleypas, McManus, & Meñez, 1999). Fish diversity (across reef habitats) varies on a latitudinal gradient together with changes happening in benthic assemblages (Bellwood & Hughes, 2001; Hillebrand, 2004; Mora, Chittaro, Sale, Kritzer, & Ludsin, 2003; Tittensor et al., 2010). Those changes can modify the trophic structure of an ecosystem (Ferreira, Floeter, Gasparini, Ferreira, & Joyeux, 2004) and eventually influence regional FD. However, there is always the possibility that a local variation has been largely overlooked in previous studies.

Taiwan is located to the north of the East Indies Triangle, the region hosting the highest marine diversity in the world (Briggs, 2005). It encompasses tropical and subtropical latitudes and spans the transition of two marine realms where three provinces and ecoregions converge (Spalding et al., 2007). The Taiwanese reef fish fauna is highly diverse (Allen, 2008), and species richness decreases steeply from southeast to northwest in correlation to the patterns in sea surface temperatures that shape reef habitats (Dai, Soong, Chen, Fan, & Hsieh, 2005). Therefore, Taiwan constitutes an ideal location for investigating the potential effects of environmental settings on the FD of reef fish assemblages. The objectives of this study were to (a) compare FD among the main regions of coral development around Taiwan with respect to prevailing environmental conditions, and (b) examine the local spatial patterns of FD of Taiwanese reef fish fauna in comparison to global patterns. To achieve this, the present study compares the richness, redundancy, over-redundancy, and vulnerability of the unique functional trait combinations identified in four coral regions around Taiwan characterized by contrasting environmental conditions.

# 2 | MATERIALS AND METHODS

#### 2.1 | Reef fish fauna

Our study focuses on four well-supported reef fish assemblages (so-called regions, Figure 1a): North, East (including Ludao), West (Penghu Archipelago), and South of Taiwan (including Xiaoliuqiu). This distinction is justified by environmental conditions producing



**FIGURE 1** (a) The four studied regions around Taiwan. (b) Their respective species richness and number of unique species, unique trait combinations (*FEs*), range of unique trait combinations (*FRic*). The y-axis is the percentage of richness in all bar plots

3

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contrasting benthic assemblages (Chen, 1999; Chen & Shashank, 2009; Ribas-Deulofeu et al., 2016) and hosting distinct reef fish assemblages (Shao, Chen, & Wang, 1999). The warm waters of the Kuroshio Current flow from the southern point of Taiwan along its east coast toward the Ryukyu Archipelago, pushing tropical organisms (e.g., scleractinian corals: Chen, 1999) northward, Accordingly, fringing reefs only occur on the East and South coasts of Taiwan, where average monthly seawater temperatures remain above 20°C. In contrast, the frequent occurrence of waters lower than 18°C in winter prevents the accretion of reefs to the West and the North of Taiwan (Chen, 1999; Klevpas et al., 1999; Wang & Chern, 1988). There, only non-reefal and less diversified coral assemblages develop directly on the basalt substrate that shapes the coastline. From a benthic survey throughout 2011–2012 at selected sites, coral cover at 5 m in depth were estimated to 24.1% at Kenting (South), 39.6% at Green Island (East), 21% at Penghu (West), and 18.8% to the North of Taiwan (Ribas-Deulofeu et al., 2016).

Reef fish composition in these four regions were obtained by corroborating information from Taiwan FishBase (Shao, 2018), FishBase (Froese & Pauly, 2018), and additional references documenting local reef fish diversity (Chen, 2004a, 2004b; Chen, Jan, Kuo, Huang, & Chen, 2009; Chen, Shao, Jan, Kuo, & Chen, 2010; Shao et al., 2008). Here, a fish was considered as a coral reef species if it is partly or strictly associated, during its lifetime, to shallow coastal coral habitat, including the non-reefal environments to the West and the North of Taiwan. For each species, its synonymy and status validity were checked in the World Register of Marine Species (WoRMS Editorial Board, 2018), and only valid species were used for further analysis. This step proved to be of critical importance in the preparation of data extracted from databases.

#### 2.2 | Trait selection

Our final list of reef fish species among the four regions in Taiwan encompassed 1,484 species. A species was considered as unique to a region when it was absent from the others. Each species was classified into six categorical traits reflecting its possible functions (the role of the species) in the ecosystem: size, diet, mobility, gregariousness, period of activity, and vertical position in the water column (following Mouillot et al., 2014). These traits and their levels are detailed in Supporting Information Appendix S1. Their relevance in describing reef fish functions has been previously evaluated tentatively by Mouillot et al. (2014). The strength of the application of these six traits depends on the extent to which these traits really are indicative of functional attributes (Mouillot, Graham, et al., 2013b). However, their extensive use in the broader field of reef fish functional ecology does not imply their links to ecosystem functioning are clear (see Beauchard, Veríssimo, Queirós, & Herman, 2017; Villéger et al., 2017 for discussion). Species trait information was extracted from databases and/or relevant references (Froese & Pauly, 2018; Shao, 2018). Because trait values can differ among sources, priority was first given to FishBase, followed by Taiwan FishBase, then other references. Missing species trait values were infilled with specific literature on the given species or with the expertise of the

authors considering the phylogenetic position of the species. The final dataset provides a species list, regional distributions, and selected traits.

#### 2.3 | Trait entities, space, and indices

#### 2.3.1 | Trait entities

Functional trait entity (*FE*) was defined as a unique combination of categorical trait values. Theoretically, the six traits and their categories yield a total number of 5,670 unique combinations. Yet many combinations do not occur and only 646 FEs (11.4%) were defined based on an analysis of global reef fish fauna (Mouillot et al., 2014).

#### 2.3.2 | Trait space

A dissimilarity (trait) matrix was produced by a pairwise comparison of the *FEs* using Gower distance (S15). S15 has the advantage of being suitable for mixed (continuous and categorical) variables, and is therefore applicable here. A Principal Coordinates Analysis (PCoA) was then computed on the basis of this trait matrix applying a Caillez correction to correct any potential negative eigenvalues generated (Cailliez, 1983). Euclidean distances among *FEs* on the first four axes of this PCoA were firmly correlated with the initial Gower matrix (Mantel test, r = 0.77, p < 0.001), and the addition of an extra axis only marginally increased this resolution. Therefore, information on the first four axes was considered the most pragmatic representation of variation among *FEs*. The coordinates of *FEs* on these axes were used to represent synthetic trait space (to visualize relationship among *FEs*) and later calculate *FD* indices.

#### 2.3.3 | Trait diversity indices

In addition to the number of FEs, the range of unique trait combinations (FRic, Mason, Mouillot, Lee, & Wilson, 2005) was computed for each region. FRic was calculated in the four dimensions of our trait space. It expresses for the fish assemblage in each region the proportion of the volume of the convex hull (%) that the species occupied. In addition, we described the four reef fish assemblages around Taiwan based on the characteristics of their FEs. The redundancy of trait combinations (FR) corresponds to the average number of species sharing similar sets of traits (Equation 1); the vulnerability of trait combinations (FV, %) represents the proportion of trait combinations supported by only one species (Equation 2); and the over-redundancy of trait combinations (FOR, %) indicates the percentage of species in FEs having more species than expected from FR (Equation 3). Following Mouillot et al. (2014), those indices are expressed as:

$$FR = \frac{\sum_{i=1}^{FE} n_i}{FE} = \frac{S}{FE}$$
(1)

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$$FV = \frac{FE - \sum_{i=1}^{FE} \min(n_i - 1, 1)}{FE}$$
(2)

$$FOR = \frac{\sum_{i=1}^{FE} \left[ \max \left( n_i, FR \right) - FR \right]}{S}$$
(3)

where *FE* is the total number of trait entities, *S* is the total number of fish species, and is the number of species in a trait entity.

FV and FOR are influenced by the number of species, the number of FEs, and the distribution of species among FEs. Therefore, to test whether the observed values were significantly different from theoretical values, the observed values of FV and FOR were compared with the null hypothesis that species were distributed randomly in each FE. Briefly, for each of the four study assemblages, we randomly assigned species in the previously determined number of FEs (ensuring that at least one species occupied each FE). In this way, we simulated 9,999 assemblages and computed FV and FOR in each of them. We compared observed FV and FOR values to theoretical values with a bilateral test ( $\alpha$  = 0.05) and further recognized how significantly different they were from a random distribution by examining the standardized effect of size. For instance, if the observed values of FV and FOR were higher than simulated, it indicated that the distribution of species was concentrated within particular FEs, leaving a number of FEs supported by only a single species.

The sensitivity of our results to the extent of trait categories was further assessed using a crude categorization following Mouillot et al. (2014). It consisted of reducing trait categories as detailed in Supporting Information Appendix S1, rerunning all the analyses on this basis, and comparing the results with the ones from the finer resolution in traits categories. Theoretically, the six traits and their crude categorization yield a total number of 216 unique combinations.

All data analyses were performed in R software (v3.4.3, R Core Team, 2017) using worms (Holstein, 2018), ape (Paradis, Claude, & Strimmer, 2004), cluster (Maechler, Rousseeuw, Struyf, Hubert, & Hornik, 2017), and FD (Laliberté & Legendre, 2010; Laliberté, Legendre, & Shipley, 2014) packages as well as the functions "quality\_funct\_space" and "multidimFD" (Mouillot, Graham, et al., 2013b) available at http://villeger.sebastien.free.fr/Rscripts.html (last accessed 2018/07/19).

#### 3 | RESULTS

#### 3.1 | Fish richness and unique trait combinations

Fish species richness ranges between 618 species in the North to 1,278 species in the South, which respectively represent 41.6% and 86.1% of Taiwan's total richness (Figure 1b). The South has the highest number of unique species (306 species, 23.9% of the regional richness), followed by the North and East with 78 (12.6%) and 65 (7.4%)

unique species, respectively. The western region hosts only 14 unique species (2.2% of the regional richness). The classification of those species from all four regions into six traits yields a total of 416 FEs representing 7.3% of the total theoretical number of unique combinations (*i.e.*, 5,670 combinations). The number of FEs follows the species richness pattern and is higher in the South (393 FEs) compared to the East (338 FEs), West (274 FEs), and North (258 FEs; Figure 1b).

#### 3.2 | Trait space and richness

Most of the variation in our Gower matrix comparing the FEs was caught in the first four axes of our trait space (Mantel test,  $R^2 = 0.58$ , p < 0.001). It also reduced the mean squared deviation considerably between the initial distance and, the standardized final distance in the trait space (quality\_funct\_space function, mSD = 0.01). Therefore, the dispersions among FEs and the computation of FRic were considered in four dimensions. The positions of our 416 FEs were determined by unique combinations of trait values (Supporting Information Appendix S2). FRic represents the hypervolume delimited by the most extreme entities in these four dimensions and was computed for each region as the proportion that local fauna occupied (Supporting Information Appendix S3). Regional FRic are 99.6% in the South, 97.7% in the East, 92.1% in the West, and 91.7% in the North (Figure 1b). Therefore, despite variable species richness and FEs, the range of unique trait combinations supported by each assemblage, remain relatively stable among regions with only a few differences noticeable in the 3rd and 4th dimensions of our trait space (Supporting Information Appendix S3). Similar parts of the trait space were lost in the West and the North explaining the slightly lower FRic values in these regions.

### 3.3 | Vulnerability, redundancy, and overredundancy of trait combinations

The distribution of fish species by FEs displays a characteristic positive skewness in all four regions (Figure 2). Testing the species richness-related indices (i.e., FOR and FV) against null models (Supporting Information Appendix S4) reveals that the distributions observed are nonrandom (p < 0.001) and that species are disproportionally packed into a low number of FEs. The average (±SE) number of species per FE (FR) is the highest in the South  $(3.3 \pm 0.2 \text{ species})$  and the lowest in the West (2.3 ± 0.2 species; Figure 2). FOR indicates that about 42.4% of the species contribute to the over-representation of some FEs at the national scale. FOR ranges between 41.0% in the South and 35.2% in the North, with intermediate levels of 38.5% and 37.4% for the East and West, respectively (Figure 2). Notably, many FEs are left without insurance, as almost half (49.0%) are supported by only a single species at the national scale. At a regional level, FV values are lower in the South (51.2%) and intermediate in the North (55.4%). FV reaches maximum values at intermediate latitudes with 61.7% and 58.3%, in the West and East of Taiwan, respectively (Figure 2). Unique species contribute 8.7% in the South, 3.6% in the East, 0.0% in the West, and 7.0% in the North, to FV (Figure 2).



**FIGURE 2** Distribution of species among unique trait combinations (*FEs*) with their redundancy (*FR*), vulnerability (*FV*), and overredundancy (*FOR*) among regions around Taiwan. *FR* is the mean number of species among all *FEs*. *FV* is the percentage of *FEs* possessing only one species. *FOR* is the percentage of species contributing to the over-representation of some sets of traits (*i.e.*, gray oblique lines). The black stair line corresponds to the distribution of species in *FEs* in the overall Taiwanese reef fish fauna. The black oblique dashed line area represents the contribution of unique species to *FV* 

#### 3.4 | Sensitivity analysis

<sup>6</sup> WILEY Ecology and Evolution

The reduction of trait categories decreased the number of *FEs* by 4.4-fold (22.6%) in the overall Taiwanese reef fish fauna (94 vs. 416 *FEs*). As a consequence, it increased the redundancy in trait combinations (15.8  $\pm$  2.7), but species remained disproportionally packed into a low number of *FEs* (*FOR*: 49.4%) leaving many of them (18.1%) with no insurance. Crude categorization slightly increases *FOR* (+7.0%) and decreases *FV* (-30.9%). Regionally, it reduced the number of *FEs* by 4.2-fold in the South, 3.7-fold in the East, and 3.1-fold in both the West and the North of Taiwan (Table 1). Patterns in regional variation of *FE* and *FRic*, *FR* and *FOR* are consistent with the one using a finer resolution in trait categories. However, *FV* was found to increase with latitudes whereas it was maximum at intermediate latitudes using a finer resolution in trait categories. None of the unique species at a regional level contributes to the *FV* defined using the coarser categorization (Supporting Information Appendix S4).

# 4 | DISCUSSION

Around Taiwan, the reef fish fauna encompasses 1,484 species, which is exceptionally high considering the small overall area of reefs (~940 km<sup>2</sup>) around the island (Allen, 2008). It yields 416 FEs, which correspond to 7.3% of the theoretical maximum possible. In comparison, using the same traits and categories, the world reef fish fauna (6,316 species) yields 646 FEs, and 11.4% of the theoretical combinations (Mouillot et al., 2014). Taiwan is located at the northern limit of the Central Indo-Pacific province, in which the fish fauna yields 468 FEs across 3,600 species (Mouillot et al., 2014). Although there is a two-fold difference in species richness between the overall province and Taiwan, there is an 89% similarity in the total number of FEs observed. This similarity in FEs suggests that the poorer Taiwanese reef fish fauna may be able to maintain the ecological processes necessary for the growth and persistence of reef ecosystems (Bellwood et al., 2004; Johnson et al., 2008).

#### TABLE 1 Sensitivity analyses

	FEs (% richness)	FRic (%)	FR	FOR (%)	FV (%)
North	83 (88.3)	91.0	7.5	46.3	30.1
West	87 (92.6)	93.9	7.4	48.6	28.7
East	91 (96.8)	98.5	9.6	47.0	22.0
South	94 (100)	100	13.6	49.1	20.2

Note. Unique trait combinations (FEs), range of unique trait combination (FR) as well as their redundancy (FR), vulnerability (FV), and over-redundancy (FOR) among regions around Taiwan after reducing the number of trait categories. Overall, this coarse categorization reduced the number of possible FEs by 4.4 fold (94 vs. 416 using a fine resolution in trait categories).

The four studied regions also had similar *FRic* values despite exhibiting large differences in species richness. The four regions fill between 91.7% and 99.6% of the trait space defined at the national scale. The lowest values to the West and the North are mostly explained by the loss of some given *FEs*, for which the possible link with prevailing environmental conditions warrants further investigation. Congruent with patterns observed among global biogeographic provinces (Mouillot et al., 2014), variation in species richness or in the number of *FEs*, seems to have little influence on *FRic* levels estimated in the four fish assemblages assessed. This pattern is especially interesting in light of the scale of our study and the contrasting coral richness among the four regions (Dai et al., 2005).

While coral trait diversity is also remarkably conserved among global biogeographic provinces and along gradients of species richness (McWilliam et al., 2018), FRic that characterizes coral assemblages around Singapore responds strongly to patterns in species distribution (Wong et al., 2018). Based on the functional relationships identified between coral and reef fish assemblages among sites (Darling et al., 2017), we could have expected a similar reduction of FRic in marginal locations where coral richness is impoverished. We hypothesized that regional consideration of species occurrence (gamma trait diversity-regional pool) could confound fine variation in FRic (alpha trait diversity-local pool), which might be relevant in interpreting local, and more specifically, human-mediated loss of FD (D'Agata et al., 2014). Site-specific, local surveys of reef fish abundance or biomass (Stuart-Smith et al., 2013) may provide further insight into the FRic pattern within regions, in particular its responses to local degradation in benthic assemblages, and their associated traits, as observed around Singapore (Wong et al., 2018).

Environmental conditions in the North and the West of Taiwan are marginal for corals (Dai & Horng, 2009) and thus limiting for reef accretion (Kleypas et al., 1999). Although these two regions only host about half of the fish species recorded from the South of Taiwan, they sustain a relatively high proportion (66% and 70% for the North and West, respectively) of the total number of *FEs* observed in the South. It suggests that these communities could have the ability to maintain a wide range of ecological processes present in more tropical locations. Consequently, these two impoverished regions have the lowest *FR* and *FOR* values. Even where these values were highest in the South (averaging  $3.3 \pm 0.2$  species per *FE* and 41.0%, respectively), they remain largely below those documented in the Central Indo-Pacific (7.9 species per *FE* and 57.9%, respectively) and are comparable to values from the impoverished reef fish fauna in the Tropical Eastern Pacific (2.8 species per *FE* and 40.0%, respectively; Mouillot et al., 2014). This supports the notion that the addition of new species does not contribute substantially toward generating new *FEs* (Halpern & Floeter, 2008). Instead, it tends to make existing functional fish assemblages more robust by offering high "insurance" in a small number of *FEs*.

The over-representation of species in a few *FEs* was further robust to a drastic reduction of the possible number of *FEs* as in the global reef fish fauna (Mouillot et al., 2014). In addition, we demonstrated that the over-representation of a low number of *FEs* is better preserved than the *FR* along a decreasing gradient of species richness. Despite a similar range of *FEs* offered, it makes the functionality of the fish assemblages from northern, western, and eastern regions potentially more vulnerable and less resilient to disturbances than the southern region.

The location of Taiwan at the periphery of the Central Indo-Pacific province is reflected in the relatively high vulnerability of its fish fauna. Regional FVs ranged between 51.2% and 61.7%, which is actually in the range reported from the Tropical Eastern Pacific province (54.2%) and much higher than the FV of the Central Indo-Pacific (38.5%; Mouillot et al., 2014). This suggests that, independently of the location around Taiwan, at least half of the FEs are supported by only one species, leaving many FEs without insurance when facing species loss (Mouillot, Bellwood, et al., 2013a). Rare species, both in terms of local abundance and regional occupancy, often support unique and distinct functions (Mouillot, Bellwood, et al., 2013a) that could be critical for the resilience of coral reefs (Bellwood, Wainwright, et al., 2006b). Unfortunately such species are usually the most susceptible to extraction and extirpation (Graham et al., 2011), which may result in the functional homogenization of fish assemblages and a proliferation of generalist species (Munday, 2004; Richardson et al., 2018).

Characteristic of a nonrandom fish assemblage, species loss should result in an increase in FV and/or an increase in FOR (Halpern & Floeter, 2008). Yet, regional patterns observed at large spatial scales support a global mismatch between species richness and FV (Parravicini et al., 2014). Our study further uncovers an interesting pattern in FV that macro-ecological research may have overlooked. Fish assemblages from regions at intermediate latitudes have the highest FV values (in the West and East in this case). From the previous observation that FR will be more affected by species loss than FOR, it suggests that the FV of fish assemblages might first increase as species are lost and then drop further after singular functions are removed, ultimately resulting in a functional homogenization of the assemblage (Clavel, Julliard, & Devictor, 2011). In this scenario, functionally unique species could tend to be rarer in tropical than in temperate reef fish assemblages as reported globally by Stuart-Smith et al. (2013).

Ecology and Evolution

7

A high FV has further been interpreted as a pristine state and a baseline for assemblages untouched by human for priority conservation (Quimbayo, Mendes, Kulbicki, Floeter, & Zapata, 2017). In this scenario, FV could act as an indicator of a functional "tipping-point," a stage prior to more severe functional changes occurring along environmental gradients or from human disturbances. An exciting outcome of these observations will be to test whether FV can be applied to temporal community dynamics and if it could serve as a dimension for assessing ecological shifts. However, FV level is sensitive to the traits considered as well as their categorization (Mouillot et al., 2014). Overall, FV decreases with a crude categorization of traits and raises with latitude, which makes the FEs recorded in marginal species-poor regions more vulnerable to species loss than in richer assemblages. This higher FV observed to the North was not caused by unique species. Instead, coarse trait resolution could decrease the susceptibility of identifying FEs that are vulnerable and overlook subtle functional changes in the assemblages. Therefore, the resolution in trait categorization and its outcomes should be considered and interpreted carefully according to the scale of changes expected.

In this study, the Taiwanese reef fish fauna offered new insights into the biogeography of reef fish trait diversity, which has recently raised as an important facet of the diversity. It constitutes the first step toward a better understanding of the fishes' role in the Taiwanese reefal areas. Taiwanese reef fish fauna retains a high proportion of trait combinations defined at the scale of the Central Indo-Pacific province, but remains highly vulnerable to species loss. Possible overfishing and the consequences of human activities on benthic habitats are likely to have already modified historic diversity (Liu et al., 2009; Ribas-Deulofeu et al., 2016). Therefore, this study proposes a baseline upon which the gain and loss of unique trait combinations could be immediately assessed after updating information on species richness. The integration of abundance and/or biomass to the current framework (see Chen et al., 2015) would improve the present trait assessment and the functional interpretation of the Taiwanese reef fish fauna. This would further enhance the prospects for exploring functional relationships between reef fish assemblages and coral habitats.

#### ACKNOWLEDGMENTS

We thank all members and collaborators of the Functional Reef Ecology Lab (Institute of Oceanography, National Taiwan University) for their useful comments and discussion on this project. This study was funded by the Ministry of Science and Technology of Taiwan (MOST, No. 104-2611-M-002-020-MY2). VD is the recipient of MOST grants (Nos 106-2611-M-002-008 and 107-2611-M-002-011). NS was a recipient of a National Taiwan University postdoctoral fellowship (Taiwan, 106R4000). The authors are particularly grateful for the suggestions of two anonymous reviewers that greatly improved an earlier version of this manuscript.

#### CONFLICT OF INTEREST

The authors have no conflicts of interest to declare.

#### AUTHOR CONTRIBUTIONS

VD designed the experiment. VD and JWC collected the data. VD, JWC, and CWW analyzed the data. VD, JWC, QC, YEH, YVL, CWW, HYW, and NS wrote the paper and revised the drafts.

#### DATA ACCESSIBILITY

All data and script for data analysis are available from the Dryad Digital Repository: https://doi.org/10.5061/dryad.838v1j5.

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#### REFERENCES

- Allen, G. R. (2008). Conservation hotspots of biodiversity and endemism for Indo-Pacific coral reef fishes. Aquatic Conservation: Marine and Freshwater Ecosystems, 18, 541–556. https://doi.org/10.1002/ aqc.880
- Alvarez-Filip, L., Dulvy, N. K., Gill, J. A., Côté, I. M., & Watkinson, A. R. (2009). Flattening of Caribbean coral reefs: Region-wide declines in architectural complexity. *Proceedings of the Royal Society B-Biological Sciences*, 276, 3019–3025. https://doi.org/10.1098/ rspb.2009.0339
- Beauchard, O., Veríssimo, H., Queirós, A. M., & Herman, P. M. J. (2017). The use of multiple biological traits in marine community ecology and its potential in ecological indicator development. *Ecological Indicators*, 76, 81-96. https://doi.org/10.1016/j. ecolind.2017.01.011
- Bellwood, D. R., & Hughes, T. P. (2001). Regional-scale assembly rules and biodiversity of coral reefs. *Science*, 292, 1532–1534. https://doi. org/10.1126/science.1058635
- Bellwood, D. R., Hughes, T. P., Folke, C., & Nyström, M. (2004). Confronting the coral reef crisis. *Nature*, 429, 827–833. https://doi. org/10.1038/nature02691
- Bellwood, D. R., Hughes, T. P., & Hoey, A. S. (2006a). Sleeping functional group drives coral-reef recovery. *Current Biology*, 16, 2434–2439.
- Bellwood, D. R., Wainwright, P. C., Fulton, C. J., & Hoey, A. S. (2006b). Functional versatility supports coral reef biodiversity. *Proceedings of* the Royal Society B-Biological Sciences, 273, 101–107.
- Birrell, C. L., McCook, L. J., & Willis, B. L. (2005). Effects of algal turfs and sediment on coral settlement. *Marine Pollution Bulletin*, 51, 408–414. https://doi.org/10.1016/j.marpolbul.2004.10.022
- Bozec, Y. M., Alvarez-Filip, L., & Mumby, P. J. (2015). The dynamics of architectural complexity on coral reefs under climate change. *Global Change Biology*, 21, 223–235. https://doi.org/10.1111/gcb.12698
- Brandl, S. J., Emslie, M. J., & Ceccarelli, D. M. (2016). Habitat degradation increases functional originality in highly diverse coral reef fish assemblages. *Ecosphere*, 7, e01557. https://doi.org/10.1002/ ecs2.1557
- Briggs, J. C. (2005). Coral reefs: Conserving the evolutionary sources. Biological Conservation, 126, 297–305. https://doi.org/10.1016/j. biocon.2005.06.018

\_Ecology and Evolution

- Briggs, J. C., & Bowen, B. W. (2012). A realignment of marine biogeographic provinces with particular reference to fish distributions. *Journal of Biogeography*, *39*, 12–30. https://doi. org/10.1111/j.1365-2699.2011.02613.x
- Cadotte, M. W., Carscadden, K., & Mirotchnick, N. (2011). Beyond species: Functional diversity and the maintenance of ecological processes and services. *Journal of Applied Ecology*, 48, 1079–1087. https://doi.org/10.1111/j.1365-2664.2011.02048.x
- Cailliez, F. (1983). The analytical solution of the additive constant problem. *Psychometrika*, 48, 305–308. https://doi.org/10.1007/BF02294026
- Casey, J. M., Baird, A. H., Brandl, S. J., Hoogenboom, M. O., Rizzari, J. R., Frisch, A. J., ... Connolly, S. R. (2017). A test of trophic cascade theory: Fish and benthic assemblages across a predator density gradient on coral reefs. *Oecologia*, 183, 161–175. https://doi.org/10.1007/ s00442-016-3753-8
- Ceccarelli, D. M., Jones, G. P., & McCook, L. J. (2011). Interactions between herbivorous fish guilds and their influence on algal succession on a coastal coral reef. *Journal of Experimental Marine Biology and Ecology*, 399, 60–67. https://doi.org/10.1016/j.jembe.2011.01.019
- Chapin, F. S. III, Zavaleta, E. S., Eviner, V. T., Naylor, R. L., Vitousek, P. M., Reynolds, H. L., ... Díaz, S. (2000). Consequences of changing biodiversity. *Nature*, 405, 234–242. https://doi.org/10.1038/ 35012241
- Cheal, A. J., MacNeil, M. A., Cripps, E., Emslie, M. J., Jonker, M., Schaffelke, B., & Sweatman, H. (2010). Coral-macroalgal phase shifts or reef resilience: Links with diversity and functional roles of herbivorous fishes on the Great Barrier Reef. *Coral Reefs*, 29, 1005–1015. https://doi.org/10.1007/s00338-010-0661-y
- Chen, C. A. (1999). Analysis of scleractinian distribution in Taiwan indicating a pattern congruent with sea surface temperatures and currents: Examples from *Acropora* and Faviidae corals. *Zoological Studies*, 38, 119–129.
- Chen, C. (2004a). Checklist of the fishes of Penghu (Vol. 4). COA, Taiwan: Fisheries Research Institute (In Chinese).
- Chen, C. A., & Shashank, K. (2009). Taiwan as a connective steppingstone in the Kuroshio triangle and the conservation of coral ecosystems under the impacts of climate change. *Kuroshio Science*, 3, 15–22.
- Chen, H., Liao, Y. C., Chen, C. Y., Tsai, J. I., Chen, L. S., & Shao, K. T. (2015). Long-term monitoring dataset of fish assemblages impinged at nuclear power plants in northern Taiwan. *Scientific Data*, *2*, 150071. https://doi.org/10.1038/sdata.2015.71
- Chen, J. P. (2004b) Biodiversity and conservation of Kenting National Park fish fauna. Kenting Natl Park Rep (In Chinese).
- Chen, J. P., Jan, R., Kuo, J. W., Huang, C. H., & Chen, C. Y. (2009). Fish fauna around Green Island. J Natl Park, 19, 23-45 (In Chinese).
- Chen, J. P., Shao, K. T., Jan, R. Q., Kuo, J. W., & Chen, J. Y. (2010). Marine fishes in Kenting National Park. Kenting Natl Park Rep (In Chinese).
- Chong-Seng, K. M., Nash, K. L., Bellwood, D. R., & Graham, N. A. J. (2014). Macroalgal herbivory on recovering versus degrading coral reefs. *Coral Reefs*, 33, 409–419. https://doi.org/10.1007/ s00338-014-1134-5
- Clavel, J., Julliard, R., & Devictor, V. (2011). Worldwide decline of specialist species: Toward a global functional homogenization? *Frontiers in Ecology and the Environment*, 9, 222-228. https://doi. org/10.1890/080216
- D'Agata, S., Mouillot, D., Kulbicki, M., Andrefouet, S., Bellwood, D. R., Cinner, J. E., ... Vigliola, L. (2014). Human-mediated loss of phylogenetic and functional diversity in coral reef fishes. *Current Biology*, 24, 555–560. https://doi.org/10.1016/j.cub.2014.01.049
- Dai, C. F., & Horng, S. (2009). Scleractinia fauna of Taiwan I. The complex group. Taipei, China: National Taiwan University.
- Dai, C. F., Soong, K., Chen, C. A., Fan, T. Y., & Hsieh, H. (2005). Status of coral reefs of Taiwan in East Asian Seas region: 2004 (pp 153–164). Japan: Ministry of the Environment.

- Darling, E. S., Graham, N. A. J., Januchowski-Hartley, F. A., Nash, K. L., Pratchett, M. S., & Wilson, S. K. (2017). Relationships between structural complexity, coral traits, and reef fish assemblages. *Coral Reefs*, 36, 561–575. https://doi.org/10.1007/s00338-017-1539-z
- Denis, V., Ribas-deulofeu, L., Sturaro, N., Kuo, C. Y., & Chen, C. A. (2017). A functional approach to the structural complexity of coral assemblages based on colony morphological features. *Scientific Reports*, 7, 9849. https://doi.org/10.1038/s41598-017-10334-w
- Dulvy, N. K., Freckleton, R. P., & Polunin, N. V. C. (2004). Coral reef cascades and the indirect effects of predator removal by exploitation. *Ecology Letters*, 7, 410-416. https://doi. org/10.1111/j.1461-0248.2004.00593.x
- Ferreira, C. E. L., Floeter, S. R., Gasparini, J. L., Ferreira, B. P., & Joyeux, J. C. (2004). Trophic structure patterns of Brazilian reef fishes: A latitudinal comparison. *Journal of Biogeography*, 31, 1093–1106. https:// doi.org/10.1111/j.1365-2699.2004.01044.x
- Froese, R., & Pauly, D. (2018) FishBase. http://www.fishbase.org.
- Fung, T., Seymour, R. M., & Johnson, C. R. (2011). Alternative stable states and phase shifts in coral reefs under anthropogenic stress. *Ecology*, 92, 967–982. https://doi.org/10.1890/10-0378.1
- Goatley, C. H. R., Bonaldo, R. M., Fox, R. J., & Bellwood, D. R. (2016). Sediments and herbivory as sensitive indicators of coral reef degradation. *Ecology and Society*, 21, 29. https://doi.org/10.5751/ ES-08334-210129
- Graham, N. A. J., Jennings, S., MacNeil, M. A., Mouillot, D., & Wilson, S. K. (2015). Predicting climate-driven regime shifts versus rebound potential in coral reefs. *Nature*, 518, 94–97. https://doi.org/10.1038/ nature14140
- Graham, N. A. J., & Nash, K. L. (2013). The importance of structural complexity in coral reef ecosystems. *Coral Reefs*, 32, 315–326. https:// doi.org/10.1007/s00338-012-0984-y
- Graham, N. A. J., Wilson, S. K., Jennings, S., Polunin, N. V. C., Robinson, J., Bijoux, J. P., & Daw, T. M. (2007). Lag effects in the impacts of mass coral bleaching on coral reef fish, fisheries, and ecosystems. *Conservation Biology*, 21, 1291–1300. https://doi. org/10.1111/j.1523-1739.2007.00754.x
- Graham, N. A. J., Chabanet, P., Evans, R. D., Jennings, S., Letourneur, Y., MacNeil, M. A., ... Wilson, S. K. (2011). Extinction vulnerability of coral reef fishes. *Ecology Letters*, 14, 341–348. https://doi. org/10.1111/j.1461-0248.2011.01592.x
- Guillemot, N., Kulbicki, M., Chabanet, P., & Vigliola, L. (2011). Functional redundancy patterns reveal non-random assembly rules in a species-rich marine assemblage. *PLoS ONE*, *6*, e26735. https://doi. org/10.1371/journal.pone.0026735
- Halpern, B. S., & Floeter, S. R. (2008). Functional diversity responses to changing species richness in reef fish communities. *Marine Ecology Progress Series*, 364, 147–156. https://doi.org/10.3354/ meps07553
- Hillebrand, H. (2004). Strength, slope and variability of marine latitudinal gradients. Marine Ecology Progress Series, 273, 251–267. https://doi. org/10.3354/meps273251
- Hoey, A. S., & Bellwood, D. R. (2011). Suppression of herbivory by macroalgal density: A critical feedback on coral reefs? *Ecology Letters*, 14, 267–273. https://doi.org/10.1111/j.1461-0248.2010.01581.x
- Holstein, J. (2018). worms: Retriving aphia information from World Register of Marine Species. R package version 0.2.2. https:// CRAN.R-project.org/package=worms.
- Hughes, T. P., Barnes, M. L., Bellwood, D. R., Cinner, J. E., Cumming, G. S., Jackson, J. B. C., ... Scheffer, M. (2017). Coral reefs in the Anthropocene. *Nature*, 546, 82–90. https://doi.org/10.1038/ nature22901
- Hughes, T. P., Kerry, J. T., Baird, A. H., Connolly, S. R., Dietzel, A., Eakin, C. M., ... Torda, G. (2018). Global warming transforms coral reef assemblages. *Nature*, 556, 492–496. https://doi.org/10.1038/ s41586-018-0041-2

9

- Johnson, K. G., Jackson, J. B. C., & Budd, A. F. (2008). Caribbean reef development was independant of coral diversity over 28 million years. *Science*, 319, 1521–1523.
- Jompa, J., & McCook, L. J. (2002). The effects of nutrients and herbivory on competition between a hard coral (*Porites cylindrica*) and a brown alga (*Lobophora variegata*). *Limnology and Oceanography*, 47, 527–534.
- Kleypas, J. A., McManus, J. W., & Meñez, L. A. B. (1999). Environmental limits to coral reef development: Where do we draw the line? *American Zoologist*, 39, 146–159. https://doi.org/10.1093/icb/39.1.146
- Kulbicki, M., Parravicini, V., Bellwood, D. R., Arias-Gonzàlez, E., Chabanet, P., Floeter, S. R., ... Mouillot, D. (2013). Global biogeography of reef fishes: A hierarchical quantitative delineation of regions. *PLoS ONE*, *8*, e81847. https://doi.org/10.1371/journal.pone.0081847
- Laliberté, E., & Legendre, P. (2010). A distance-based framework for measuring functional diversity from multiple traits. *Ecology*, 91, 299– 305. https://doi.org/10.1890/08-2244.1
- Laliberté, E., Legendre, P., & Shipley, B. (2014). FD: Measuring functional diversity from multiple traits, and other tools for functional ecology. R package version 1.0-12.
- Liu, P., Shao, K. T., Jan, R. Q., Fan, T. Y., Wong, S. L., Hwang, J. S., ... Lin, H. J. (2009). A trophic model of fringing coral reefs in Nanwan Bay, southern Taiwan suggests overfishing. *Marine Environment Research*, 68, 106–117. https://doi.org/10.1016/j.marenvres.2009.04.009
- Maechler, M., Rousseeuw, P., Struyf, A., Hubert, M., & Hornik, K. (2017). cluster: Cluster analysis basics and extensions. R package version 2.0.6.
- Mason, N. W. H., Mouillot, D., Lee, W. G., & Wilson, J. B. (2005). Functional richness, functional evenness and functional divergence: The primary components of functional diversity. Oikos, 111, 112– 118. https://doi.org/10.1111/j.0030-1299.2005.13886.x
- McCann, K. S. (2000). The diversity-stability debate. *Nature*, 405, 228–233. https://doi.org/10.1038/35012234
- McClanahan, T. R. (1994). Coral-eating snail Drupella cornus population increases in Kenyan coral reef lagoons. Marine Ecology Progress Series, 115, 131–138. https://doi.org/10.3354/meps115131
- McWilliam, M., Hoogenboom, M. O., Baird, A. H., Kuo, C.-Y., Madin, J. S., & Hughes, T. P. (2018). Biogeographical disparity in the functional diversity and redundancy of corals. *Proceedings of the National Academy* of Sciences United States of America, 115, 3084–3089. https://doi. org/10.1073/pnas.1716643115
- Mora, C., Chittaro, P. M., Sale, P. F., Kritzer, J. P., & Ludsin, S. A. (2003). Patterns and processes in reef fish diversity. *Nature*, 421, 933–936. https://doi.org/10.1038/nature01393
- Mouillot, D., Bellwood, D. R., Baraloto, C., Chave, J., Galzin, R., Harmelin-Vivien, M., ... Thuiller, W. (2013a). Rare species support vulnerable functions in high-diversity ecosystems. *PLoS Biology*, 11, e1001569.
- Mouillot, D., Graham, N. A. J., Villéger, S., Mason, N. W. H., ... D. R. (2013b). A functional approach reveals community responses to disturbances. *Trends in Ecology & Evolution*, 28, 167–177.
- Mouillot, D., Villéger, S., Parravicini, V., Kulbicki, M., Arias-González, J. E., Bender, M., ... Bellwood, D. R. (2014). Functional over-redundancy and high functional vulnerability in global fish faunas on tropical reefs. Proceedings of the National Academy of Sciences United States of America, 111, 13757–13762. https://doi.org/10.1073/ pnas.1317625111
- Munday, P. L. (2004). Habitat loss, resource specialization, and extinction on coral reefs. *Global Change Biology*, 10, 1642–1647. https://doi. org/10.1111/j.1365-2486.2004.00839.x
- Paradis, E., Claude, J., & Strimmer, K. (2004). APE: Analyses of phylogenetics and evolution in R language. *Bioinformatics*, 20, 289–290. https://doi.org/10.1093/bioinformatics/btg412
- Parravicini, V., Villéger, S., McClanahan, T. R., Arias-González, J. E., Bellwood, D. R., Belmaker, J., ... Mouillot, D. (2014). Global mismatch between species richness and vulnerability of reef fish assemblages. *Ecology Letters*, 17, 1101–1110. https://doi.org/10.1111/ele.12316

- Petchey, O. L., & Gaston, K. J. (2006). Functional diversity: Back to basics and looking forward. *Ecology Letters*, 9, 741–758. https://doi. org/10.1111/j.1461-0248.2006.00924.x
- Pratchett, M. S., Hoey, A. S., & Wilson, S. K. (2014). Reef degradation and the loss of critical ecosystem goods and services provided by coral reef fishes. *Current Opinion in Environmental Sustainability*, *7*, 37-43. https://doi.org/10.1016/j.cosust.2013.11.022
- Pratchett, M. S., Hoey, A. S., Wilson, S. K., Messmer, V., & Graham, N. A. J. (2011). Changes in biodiversity and functioning of reef fish assemblages following coral bleaching and coral loss. *Diversity*, *3*, 424–452. https://doi.org/10.3390/d3030424
- Purvis, A., & Hector, A. (2000). Getting the measure of biodiversity. *Nature*, 405, 212-219. https://doi.org/10.1038/35012221
- Quimbayo, J. P., Mendes, T. C., Kulbicki, M., Floeter, S. R., & Zapata, F. A. (2017). Unusual reef fish biomass and functional richness at Malpelo, a remote island in the Tropical Eastern Pacific. *Environmental Biology of Fishes*, 100, 149-162. https://doi. org/10.1007/s10641-016-0557-y
- R Core Team. (2017). R: A language and environment for statistical computing. Vienna, Austria: R Foundation for Statistical Computing. https:// www.R-project.org/.
- Raymundo, L. J., Halford, A. R., Maypa, A. P., & Kerr, A. M. (2009). Functionally diverse reef-fish communities ameliorate coral disease. Proceedings of the National Academy of Sciences United States of America, 106, 17067–17070. https://doi.org/10.1073/ pnas.0900365106
- Ribas-Deulofeu, L., Denis, V., De Palmas, S., Kuo, C. Y., Hsieh, H. J., & Chen, C. A. (2016). Structure of benthic communities along the Taiwan latitudinal gradient. *PLoS ONE*, 11, e0160601. https://doi. org/10.1371/journal.pone.0160601
- Richardson, L. E., Graham, N. A. J., Pratchett, M. S., Eurich, J. G., & Hoey, A. S. (2018). Mass coral bleaching causes biotic homogenization of reef fish assemblages. *Global Change Biology*, 24, 3117–3129. https:// doi.org/10.1111/gcb.14119
- Shao, K. T. (2018). The fish database of Taiwan. http://fishdb.sinica.edu. tw.
- Shao, K. T., Chen, J. P., & Wang, S. C. (1999). Biogeography and database of marine fishes in Taiwan waters. Proceedings of the 5th Indo-Pacific Fish Conference, 673–680.
- Shao, K. T., Ho, H. C., Lin, P. L., Lee, P. F., Lee, M. Y., Tsai, C. Y., ... Yeh, H. M. (2008). A checklist of the fishes of southern Taiwan, northern South China Sea. *Raffles Bulletin of Zoology*, 19, 233–271.
- Spalding, M. D., Fox, H. E., Allen, G. R., Davidson, N., Ferdaña, Z. A., Finlayson, M., ... Robertson, J. (2007). Marine ecoregions of the world: A bioregionalization of coastal and shelf areas. *BioScience*, 57, 573–583. https://doi.org/10.1641/B570707
- Strong, D. R. (1992). Are trophic cascades all wet? Differentiation and donor-control in speciose ecosystems. *Ecology*, 73, 747–754. https:// doi.org/10.2307/1940154
- Stuart-Smith, R. D., Brown, C. J., Ceccarelli, D. M., & Edgar, G. J. (2018). Ecosystem restructuring along the Great Barrier Reef following mass coral bleaching. *Nature*, 560, 92–96. https://doi.org/10.1038/ s41586-018-0359-9
- Stuart-Smith, R. D., Bates, A. E., Lefcheck, J. S., Duffy, J. E., Baker, S. C., Thomson, R. J., ... Edgar, G. J. (2013). Integrating abundance and functional traits reveals new global hotspots of fish diversity. *Nature*, 501, 539–542. https://doi.org/10.1038/nature12529
- Tebbett, S. B., Goatley, C. H. R., & Bellwood, D. R. (2017). Clarifying functional roles: Algal removal by the surgeonfishes *Ctenochaetus striatus* and *Acanthurus nigrofuscus*. *Coral Reefs*, 36, 803–813. https://doi. org/10.1007/s00338-017-1571-z
- Tittensor, D. P., Mora, C., Jetz, W., Lotze, H. K., Ricard, D., Vanden, B. E., & Worm, B. (2010). Global patterns and predictors of marine biodiversity across taxa. *Nature*, 466, 1098–1101. https://doi.org/10.1038/ nature09329

- Villéger, S., Brosse, S., Mouchet, M., Mouillot, D., & Vanni, M. J. (2017). Functional ecology of fish: Current approaches and future challenges. *Aquatic Sciences*, 79, 783–801. https://doi.org/10.1007/ s00027-017-0546-z
- Walther, G., Post, E., Convey, P., Menzel, A., Parmesan, C., Beebee, T. J. C., ... Bairlein, F. (2002). Ecological responses to recent climate change. *Nature*, 416, 389–395. https://doi.org/10.1038/416389a
- Wang, J., & Chern, C. S. (1988). On the Kuroshio branch in the Taiwan strait during wintertime. Progress in Oceanography, 21, 469–491. https://doi.org/10.1016/0079-6611(88)90022-5
- Wilson, S. K., Fisher, R., Pratchett, M. S., Graham, N. A. J., Dulvy, N. K., Turner, R. A., ... Polunin, N. V. C. (2010). Habitat degradation and fishing effects on the size structure of coral reef fish communities. *Ecological Applications*, 20, 442–451. https://doi.org/10.1890/08-2205.1
- Wong, J. S. Y., Chan, Y. K. S., Ng, C. S. L., Tun, K. P. P., Darling, E. S., & Huang, D. (2018). Comparing patterns of taxonomic, functional and phylogenetic diversity in reef coral communities. *Coral Reefs*, 37, 737– 750. https://doi.org/10.1007/s00338-018-1698-6

WoRMS Editorial Board. (2018). World Register of Marine Species. Available from http://www.marinespecies.org at VLIZ.

#### SUPPORTING INFORMATION

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How to cite this article: Denis V, Chen J-W, Chen Q, et al. Biogeography of functional trait diversity in the Taiwanese reef fish fauna. *Ecol Evol*. 2018;00:1–11. <u>https://doi.</u> org/10.1002/ece3.4771