

1 **Biogenic volatile organic compound emissions from senescent**
2 **maize leaves and a comparison with other leaf developmental**
3 **stages**

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22 **Abstract**

23 Plants are the major source of Biogenic Volatile Organic Compounds (BVOCs) which have a
24 large influence on atmospheric chemistry and the climate system. Therefore, understanding of
25 BVOC emissions from all abundant plant species at all developmental stages is very important.
26 Nevertheless, investigations on BVOC emissions from even the most widespread agricultural
27 crop species are rare and mainly confined to the healthy green leaves. Senescent leaves of grain
28 crop species could be an important source of BVOCs as almost all the leaves senesce on the field
29 before being harvested. For these reasons, BVOC emission measurements have been performed
30 on maize (*Zea mays* L.), one of the most cultivated crop species in the world, at all the leaf
31 developmental stages. The measurements were performed in controlled environmental conditions
32 using dynamic enclosures and proton transfer reaction mass spectrometry (PTR-MS). The main
33 compounds emitted by senescent maize leaves were methanol (31% of the total cumulative
34 BVOC emission on a mass of compound basis) and acetic acid (30%), followed by acetaldehyde
35 (11%), hexenals (9%) and *m/z* 59 compounds (acetone/propanal) (7%). Important differences
36 were observed in the temporal emission profiles of the compounds, and both yellow leaves
37 during chlorosis and dry brown leaves after chlorosis were identified as important senescence-
38 related BVOC sources. Total cumulative BVOC emissions from senescent maize leaves were
39 found to be among the highest for senescent Poaceae plant species. BVOC emission rates varied
40 strongly among the different leaf developmental stages, and senescent leaves showed a larger
41 diversity of emitted compounds than leaves at earlier stages. Methanol was the compound with
42 the highest emissions for all the leaf developmental stages and the contribution from the young-
43 growing, mature, and senescent stages to the total methanol emission by a typical maize leaf was
44 61, 13, and 26%, respectively. This study shows that BVOC emissions from senescent maize
45 leaves cannot be neglected and further investigations in field conditions are recommended to
46 further constrain the BVOC emissions from this important C4 crop species.

47

48

49 **1 Introduction**

50

51 Terrestrial vegetation is a huge source of volatile organic compounds (VOCs) in the Earth's
52 atmosphere. Besides playing a role in plant biology and ecology (Pierik et al., 2014), biogenic
53 VOCs (BVOCs) are generally highly reactive with the major atmospheric oxidants, thus
54 affecting the oxidation capacity of the atmosphere, air quality and climate (Atkinson, 2000;
55 Laothawornkitkul et al., 2009; Pacifico et al., 2009). BVOC emissions from plants are highly
56 species-specific and not only depend on environmental conditions (Guenther et al., 2012) and
57 abiotic or biotic stress factors (Holopainen et al., 2010) but also on plant ontogeny (Bracho-
58 Nunez et al., 2011). Therefore they should be investigated at all developmental stages, including
59 senescence.

60 Leaf senescence is the final stage of leaf development. It is a complex energy-dependent self-
61 digesting process that facilitates the remobilisation of nutrients from the senescing leaf to
62 growing vegetative plant organs or developing seeds and fruits, where they are reused for
63 biosynthesis (Woo et al., 2013; Keskitalo, 2005; Gan and Amasino, 1997; Taiz et al., 2015).
64 Under normal growing conditions leaf senescence is governed by the developmental age of the
65 leaves, which is a function of hormones and other regulatory factors (Taiz et al., 2015). Under
66 unfavourable environmental conditions (e.g. drought or enhanced ozone concentrations) or biotic
67 stress (e.g. pathogen infestation), however, the leaf senescence process can occur prematurely.
68 Three main phases are generally distinguished in the developmental leaf senescence process
69 (Taiz et al., 2015). The initiation phase is characterised by a gradual decline in photosynthesis
70 and a transition of the leaf from being a nitrogen sink to a nitrogen source. Self-digestion of
71 cellular constituents and macromolecules mainly occurs during the second phase, the
72 degenerative phase. The third phase, the terminal phase, is characterised by loss of cellular
73 integrity, cell death, and finally (in most cases) leaf abscission.

74 Although numerous studies have already been performed on BVOC emissions from healthy
75 and growing leaves where cells were developing (Kuhn et al., 2002; Harley et al., 2007; Hüve et
76 al., 2007; Folkers et al., 2008; Bracho-Nunez et al., 2011; Mozaffar et al., 2017), studies on
77 senescent leaves where cells are breaking down (Gan and Amasino, 1997) are very rare. As far

78 as we know, there is only one leaf-scale study, performed under controlled conditions
79 (Holopainen et al., 2010), in which VOC emissions from undetached senescent leaves (of *Betula*
80 *pendula Roth*) have been measured, but the measurement frequency was too low (1 Gas
81 Chromatography – Mass Spectrometry (GC-MS) sample every 3 days) to adequately represent
82 the emission dynamics. Therefore, additional studies at increased time resolution are required for
83 a better characterisation of BVOC exchanges between senescent leaves and the atmosphere
84 during the whole senescence period.

85 To assess the importance of BVOC emission rates from the senescent leaves of a plant,
86 information about BVOC emission rates from other leaf developmental stages (young, semi-
87 mature, mature) is also necessary. In this study we will mainly focus on BVOC emissions from
88 senescent maize (*Zea Mays L.*) leaves, but we will also compare them with BVOC emissions
89 from other developmental stages of maize leaves measured under the same environmental
90 conditions. Despite being a vastly cultivated crop species worldwide, only a few literature
91 studies have been devoted to BVOC emissions for this species (MacDonald and Fall, 1993; Das
92 et al., 2003; Graus et al., 2013; Bachy et al., 2016; Mozaffar et al., 2017) and none of them cover
93 all the leaf developmental stages. In particular, data on BVOC emission rates from senescent
94 maize leaves are missing in the abovementioned literature.

95 Maize is a monocarpic (a plant which only flowers and bears fruit once in its lifetime)
96 herbaceous C4 plant for which whole plant senescence occurs with seed maturation (Lim et al.,
97 2007). However, the first leaf at the base of the plant starts senescing long before flowering and
98 this process continues for all the leaves from the base to the top of the plant throughout the
99 growing season. Therefore, emissions from senescent leaves could provide a significant
100 contribution to the total BVOC emission budget from a maize field, as suggested by de Gouw et
101 al. (2000).

102 In order to improve the knowledge on BVOC emissions from this important crop species we
103 aim to provide answers to the following specific questions: 1) which BVOCs are emitted during
104 the senescence process and in what proportions, 2) how do BVOC emissions from senescent
105 maize leaves compare to those from other species of the Poaceae family, 3) how do BVOC
106 emission rates vary among the different leaf developmental stages of maize, and 4) what are the

107 contributions of the different developmental stages to the total emission of individual BVOC
108 compounds by a maize leaf/plant.

109 **2 Materials and methods**

110 ***2.1 Plants and environmental conditions***

111 The experiments were performed on maize leaves (*Zea mays* L., variety Prosil, Caussade
112 Semences, France) at four different leaf developmental stages: young, semi-mature, mature, and
113 senescent. To measure BVOC exchanges between young leaves and the atmosphere, 8 to 14 day
114 old maize plants (age counting began at seed germination) were used. The shoots of the young
115 maize plants were completely enclosed because it was not feasible to enclose a single young leaf
116 for a sufficiently long period without damaging it due to the fast elongation rate of both leaves
117 and stem. The upper part (ca. 55 cm starting from the tip) of almost fully developed, healthy 7th
118 leaves of 30-40 day old plants (around 120 cm tall) and the fully developed (7th, 8th or 9th) leaves
119 of 60-70 day old, fully grown plants (around 180 cm tall) were enclosed to measure BVOC
120 emission rates from semi-mature and mature leaves during 5 consecutive days, respectively. The
121 above-mentioned leaf numbers refer to the order in which new leaves appear during plant
122 development and leaf numbering thus starts from the base. The choice for (mainly) 7th leaves
123 was determined by their position with respect to the ground and the ceiling level of the
124 environmental chamber and the possibility to enclose them without damaging them. Senescence-
125 induced BVOC emission rates were measured from senescing (7th, 8th or 10th) leaves of fully
126 grown 60-95 day old plants. The upper half of those leaves was enclosed a few days before the
127 start of the senescence initiation phase, which was identified by a decrease of photosynthesis
128 (Taiz et al., 2015) and preceded chlorosis, i.e. leaf yellowing due to chlorophyll degradation, by
129 a few days. Chlorosis started at the tip of the leaf and gradually progressed towards the base of
130 the enclosed leaf within 5-10 days (range for 10 replicates). The onset and evolution of leaf
131 chlorosis, however, could only be determined visually and not by a quantitative metric (e.g.
132 photochemical efficiency or chlorophyll content). When describing the BVOC emission rate
133 dynamics in Section 3.1, the term “chlorosis period” refers to the period between the first visual
134 signs of yellowing and the total absence of green areas on the enclosed part of the leaf. Whereas

135 the period between the onset of the decrease of photosynthesis and the onset of chlorosis can be
136 considered as the initiation phase of leaf senescence, the chlorosis period largely coincides with
137 the degenerative phase. The terminal phase can be mainly associated with the post-chlorosis
138 period (from the end of the chlorosis period to the end of the measurements). It is to be noted that
139 the third phase of senescence in maize does not imply leaf abscission. In order to support the
140 description of the temporal evolution of senescence and BVOC emissions in this Section and in
141 subsections 3.1.1 and 3.1.2, daily pictures of a typical senescent leaf (taken at 10 AM every day),
142 along with the temporal evolution of the daily net photosynthesis rate and the methanol and
143 acetic acid daily emission rates are shown in Fig. 1.

144

145 Figure 1

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147 Investigations on young, semi-mature, and mature leaves were performed on 5 replicates for
148 each stage, and 10 replicates were used for investigating emissions from senescent leaves. More
149 details about the enclosed leaves at the different leaf developmental stages and the associated
150 plant age and developmental status can be found in Table 1.

151 All these experiments were conducted on potted (20 L pots filled with a mixture of 75% silty
152 clay loam soil and 25% sand) plants in a temperature and light controlled environmental
153 chamber. Details about the environmental chamber can be found in Mozaffar et al. (2017). The
154 soil was fertilised before planting with NPK fertiliser (6-5-5, Substral Nutrimax, Belgium)
155 containing micronutrients (Cu, Fe, Mn, Mo, Zn). Daytime temperature in the environmental
156 chamber was maintained at 25°C throughout the experimental period and during the night the
157 temperature decreased by around 2°C due to the absence of a heat source. Seventeen hours of
158 photoperiod at three different PPFD (Photosynthetic Photon Flux Density) values (100, 330 and
159 600 $\mu\text{mol m}^{-2} \text{s}^{-1}$) were alternated with seven hours of darkness (Fig. 2). All the plants were
160 watered regularly to maintain a good soil moisture content (35-40%, field capacity was 40%).

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163 **Table 1:** Detailed information about the enclosed leaf/leaves at the different leaf developmental
 164 stages, the associated plant age and two useful metrics, BBCH code and AGDD (Accumulated
 165 Growing Degree-Days), for the plant developmental stages at which the experiments were
 166 carried out. The BBCH code and AGDD are explained in supplement S1.

Leaf developmental stage	Enclosed leaf/leaves	Plant age [days]	BBCH code *	AGDD [degree-days]
Young	first 2-5 leaves	4-14	10-14	122-274
Semi-mature (almost fully grown)	7th leaf	30-40	17-19	518-671
Mature (fully grown)	7th/8th/9th leaf	60-70	65-69	976-1128
Senescent	7th/8th/10 th leaf	60-95	65-69*	976-1510

167 * range of BBCH codes at which the onset of senescence was identified for the enclosed leaves.

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170 **2.2 Experimental setup**

171 Young maize shoots were enclosed in 22 L cylindrical dynamic flow-through enclosures made of
 172 transparent PFA (perfluoroalkoxy Teflon) envelope (Norton, Saint-Gobain Performance Plastics,
 173 NJ, USA). Similar 30 L enclosures were used to partially enclose the semi-mature, mature, and
 174 senescent leaves. In addition, a similar empty enclosure was used to measure the background
 175 VOC concentrations in the air that was supplied to the enclosures. The enclosures were equipped
 176 with a Teflon fan, a thermistor (type 10k, NTC, Omega, UK), and a relative humidity sensor
 177 (type HIH-3610, Honeywell, NJ, USA) for proper mixing of air and continuous monitoring of
 178 leaf temperature, air temperature and relative humidity, respectively. Relative humidity and CO₂
 179 concentration in the 5 SLM (standard litre per minute) air flow sent through each of the
 180 enclosures were controlled at around 40% and 400 ppm, respectively. A conventional high-
 181 sensitivity quadrupole-based Proton Transfer Reaction-Mass Spectrometry instrument (hs-PTR-

182 MS, Ionicon Analytik GmbH, Innsbruck, Austria) and a LI-7000 non-dispersive IR gas analyser
183 (LI-COR, Lincoln, Nebraska, USA) were used to determine BVOC, H₂O and CO₂
184 concentrations. In general, BVOC and H₂O calibrations were performed every 3-5 days and CO₂
185 calibrations were performed monthly. Further details about the set-up have been presented in
186 Mozaffar et al. (2017).

187 ***2.3 Tentative identification and quantification of emitted BVOCs***

188 The operating principle and technical details of the PTR-MS technique have been amply
189 described in the literature (e.g. Lindinger et al., 1998; Ellis and Mayhew, 2013). In the present
190 work, the quadrupole-based hs-PTR-MS instrument was operated at a drift tube pressure and
191 temperature of 2.1 hPa and 333 K, respectively, and at an E/N value (ratio of the electric field to
192 the air number density in the drift tube) of 130 Td (1 Td = 10⁻¹⁷ V cm²). Measurements were
193 carried out in the Multiple Ion Detection mode, in which the instrument continuously cycled
194 through a list of preset m/z values of BVOC-related ion species for which signal intensities
195 above the detection limit have been observed in PTR-MS spectra of maize leaf emissions. The
196 m/z values of ions which were used for emission rate quantification, along with a tentative
197 identification of the associated BVOCs are shown in Table 2. This identification mainly relies on
198 Karl et al. (2005) who unambiguously identified BVOCs released from drying rice (*O. sativa*) by
199 using the hyphenated GC/PTR-MS technique. In addition to the BVOC-related ions, isotopes of
200 the reactant ion species H₃O⁺ (at m/z 21) and its first hydrate H₃O⁺•H₂O (at m/z 39) were
201 monitored.

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208 Table 2: The m/z ratios and list of potential contributory compounds for which emission rates
 209 were quantified. Compounds that were present in the multi-component calibration bottles are
 210 indicated in bold and their mixing ratios ($\pm 5\%$ accuracy) in the bottles are also shown.

m/z	Compound	Mixing ratio (ppbv)	
		Bottle 1 ^a	Bottle 2 ^a
33	Methanol	1074	1013
45	Acetaldehyde	999	969
57	Hexenals		1041 ^b
59	Acetone , propanal, 2,3-butanedione	973	502
61	Acetic acid		
69	Isoprene , pentenol, n-pentanal	441	483
73	Methyl ethyl ketone (MEK), 2-methyl-propanal, butanal	494	516
81	Monoterpenes ,	851 ^c	
	Hexenals		1041 ^b
83	Hexenols , n-hexanal, hexenyl acetates		986 ^d
99	Hexenals		1041 ^b
137	Monoterpenes	851 ^c	

211

212 ^a Calibration bottle 1 was used for the experiments with young, semi-mature and mature leaves.
 213 Bottle 2 was used for the senescence experiments.

214 ^b Trans-2-hexenal was used as a proxy for hexenals and was the only compound in calibration
 215 bottle 2 which resulted in product ions at m/z 81.

216 ^c Alpha-pinene (452 ppbv) and sabinene (399 ppbv) were used as proxies for the monoterpene
 217 family and were the only compounds in calibration bottle 1 which resulted in product ions at m/z
 218 81.

219 ^d Only the cis-3-hexenol isomer was present in the calibration bottle

220

221 PTR-MS calibrations were performed by using two gravimetrically prepared multi-component
222 mixtures of BVOCs in nitrogen (Apel-Riemer Environmental Inc., Denver, CO, USA) with a
223 certified accuracy of 5%. The calibration gas was further diluted with zero air, generated by
224 sending ambient air through a catalytic converter (Parker® ChromGas® Zero Air Generator,
225 model 1001, Parker Hannifin Corporation, Haverhill, MA, USA), to obtain mixing ratios in the
226 lower ppbv range. When several compounds potentially contributed to the ion signal at a given
227 nominal m/z value (e.g. methyl ethyl ketone, 2-methyl-propanal and butanal at m/z 73), it was
228 assumed that they possessed similar calibration factors. By calibrating the PTR-MS to the
229 compound indicated in bold in Table 2 (methyl ethyl ketone for the above-mentioned example)
230 and applying the calibration factor to the measured ion signal at that m/z, we estimated the sum
231 of the concentrations of those compounds.

232 The H_3O^+ /monoterpene reaction in the drift tube reactor of the PTR-MS is known to result
233 mainly in the protonated molecule at m/z 137 and a fragment at m/z 81 and the product ion
234 distribution is isomer-dependent (Tani et al., 2003). However, since the rate constant for proton
235 transfer from H_3O^+ to monoterpenes is isomer-independent, the sum of the ion signals at m/z 81
236 and 137, weighted for their relative transmission in the mass spectrometer, is expected to be
237 isomer-independent. This sum was used to determine the emission rate of all monoterpene
238 isomers emitted by the young, semi-mature, and mature leaves. During the measurements with
239 senescent leaves the ion signals at m/z 137 were always below the detection limit, even when ion
240 signals at m/z 81 were maximal (I_{137}/I_{81} at maximum I_{81} is 0.01 ± 0.02 with $I_{137(81)}$ being the net
241 ion signal at m/z 137(81)). This indicates that monoterpene emission rates were not significant
242 during this stage and that ions at m/z 81 did not result from dissociative proton transfer of
243 monoterpenes but corresponded to dehydrated protonated hexenals. In order to obtain accurate
244 emission rates for the sum of hexenal isomers, the other main proton transfer product ions of
245 hexenals at m/z 57 and m/z 99 (Pang, 2015) were also continuously monitored, both during the
246 enclosure measurements and calibrations, and the sum of the signals of the three product ion
247 species, weighted for ion transmission in the mass spectrometer, was used to derive the emission
248 rate of the sum of hexenal isomers. As hexenals and monoterpenes both result in a product ion at
249 m/z 81, calibration of the PTR-MS for those compounds required the use of two multi-
250 component calibration mixtures containing either a hexenal or a monoterpene, in addition to the

251 other compounds to be quantified (Table 2). The calibration mixture containing the
252 monoterpenes was used for the experiments on young, mature, and semi-mature leaves, whereas
253 the mixture containing hexenal was used for the experiments on senescent leaves.

254 The calibration factor for acetic acid (at m/z 61), a compound which was not present in the
255 calibration bottles, was estimated from the factor for acetone (at m/z 59) by considering the
256 fragmentation of the protonated molecules in the drift tube (Inomata and Tanimoto, 2010;
257 Schwarz et al., 2009) and the ratio of the calculated collision rate constants for the proton
258 transfer reactions (Su, 1994), and by assuming the same PTR-MS transmission efficiency for
259 ions at m/z 59 and 61. The m/z 81 and 83 compounds emitted by the senescent leaves are
260 associated to C6 volatiles (hexenals, hexanal, hexenols, hexenyl acetates) and will be designated
261 as green leaf volatiles (GLVs) hereafter.

262 **2.4 Emission rate computation**

263 Unless explicitly mentioned otherwise, BVOC emission rates E_{BVOC} from maize leaves are
264 expressed as mass flow rates per unit of leaf dry weight [μg of the compound $\text{g}_{\text{DW}}^{-1} \text{s}^{-1}$]. They are
265 calculated according to Eq. 1 in which F_{air} is the molar flow rate of the air which is supplied to
266 the enclosure, M_{BVOC} is the molar mass of the BVOC, DW is the dry weight of the enclosed (part
267 of the) leaf, and $x_{BVOC,PE}$ and $x_{BVOC,RE}$ are the mole fractions of the BVOC in the sampled air
268 from the plant enclosure and the empty reference enclosure, respectively.

$$E_{BVOC} = \frac{F_{\text{air}} \times (x_{BVOC,PE} - x_{BVOC,RE}) \times M_{BVOC}}{DW} \quad (1)$$

269 The mole fractions of the BVOC are obtained by dividing the normalised background-corrected
270 PTR-MS BVOC ion signals $I_{m/z}$, expressed in normalised counts per second (ncps), by the
271 BVOC calibration coefficients C_{BVOC} [ncps ppbv⁻¹] (Mozaffar et al., 2017).

272 The dry weight of the leaves was determined at the end of the experiments after at least 48
273 hours of drying in an oven at 75°C until all water was evaporated and a constant weight was
274 reached. Photosynthesis and transpiration rates were obtained in a similar way but are expressed
275 in molar units and per unit of leaf area [$\text{mol m}^{-2} \text{s}^{-1}$]. Leaf area was estimated as described in
276 Mozaffar et al. (2017).

277

278 **3 Results and discussion**

279 ***3.1 BVOC emissions from senescent maize leaves***

280 Significant emissions of methanol, acetaldehyde, m/z 59 compounds, acetic acid, and hexenals
281 were observed from senescent maize leaves and their temporal evolution is shown in Fig. 2 for a
282 single leaf.

283

284 Figure 2

285

286 Although the plants were grown and investigated under the same environmental conditions, a
287 considerable variability was noticed in the temporal evolution of the emissions among the
288 different replicates as is shown in Figures S1a and S1b in Supplement S1. However, the BVOC
289 emission dynamics from the investigated leaves show similar characteristics, as will be described
290 in the following subsections. Small emission rates of m/z 69, m/z 73 and m/z 83 compounds
291 were also noticed (Table 3), but they often barely exceeded the detection limit. The emergence of
292 BVOC emissions or their increase (in case the BVOC was also emitted constitutively) took place
293 at the start of the degeneration phase. Photosynthesis and transpiration diminished gradually as
294 chlorosis progressed from the apex towards the base of the leaf and finally stopped 2-4 days
295 before the end of the chlorosis period (as shown in Fig. 2 for one replicate). While some of the
296 BVOCs were predominantly emitted during and just after the degeneration phase, others were
297 emitted during the termination phase of senescence as well, as will be discussed in detail in the
298 following sections.

299 The temporal emission dynamics of the individual BVOC species, the relative contribution of
300 their cumulative emissions to the total emission, as well as a comparison with cumulative
301 emissions for other senescent species of the Poaceae family will be discussed hereafter.

302 **3.1.1 Methanol, acetaldehyde and GLVs are predominantly emitted during the**
303 **degeneration phase of leaf senescence**

304 At the beginning of the chlorosis period, the emission rates of methanol, acetaldehyde, and
305 hexenals rose slowly as leaf discoloration moved from the tip to the base and they increased
306 faster when the leaf was turning brown and was shrinking due to the drying process. The highest
307 emission rates occurred at around 1-3, 0-2 and 0-2 days before the end of the chlorosis period,
308 respectively (Fig. 2 and Fig. S1a). By then photosynthesis and transpiration from the steadily
309 decreasing green part of the enclosed leaf were already greatly reduced. Emissions of methanol,
310 acetaldehyde, and hexenals exceeding 20% of their maximum value lasted for 3-8, 3-7, and 2-7
311 days, respectively. The temporal evolution of normalised m/z 83 compound emission rates was
312 similar to that of normalised hexenal emission rates (data not shown). The variability in the
313 duration of the high emission periods for the different compounds reflects the variability in
314 temporal emission profiles for the 10 replicates (Fig. S1a in Supplement S1) . Figure 2 also
315 shows small methanol and acetaldehyde emissions after the chlorosis period which persisted
316 until the end of the measurements. These emissions showed a fixed diurnal pattern with a
317 maximum emission rate of at most 10% of the maximum emission rate during the chlorosis
318 period.

319 As can be seen in Fig. 2 and in Fig. S1a, no distinct diurnal patterns were observed for
320 methanol, acetaldehyde, and GLV emission rates during the chlorosis period and there was no
321 clear correlation between these emission rates and the environmental conditions (temperature,
322 PPF) in the dynamic enclosures. This is in agreement with previous observations of emissions
323 of these compounds from drying leaves (Fall et al., 1999; Warneke et al., 1999; Warneke et al.,
324 2002).

325 The link between the evolution of the senescent leaf and the temporal emission profile
326 suggests that the biochemical and physical mechanisms involved in leaf chlorosis and drying are
327 major drivers of BVOC emissions during leaf senescence. Presumably collapse of the cellular
328 structure during drying (Karl et al., 2001a; Karl et al., 2001c; de Gouw et al., 1999) and
329 disintegration of cell organelles and dying cells (Keskitalo et al., 2005) are at the origin of high
330 emissions of these compounds. Ozuna et al. (1985) reported that the stomata of senescent

331 *Nicotiana glauca* leaves which already showed marked chlorosis (the degree of chlorosis varied
332 from 60 to 90% of the total leaf surface area) remained closed even in the presence of light. As
333 maximum emissions of methanol and acetaldehyde in our experiments were measured close to
334 the end of the chlorosis period, the main emission route for those compounds during the
335 degeneration phase of leaf senescence was therefore probably no longer diffusion through the
336 stomata, as is the case for healthy leaves (MacDonald and Fall, 1993; Kreuzwieser et al., 2000;
337 Rottenberger et al., 2004). As changes in the biomechanical properties of the epidermis cuticle
338 have been reported for *Sonneratia alba* leaves (Takahashi et al., 2012) during this senescence
339 phase, diffusion of those compounds through a degraded epidermal layer might be considered as
340 a potential emission route.

341

342 In leaves, methanol is mainly produced by pectin demethylation in the cell wall, catalysed by
343 pectin methylesterase (PME) enzymes (Fall and Benson, 1996; Fall 2003). This process occurs
344 during growth as well as during aging and senescence of plant tissues (Nemecek-Marshall et al.,
345 1995). Due to continuous breakdown of cellular materials for nutrient remobilisation during the
346 degeneration phase of senescence (Woo et al., 2013; Keskitalo, 2005), massive changes occur in
347 the primary cell wall structure, which potentially stimulates massive methanol production.
348 Methanol emissions from the dead, brown leaves during the termination phase could be either
349 due to the physical membranes breakdown following dehydration or result from the presence of
350 active PME in dead plant materials (Castaldo et al., 1997; Galbally and Kirstine, 2002). Indeed,
351 most of the PME formed by the plant remains active under normal environmental conditions and
352 is capable of demethylation of about 65% of the pectin in the dead plant material (Galbally and
353 Kirstine, 2002).

354 The acetaldehyde production mechanism in senescent leaves is not yet well known, but fatty
355 acid peroxidation by reactive oxygen species (ROS) has been suggested by Jardine et al. (2009)
356 as a potential mechanism based on their ¹³C isotope analysis studies on stressed leaves of
357 deciduous trees. As increased oxidative stress and lipid peroxidation have been observed
358 (Prochazkova et al., 2001) during maize leaf senescence, the suggested mechanism might be
359 responsible for high production of acetaldehyde during the degeneration phase.

360 GLV emissions from senescent leaves have been reported previously in the literature. While
361 Holopainen et al. (2010) mentioned increases in GLV emissions from senescent leaves of *Betula*
362 *pendula* Roth before abscission, others observed hexenal and hexenol emissions from leaf drying
363 experiments (Fall et al., 1999; Fall et al., 2001; de Gouw et al., 2000; Karl et al., 2001a; Karl et
364 al., 2001c; Karl et al., 2001b; Warneke et al., 2002; Karl et al., 2005). During leaf senescence,
365 most of the cellular fatty acids are oxidised to supply energy for the senescence process (Lim et
366 al., 2007). Since fatty acid oxidation, catalysed by 13-lipoxygenase (13-LOX), is the initial step
367 in GLV production in leaves (Scala et al., 2013; Hatanaka et al., 1993), high GLV emissions can
368 be expected during leaf senescence. Furthermore, recent research revealed the key role of this
369 enzyme in the degradation of chloroplasts during leaf senescence (Springer et al., 2016),
370 potentially leading to strong increases in the emission of GLVs with increasing chlorosis of the
371 leaf.

372 **3.1.2 Persistent m/z 59 compounds and acetic acid emissions during the** 373 **terminal phase of leaf senescence**

374 Whereas the temporal evolution of acetic acid emissions was well-marked and no emissions
375 were observed prior to senescence, normalized m/z 59 compounds emissions from senescent
376 leaves showed a much smaller increase with respect to the constitutive emissions by the end of
377 the chlorosis period (Fig. 2 and Fig. S1b). Normalised acetic acid and m/z 59 compounds
378 emissions increased somewhat slower than those of methanol, acetaldehyde and GLVs, resulting
379 in a delay of their maximum emissions with respect to the latter three compounds by 1-2 days.

380 Furthermore, in strong contrast to methanol, acetaldehyde and GLVs, acetic acid and m/z 59
381 compounds emission rates remained close to their maximal value for a few days after the end of
382 the chlorosis period and then slowly decreased with time, but never reached zero during the
383 entire measurement period. Early morning acetic acid emissions were generally somewhat lower
384 than at the end of the day, presumably due to a decrease in temperature in the leaf enclosures
385 during the night. A similar diurnal behaviour was observed for the m/z 59 compounds emissions
386 from senescent leaves which, in contrast to the pre-senescence constitutive emissions, remained
387 well above the zero level in dark conditions. High day and night-time m/z 59 compounds
388 emissions (acetone) from dry plant parts and litter have previously been mentioned in the

389 literature (Schade and Goldstein, 2001; Warneke et al., 2002) and acetic acid emissions from dry
390 plant material have been reported as well (Kesselmeier et al., 1998; Warneke et al., 1999; Crespo
391 et al., 2013). The production of acetone and acetic acid during the drying process following the
392 chlorosis period might be explained by the occurrence of a non-enzymatic Maillard reaction
393 (Ikan et al., 1996). After the end of the chlorosis period a significant positive correlation ($R^2 =$
394 0.53 , $P < 0.01$) was observed between the acetic acid emissions and the relative humidity of the
395 air supplied to the enclosures. Similarly, Warneke et al. (1999) observed a large increase in
396 partially oxidised VOC (POVOC) emissions after wetting of dried biomatter. They explained
397 this by the transfer of some of the POVOC molecules, produced by the Maillard reaction and
398 remaining attached to the surface of the solid structure of the dry leaf material, to the aqueous
399 phase through replacement by highly polar water molecules. The dissolved POVOCs were
400 subsequently transferred from the aqueous to the gas phase until a gas/liquid equilibrium,
401 determined by the Henry's law constant, was reached. Although the dry leaves were not wetted
402 in our experiments, release of acetic acid molecules from the biomass surface layer may have
403 been influenced by the presence of a relative humidity dependent microlayer of water on the
404 surface layer or by interactions with polar water vapour molecules. Those interactions may have
405 promoted the release of the polar BVOC molecules from the surface layer. A short additional
406 experiment was carried out in which an inert surface, coated with a pure acetic acid microlayer,
407 was inserted in an empty enclosure demonstrated that such a positive correlation between acetic
408 acid emissions and relative humidity was not restricted to senescent biomass. Indeed, the acetic
409 acid emission rate from the coated surface, which was of similar magnitude to that of the
410 enclosed senescent maize leaves, was found to increase by a factor of 2.7 when increasing the
411 relative humidity of the purge air from 40 to 64 %. Moreover, increased emissions of acetic acid
412 and other polar compounds with increasing relative humidity have been reported previously from
413 other materials such as wood boards (Steckel et al., 2013) and indoor furniture (Fechter et al.,
414 2006; Schaeffer et al., 1996).

415 **3.1.3 Absolute BVOC emissions from senescent maize leaves and relative**
416 **contribution of the emitted compounds to the total BVOC emission from**
417 **senescent leaves**

418 Cumulative emissions of the quantified BVOCs for the different senescence phases are shown in
419 Table 3, along with the maximum instantaneous emissions and the maximum daily emissions
420 that have been observed over the entire senescence period. Additional statistical information on
421 instantaneous and daily emissions and on cumulative emissions from the investigated senescent
422 leaves is provided in Table S2 and in Fig. S2 in Supplement S1. The emission rates for some
423 compounds (e.g. methanol, m/z 59 compounds, and acetic acid) were still well above the
424 detection limit at the end of the measurement period. Consequently, the reported cumulative
425 emission values from the onset of senescence to the end of the measurement period (31 days) for
426 those compounds should be considered as lower limits for their total cumulative emissions
427 during leaf senescence.

428 Table 3: BVOC emissions from maize leaves at different leaf developmental stages. Maximum values for the instantaneous ($E_{inst,max}$)
 429 and for the daily emission rate ($E_{day,max}$), and cumulative emissions for the initiation, degeneration, and termination phases are shown
 430 for the different BVOCs emitted during senescence. Also tabulated are the daily emission rates (E_{day}) from young, semi-mature and
 431 mature maize leaves. Data for the senescent leaves are averages over 10 replicates and data for the young, semi-mature and mature
 432 leaves are averages over 5 replicates. The error on the emission rates corresponds to the standard deviation (1σ).

433

Compounds	Senescent leaves				$E_{day,max}$ ($\mu\text{g g}_{DW}^{-1}$ day^{-1})	Young	Semi-	Mature
	$E_{inst,max}$ (ng g_{DW}^{-1} s^{-1})	E_{acc} ($\mu\text{g g}_{DW}^{-1}$)				leaves	mature	leaves
		Initiation phase	Degeneration phase	Termination phase		E_{day} ($\mu\text{g g}_{DW}^{-1}$ day^{-1})	E_{day} ($\mu\text{g g}_{DW}^{-1}$ day^{-1})	E_{day} ($\mu\text{g g}_{DW}^{-1}$ day^{-1})
Methanol	0.9 ± 0.4	14 ± 11	150 ± 30	50 ± 20	53 ± 16	$137 \pm 17^*$	$4.1 \pm 0.7^*$	$4.8 \pm 1.0^*$
Acetaldehyde	0.38 ± 0.10	1 ± 2	44 ± 12	32 ± 19	15 ± 3	-	-	-
m/z 59 compounds	0.05 ± 0.02	5 ± 5	12 ± 7	31 ± 20	2.3 ± 1.2	2.7 ± 1.1	0.4 ± 0.2	0.9 ± 0.3
Acetic acid	0.17 ± 0.05	6 ± 7	40 ± 20	160 ± 60	12 ± 4	<DL**	<DL	<DL
m/z 69 compounds	0.08 ± 0.02	3 ± 3	13 ± 8	22 ± 16	3.4 ± 1.3	<DL	<DL	<DL
m/z 73 compounds	0.03 ± 0.02	3 ± 4	5 ± 7	11 ± 15	1.2 ± 1.3	1.7 ± 1.0	0.4 ± 0.2	0.30 ± 0.08
Hexenals	0.21 ± 0.14	6 ± 7	30 ± 20	20 ± 20	10 ± 8	-	-	-
Monoterpenes	<DL	<DL	<DL	<DL	<DL	10 ± 6	0.6 ± 0.2	0.4 ± 0.2
m/z 83 compounds	0.07 ± 0.02	1.4 ± 1.7	9 ± 6	11 ± 9	2.6 ± 1.5	-	-	-

434 *: Mozaffar et al. (2017); ** Detection Limit

435

436

437 Methanol and acetic acid were clearly the compounds with the highest emissions (expressed in
438 mass of compound per leaf dry weight), accounting both for around 30% of the total cumulative
439 emission from the onset of senescence to the end of the investigations (Fig. 3a). They were
440 followed by acetaldehyde (11%), hexenals (9%), m/z 59 compounds (7%), and m/z 69
441 compounds (6%). The m/z 73 and m/z 83 compounds also made up a small fraction of the total
442 BVOC emissions from the senescent leaves, but their individual contributions were not higher
443 than 3%.

444

445 Figure 3

446

447 Fig. 3 clearly shows that the relative contribution of the emitted compounds varied among the
448 time periods considered. Differences in relative BVOC composition reflect the differences in
449 temporal emission dynamics among BVOCs that were discussed in Section 3.1.1 and 3.1.2.
450 Indeed, as methanol, acetaldehyde, and hexenals were predominantly emitted during the
451 degeneration phase (see Table 3), it is clear that those compounds together made up a large part
452 (75%) of the BVOC emissions for that period (Fig. 3b). Because of the strong persistence of m/z
453 59 compounds and acetic acid emissions after the chlorosis period, the relative contribution of
454 those compounds increased with time after that period. During the termination phase (Fig. 3c),
455 the m/z 59 compounds and acetic acid together made up half (55%) of the BVOC emissions, and
456 acetic acid was predominantly emitted during that period (46%).

457 While some studies (Karl et al., 2001b; Karl et al., 2005; Eller et al., 2011; Crespo et al.,
458 2013) have reported cumulative BVOC emissions from biomass drying and simulated
459 drying/senescence experiments, information about cumulative BVOC emissions from senescent
460 leaves that are still attached to the plant is, to our knowledge, not available in the literature.
461 Moreover, the abovementioned studies are the only available ones which report cumulative
462 emissions in conditions which somehow approximate senescence in real natural conditions.
463 Therefore we will compare the cumulative BVOC emissions from senescent maize leaves over
464 the entire measurement period against the results from those studies (Table 4), which were all
465 performed on different species of the Poaceae family. Overall, BVOC emissions from senescent

466 maize leaves are among the highest for the drying/senescent plant species. The cumulative
467 methanol emission from the senescent maize leaves is very similar to that of hay, reported by
468 Karl et al. (2001b). By contrast, cumulative methanol emissions from *Sorghum sudanense*
469 (sorghum) (Karl et al., 2005), *Oryza sativa* (rice) (Karl et al., 2005), *Panicum virgatum*
470 (switchgrass) (Eller et al., 2011), and *Phyllostachys nigra* (black bamboo) (Crespo et al., 2013)
471 are several orders of magnitude lower than those for maize and hay. Cumulative emissions of
472 acetaldehyde, m/z 59, and m/z 69 compounds from maize are of the same order of magnitude as
473 those from drying hay and black bamboo, and considerably higher than those from sorghum, rice,
474 and switchgrass. Total acetic acid emission from maize is several-fold higher than that of black
475 bamboo, which is the only other species of the Poaceae family for which cumulative acetic acid
476 emissions have been reported (Warneke et al. (1999) also mentioned acetic acid emission from
477 grass, but didn't report cumulative emissions). Maize, hay, rice, and black bamboo all have
478 higher cumulative GLV emissions than sorghum and switchgrass.

479 This comparison of cumulative BVOC emissions from the different plant species should,
480 however, be taken with caution because of differences in the way they have been determined
481 (from leaves still attached to the plant as in this study or from cutting of biomass, followed by
482 drying in the field or in an oven as in the other studies), in environmental conditions during the
483 experiments, and in the length of the accumulation period (see Table 4).

484 Table 4: Comparison of the cumulative BVOC emissions during senescence of maize leaves and drying of hay, *Sorghum sudanense*
 485 (sorghum), *Oryza sativa* (rice), *Panicum virgatum* (switchgrass) and *Phyllostachys nigra* (black bamboo). Also mentioned in the
 486 Table are the accumulation period and the temperature at which the experiments were performed.

	Maize	Hay	Sorghum	Rice	Switchgrass	Black bamboo
	Current study	Karl et al. (2001b)		Karl et al. (2005)	Eller et al. (2011)	Crespo et al. (2013)
	± total (~31 days)	total	total (7-9 hours)	total (7-9 hours)	partial (1 day)	total (3 days)
	25 °C	25-35 °C	30 °C	30 °C	30 °C	80 °C
Compound	(µg/g _{DW})	(µg/g _{DW})	(µg/g _{DW})	(µg/g _{DW})	(µg/g _{DW})	(µg/g _{DW})
Methanol	210 ± 30	160	2 ± 0.9	3.1 ± 0.8	9.07 ± 2.35	26.1
Acetaldehyde	80 ± 30	20-80	6.6 ± 2	8.4 ± 0.7	13.99 ± 5.44	73.2
m/z 59 compounds	50 ± 30	20-40	0.4 ± 0.3	1.7 ± 1.1	2.21 ± 0.69	13.3
Acetic acid	200 ± 80	-	-	-	-	28.6
m/z 69 compounds	40 ± 20	15 ^a	0.32 ± 0.13 ^b	3.2 ± 1.7 ^b	1.15 ± 0.53 ^c	58 ^c
m/z 73 compounds	20 ± 30	11-80	-	-	-	-
Hexenals	60 ± 50	100-240	0.53 ± 0.17	100 ± 49	1.49 ± 1.42	231 ^d
m/z 83 compounds	21 ± 15	30-60 ^e	0.36 ± 0.17	34 ± 13	1.06 ± 1.49	

487

488 ^a pentenol and 2-methyl-butanal; ^b pentenol and n-pentanal; ^c pentenol and isoprene; ^d based on the sum of ion intensities at m/z 57,
 489 81, 99, 83, 101, and 103; ^e hexenols and hexanal

490 **3.2 Contribution of leaf senescence to the total BVOC emissions from a**
491 **maize leaf/plant**

492 In order to investigate the relative importance of BVOC emission rates from senescent maize
493 leaves, these emission rates were compared with those of young, semi-mature, and mature maize
494 leaves measured under the same environmental conditions (Table 3). Strong variations in daily
495 emission rates were observed among the different leaf developmental stages for those
496 compounds that could be quantified for the three earliest stages. Methanol was clearly the
497 compound with the highest emission rates for all stages. The highest daily emission rate during
498 the senescence period was about 40% of the average daily emission rate for the young leaves and
499 about a factor of 10 higher than that of semi-mature and mature leaves. The high emission rates
500 for senescent and young leaves with respect to those of the other developmental stages are due to
501 enhanced production of methanol by PME-catalysed demethylation of pectin during cell wall
502 remodelling. Whereas complex diurnal methanol emission profiles were found for the young
503 leaves (Mozaffar et al., 2017), no clear correlation between the methanol emission rate and PPFD
504 could be observed for the senescent leaves either. Maximum daily emission rates for m/z 59 and
505 m/z 73 compounds during senescence were of similar magnitude to the corresponding average
506 daily emission rates for young leaves and about 3-6 times higher than the average daily emission
507 rates for the semi-mature and mature leaves. Daily monoterpene emission rates for young maize
508 leaves were more than ten-fold higher than for semi-mature and mature leaves and emission rates
509 from senescent leaves were below the detection limit. Higher acetone and monoterpene
510 emissions from young leaves have previously been reported in the literature and higher defence
511 requirements for the young leaves (Bracho-Nunez et al, 2011) and lower metabolic activity in
512 mature leaves resulting in low *de novo* synthesis rates (Aalto et al., 2014) have been proposed as
513 potential reasons.

514 Although the data listed in Table 3 allow a comparison between the maximum daily BVOC
515 emission rates during senescence and the average daily BVOC emission rates measured during
516 relatively short periods for the young, semi-mature, and mature leaves, additional information is
517 required to assess the contributions of the different leaf developmental stages to the total BVOC
518 emissions for an individual leaf in the course of its lifetime. By taking into account the

519 experimental daily methanol emission rate data from the young and mature leaves (Mozaffar et
520 al., 2017) and the cumulative methanol emissions from senescent leaves (Table 3), and by using
521 information about the relative leaf area growth rate and the duration of the young-growing
522 (encompassing young and semi-mature) and mature leaf developmental stages, it was possible to
523 estimate the cumulative methanol emissions from a 7th leaf of a maize plant during the young-
524 growing, mature and senescent leaf developmental stages. The leaf number again refers to the
525 order of appearance of the leaf on the maize plant. They were $360 \pm 30 \mu\text{g}$, $80 \pm 30 \mu\text{g}$ and $150 \pm$
526 $40 \mu\text{g}$, respectively, and the resulting total amount of methanol emitted by a 7th leaf in the course
527 of its lifetime was therefore $590 \pm 60 \mu\text{g}$. The young-growing, mature and senescent stages
528 consequently contributed 61 ± 3 , 13 ± 5 and $26 \pm 6\%$ to the total methanol emission from a 7th
529 leaf. Details about the estimation procedure are provided in Supplement S1.

530 Based on the emission rates obtained for a 7th leaf, the total amount of methanol emitted by all
531 leaves of a representative maize plant, grown in the environmental chamber at 25 °C and exposed
532 to the diurnal light pattern described in Section 2.1, has been estimated for the different leaf
533 developmental stages. Values of 4.8 ± 0.4 , 0.8 ± 0.5 and $1.9 \pm 0.6 \text{ mg}$ were obtained for the
534 cumulative methanol emission from all young-growing, mature, and senescent leaves over the
535 entire lifetime of the plant, respectively. The total amount of methanol emitted by the leaves of a
536 whole maize plant from shoot emergence to full senescence was therefore equal to $7.5 \pm 0.9 \text{ mg}$
537 and the contributions of the different stages to the total methanol emission were 64 ± 3 , 10 ± 8
538 and $26 \pm 7\%$, respectively. The estimation of the total cumulative methanol emission from a
539 whole maize plant over the course of its lifetime should however be considered with caution. As
540 already mentioned it was assumed that all leaves were exposed to the same PPFD and the effect
541 of shading by leaves with a higher leaf number was not taken into account. Moreover, a large
542 variability in the duration of the mature stage was noticed among leaves with different leaf
543 numbers, which may be related to the demands of the plant for nutrient relocation. The total leaf
544 emissions of methanol from a maize plant presented in this work could be used as a first estimate
545 for upscaling to field scale in regions characterised by environmental conditions close to those of
546 the growth chamber. However, a general upscaling from the leaf/plant level to ecosystem level
547 would definitely benefit from methanol emission measurements at different temperatures at all
548 leaf developmental stages. Moreover, uptake of atmospheric methanol by the ecosystem may

549 also be important in the field (Laffineur et al., 2012; Wohlfahrt et al., 2015) and should be taken
550 into account.

551 For the other BVOCs emitted by maize leaves, assumptions about the temporal evolution of
552 the emission rates at the different leaf developmental stages are less straightforward than for
553 methanol. Indeed, whereas methanol emission from young leaves has been associated with leaf
554 growth (Hüve et al. 2007), this is not the case for other BVOCs and therefore a simple
555 relationship between the daily emission rate of those compounds and the daily relative leaf area
556 growth rate (eq. S3 in Supplement S1) cannot be put forward. Consequently, an estimation of the
557 relative contribution of the different leaf developmental stages to the total emission from a 7th
558 maize leaf for other BVOCs than methanol- from the emission rate data obtained in this work
559 would have been prone to very large errors and has therefore not been accomplished. The same
560 applies to extrapolation to an entire maize plant.

561 **4 Conclusions**

562 Maize is one of the most cultivated crop species worldwide, but only 5 studies on BVOC
563 exchanges from maize are available in the literature and none of them deal with emissions from
564 senescent leaves. In contrast to most studies on BVOC emissions from artificially senescing
565 leaves (cutting and drying), the senescent maize leaves in our experiments were still attached to
566 the stems.

567 The main emitted compounds, ranked according to their cumulative emissions over the
568 senescence period, were found to be methanol, acetic acid, acetaldehyde, hexenals, m/z 59
569 compounds, m/z 69 compounds, m/z 73 compounds, and m/z 83 compounds. Important
570 differences were observed in the temporal emission profiles of these compounds. Whereas
571 methanol, acetaldehyde, and GLVs (hexenals and m/z 83 compounds) were emitted mainly
572 during the degeneration phase of leaf senescence, m/z 59 compounds and acetic acid emission
573 rates increased at the end of that phase and their emissions remained high during the termination
574 phase, even when the leaves were already completely dry. Beside m/z 59 compounds and acetic
575 acid, the rest of the abovementioned compounds were also emitted in small but significant
576 amounts from dry leaves during the termination phase. Therefore, not only the yellow senescent
577 maize leaves but also the dry brown leaves, which remain attached to the plant, are an important

578 source of BVOCs. By comparing cumulative BVOC emissions from senescent maize leaves with
579 those of artificially senescing species of the Poaceae family, it was found that maize leaves were
580 clearly among the strongest emitting species during that leaf developmental stage. Nevertheless,
581 an improved comparison of cumulative BVOC emissions among senescent species of the
582 Poaceae family might benefit from additional measurements on naturally senescing plants instead
583 of results obtained from cutting and drying experiments.

584 Whereas senescent leaves showed a large diversity of emitted compounds, BVOC emission
585 rates for young, semi-mature, and mature leaves were limited to methanol, m/z 59 and m/z 73
586 compounds, and monoterpenes. Methanol was clearly the highest emitted compound for all
587 stages and showed a strong variation in intensity and diurnal emission patterns among the
588 different leaf developmental stages. The contributions from the young-growing, mature, and
589 senescent stages to the total methanol emission from a typical leaf of a maize plant were
590 estimated to be 61, 13, and 26%, respectively.

591 Although our growth chamber study provided new information on the contribution of
592 senescence and other leaf developmental stages to the BVOC emissions from maize, additional
593 studies, preferably in field conditions and at a wide variety of meteorological conditions, are
594 definitely required to better constrain BVOC emissions for this important C4 crop species for use
595 in regional and global atmospheric chemistry and climate models.

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602

603 **References**

604

605 Atkinson, R.: Atmospheric chemistry of VOCs and NO_x, *Atmospheric Environment*, 34(12-14),
606 2063–2101, doi: 10.1016/S1352-2310(99)00460-4, 2000.

607 Aalto, J., Kolari, P., Hari, P., Kerminen, V.-M., Schiestl-Aalto, P., Aaltonen, H., Levula, J.,
608 Siivola, E., Kulmala, M. and Bäck, J.: New foliage growth is a significant, unaccounted
609 source for volatiles in boreal evergreen forests, *Biogeosciences*, 11, 1331-1344,
610 doi:10.5194/bg-11-1331-2014, 2014.

611 Bachy, A., Aubinet, M., Schoon, N., Amelynck, C., Bodson, B., Moureaux, C. and Heinesch, B.:
612 Are BVOC exchanges in agricultural ecosystems overestimated? Insights from fluxes
613 measured in a maize field over a whole growing season, *Atmospheric Chemistry and*
614 *Physics*, 16, 5343-5356, doi: 10.5194/acp-16-5434-2016, 2016.

615 Bracho-Nunez, A., Welter, S., Staudt, M. and Kesselmeier, J.: Plant-specific volatile organic
616 compound emission rates from young and mature leaves of Mediterranean vegetation,
617 *Journal of Geophysical Research*, 116(D16), doi: 10.1029/2010JD015521, 2011.

618 Castaldo, D., Laratta, B., Loiudice, R., Giovane, A., Quagliuolo, L., and Servillo, L.: Presence of
619 residual pectin methylesterase activity in thermally stabilized industrial fruit preparations,
620 *LWT - Food Science and Technology*, 30, 479–484, doi: 10.1006/fstl.1996.0211, 1997.

621 Crespo, E., Graus, M., Gilman, J. B., Lerner, B. M., Fall, R., Harren, F. J. M. and Warneke, C.:
622 Volatile organic compound emissions from elephant grass and bamboo cultivars used as
623 potential bioethanol crop, *Atmospheric Environment*, 65, 61-68, doi:
624 10.1016/j.atmosenv.2012.10.009, 2013.

625 Das, M., Kang, D., Aneja, V.P., Lonneman, W., Cook, D.R. and Wesely, M.L.: Measurements of
626 hydrocarbon air–surface exchange rates over maize, *Atmospheric Environment* 37(16),
627 2269-2277, doi: 10.1016/S1352-2310(03)00076-1, 2003.

628 de Gouw, J. D., Howard, C. J., Custer, T. G. and Fall, R.: Emissions of volatile organic
629 compounds from cut grass and clover are enhanced during the drying process, *Geophysical*
630 *Research Letters*, 26(7), 811-814, doi: 10.1029/1999GL900076, 1999.

631 de Gouw, J. D., Howard, C. J., Custer, T. G., Baker, B. M. and Fall, R.: Proton-Transfer
632 Chemical-Ionization Mass Spectrometry allows real-time analysis of volatile organic
633 compounds released from cutting and drying of crops, *Environmental Science &*
634 *Technology*, 34(12), 2640-2648, doi: 10.1021/es991219k, 2000.

635 Eller, A.S.D., Sekimoto, K., Gilman, J.B., Kuster, W.C., de Gouw, J.A., Monson, R.K., Graus,
636 M., Crespo, E., Warneke, C. and Fall, R.: Volatile organic compound emissions from
637 switchgrass cultivars used as biofuel crops, *Atmospheric Environment* 45 (19), 3333-3337,
638 doi: 10.1016/j.atmosenv.2011.03.042, 2011.

639 Ellis, A. M. and Mayhew, C. A.: Proton transfer reaction mass spectrometry: principles and
640 applications, Wiley, Chichester, 2013.

641 Fechter, J.-O., Englund, F. and Lundin, A.: Association between temperature, relative humidity
642 and concentration of volatile organic compounds from wooden furniture in a model room,
643 *Wood Material Science and Engineering* 1 (2), 69-75, doi: 10.1080/17480270600900551,
644 2006.

645 Fall, R.: Abundant oxygenates in the atmosphere: a biochemical perspective, *Chemical Reviews*,
646 103(12), 4941-4952, doi: 10.1021/cr0206521, 2003.

647 Fall, R. and Benson, A.A.: Leaf methanol - the simplest natural product from plants, *Trends in*
648 *Plant Science*, 1, 296-301, doi: 10.1016/S1360-1385(96)88175-0, 1996.

649 Fall, R., Karl, T., Hansel, A., Jordan, A. and Lindinger, W.: Volatile organic compounds emitted
650 after leaf wounding: On-line analysis by proton-transfer-reaction mass spectrometry,
651 *Journal of Geophysical Research: Atmospheres*, 104(D13), 15963–15974, doi:
652 10.1029/1999JD900144, 1999.

653 Fall, R., Karl, T., Jordan, A. and Lindinger, W.: Biogenic C5 VOCs: release from leaves after
654 freeze–thaw wounding and occurrence in air at a high mountain observatory, *Atmospheric*
655 *Environment*, 35(22), 3905-3916, doi: 10.1016/S1352-2310(01)00141-8, 2001.

- 656 Folkers, A., Hüve, K., Ammann, C., Dindorf, T., Kesselmeier, J., Kleist, E., Kuhn, U., Uerlings,
657 R. and Wildt, J.: Methanol emissions from deciduous tree species: dependence on
658 temperature and light intensity, *Plant Biology*, 10, 65-75, doi: 10.1111/j.1438-
659 8677.2007.00012.x, 2008.
- 660 Galbally, I. E. and Kirstine, W.: The production of methanol by flowering plants and the global
661 cycle of methanol, *Journal of Atmospheric Chemistry*, 43, 195-229, doi:
662 10.1023/A:1020684815474, 2002.
- 663 Gan, S. and Amasino, R. M.: Making sense of senescence (molecular genetic regulation and
664 manipulation of leaf senescence), *Plant Physiology*, 113(2), 313-319, 1997.
- 665 Graus, M., Eller, A., Fall, R., Yuan, B., Qian, Y., Westra, P., de Gouw, J. and Warneke, C.:
666 Biosphere-atmosphere exchange of volatile organic compounds over C4 biofuel crops,
667 *Atmospheric Environment*, 66, 161-168, doi: 10.1016/j.atmosenv.2011.12.042, 2013.
- 668 Guenther, A. B., Jiang, X., Heald, C., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K. and
669 Wang, X., The Model of Emissions of Gases and Aerosols from Nature version 2.1
670 (MEGAN 2.1): an extended and updated framework for modelling biogenic emissions,
671 *Geoscientific Model Development*, 5, 1471-1492, doi: 10.5194/gmd-5-1471-2012, 2012.
- 672 Harley, P., Greenberg, J., Niinemets, Ü. and Guenther, A.: Environmental controls over methanol
673 emission from leaves, *Biogeosciences*, 4, 1083-1099, doi: 10.5194/bg-4-1083-2007, 2007.
- 674 Hatanaka, A.: The biogeneration of green odour by green leaves, *Phytochemistry*, 34, 1201-1218,
675 doi: 10.1016/0031-9422(91)80003-J, 1993.
- 676 Holopainen, J. K., Heijari, J., Oksanen, E. and Alessio, G. A.: Leaf volatile emissions of *betula*
677 *pendula* during autumn coloration and leaf fall. *Journal of Chemical Ecology*, 36(10),
678 1068-1075, doi: 10.1007/s10886-010-9857-4, 2010.
- 679 Holopainen, J. K., and Gershenson, J., Multiple stress factors and the emission of plant VOCs,
680 *Trends in Plant Science*, 15, 176-184, doi: 10.1016/j.tplants.2010.01.006, 2010.
- 681 Hüve, K., Christ, M. M., Kleist, E., Uerlings, R., Niinemets, U., Walter, A. and Wildt, J.:
682 Simultaneous growth and emission measurements demonstrate an interactive control of

683 methanol release by leaf expansion and stomata, *Journal of Experimental Botany*, 58,
684 1783-1793, doi: 10.1093/jxb/erm038, 2007.

685 Ikan, R., Rubinszaln, Y., Nissenbaum, A. and Kaplan, I. R.: The maillard reaction:
686 consequences for the chemical and life sciences, edited by Ikan, R., 1-25, John Wiley, New
687 York, 1996.

688 Inomata, S. and Tanimoto, H.: A quantitative examination of the detection sensitivities of proton-
689 transfer reaction mass spectrometry for gaseous 2-propanol and acetic acid, *Bulletin of the*
690 *Chemical Society of Japan*, 83, 900–904, doi: 10.1246/bcsj.20100043, 2010.

691 Jardine, K., Karl, T., Lerdau, M., Harley, P., Guenther, A., and Mak, J.E.: Carbon isotope
692 analysis of acetaldehyde emitted from leaves following mechanical stress and anoxia, *Plant*
693 *Biology*, 11(4), 591-597, doi: 10.1111/j.1438-8677.2008.00155.x, 2009.

694 Karl, T., Fall, R., Jordan, A. and Lindinger, W.: On-line analysis of reactive VOCs from urban
695 lawn mowing, *Environmental Science & Technology*, 35(14), 2926-2931, doi:
696 10.1021/es010637y, 2001a.

697 Karl, T., Guenther, A., Jordan, A., Fall, R. and Lindinger, W.: Eddy covariance measurement of
698 biogenic oxygenated VOC emissions from hay harvesting, *Atmospheric Environment*, 35,
699 491-495, doi: 10.1016/S1352-2310(00)00405-2, 2001b.

700 Karl, T., Guenther, A., Lindinger, C., Jordan, A., Fall, R. and Lindinger, W.: Eddy covariance
701 measurements of oxygenated volatile organic compound fluxes from crop harvesting using
702 a redesigned proton-transfer-reaction mass spectrometer, *Journal of Geophysical Research:*
703 *Atmospheres*, 106(D20), 24157-24167, 2001c.

704 Karl, T., Harren, F., Warneke, C., Gouw, J. D., Grayless, C. and Fall, R.: Senescing grass crops
705 as regional sources of reactive volatile organic compounds. *Journal of Geophysical*
706 *Research*, 110(D15), doi: 10.1029/2005JD005777, 2005.

707 Keskitalo, J., Bergquist, G., Gardeström, P. and Jansson, S.: A cellular timetable of autumn
708 senescence, *Plant Physiology*, 139(4), 1635-1648, doi: 10.1104/pp.105.066845, 2005.

709 Kesselmeier, J., Bode, K., Gerlach, C., and Hork, E.-M.: Exchange of atmospheric formic and
710 acetic acids with trees and crop plants under controlled chamber and purified air

- 711 conditions, *Atmospheric Environment*, 32(10), 1765-1775, doi: 10.1016/S1352-
712 2310(97)00465-2, 1998.
- 713 Kreuzwieser, J., Kuhnemann, F., Martis, A., Rennenberg, H. and Urban, W.: Diurnal pattern of
714 acetaldehyde emission by flooded poplar trees, *Physiologia Plantarum*, 108(1), 79-86, doi:
715 10.1034/j.1399-3054.2000.108001079.x, 2000.
- 716 Kuhn, U., Rottenberger, S., Biesenthal, T., Wolf, A., Schebeske, G., Ciccioli, P., Brancaleoni, E.,
717 Frattoni, M., Tavares, T. M. and Kesselmeier, J.: Isoprene and monoterpene emissions of
718 Amazonian tree species during the wet season: direct and indirect investigations on
719 controlling environmental functions, *Journal of Geophysical Research: Atmosphere*, 107
720 (D20), LBA 38-1–LBA 38-13, doi: 10.1029/2001JD000978, 2002.
- 721 Laffineur, Q., Aubinet, M., Schoon, N., Amelynck, C., Müller, J.-F., Dewulf, J., Van
722 Langenhove, H., Steppe, K., and Heinesch, B.: Abiotic and biotic control of methanol
723 exchanges in a temperate mixed forest, *Atmospheric Chemistry and Physics*, 12, 577-590,
724 doi: 10.5194/acp-12-577-2012, 2012.
- 725 Laothawornkitkul, J., Taylor, J. E., Paul, N.D., Hewitt, C.N.: Biogenic volatile organic
726 compounds in the Earth system, *New Phytologist*, 183, 27–51, 2009.
- 727 Lim, P. O., Kim, H. J. and Nam, H. G.: Leaf senescence, *Annual Review of Plant Biology*, 58,
728 115-136, doi: 10.1146/annurev.arplant.57.032905.105316, 2007.
- 729 Lindinger, W., Hansel, A., Jordan, A.: On-line monitoring of volatile organic compounds at pptv
730 levels by means of proton-transfer-reaction mass-spectrometry (PTR-MS) – medical
731 applications, food control and environmental research, *International Journal of Mass
732 Spectrometry and Ion Processes*, 173(3), 191-241, 1998.
- 733 MacDonald, R.C. and Fall, R.: Detection of substantial emissions of methanol from plants to the
734 atmosphere, *Atmospheric Environment, Part A General Topics*, 27, 1709-1713, doi:
735 10.1016/0960-1686(93)90233-O, 1993.
- 736 Mozaffar, A., Schoon, N., Digrao, A., Bachy, A., Delaplace P., du Jardin, P., Fauconnier, M.-L.,
737 Aubinet, M., Heinesch, B. and Amelynck, C.: Methanol emissions from maize: ontogenetic
738 dependence to varying light conditions and guttation as an additional factor constraining

739 the flux, *Atmospheric Environment*, 43, 405-417, doi: 10.1016/j.atmosenv.2016.12.041,
740 2017.

741 Nemecek-Marshall, M., MacDonald, R. C., Franzen, J. J., Wojciechowski, C. L. and Fall, R.:
742 Enzymatic detection of gas-phase methanol and relation of methanol fluxes to stomatal
743 conductance and leaf development, *Plant Physiology*, 108, 1359–1368, 1995.

744 Ozuna, R., Yera, R., Ortega, K., and Tallman, G.: Analysis of Guard Cell Viability and Action in
745 Senescing Leaves of *Nicotiana glauca* (Graham), *Tree Tobacco*, *Plant Physiology*, 79, 7-
746 10, 1985.

747 Pacifico, F., Harrison, S. P., Jones, C. D., and Sitch, S.: Isoprene emissions and climate,
748 *Atmospheric Environment*, 43, 6121-6135, doi: 10.1016/j.atmosenv.2009.09.002, 2009.

749 Pang, X.: Biogenic volatile organic compound analyses by PTR-TOF-MS: Calibration, humidity
750 effect and reduced electric field dependency, *Journal of Environmental Sciences*, 32, 196-
751 206, doi: 10.1016/j.jes.2015.01.013, 2015.

752 Pierik, R., Ballaré, C. L., and Dicke, M.: Ecology of plant volatiles: taking a plant community
753 perspective, *Plant, Cell and Environment*, 37, 1845-1853, doi: 10.1111/pce.12330, 2014.

754 Prochazkova, D., Sairam, R.K., Srivastava, G.C., and Singh, D.V.: Oxidative stress and
755 antioxidant activity as the basis of senescence in maize leaves, *Plant Science*, 161(4), 765-
756 771, doi: 10.1016/S0168-9452(01)00462-9, 2001.

757 Rottenberger, S., Kuhn, U., Wolf, A., Schebeske, G., Oliva, S. T., Tavares, T. M. and
758 Kesselmeier, J.: Exchange of short-chain aldehydes between Amazonian vegetation and the
759 atmosphere. *Ecological Applications*, 14, 247-262, doi: 10.1890/01-6027, 2004.

760 Scala, A., Allmann, S., Mirabella, R., Haring, M. and Schuurink, R.: Green Leaf Volatiles: A
761 plant's multifunctional weapon against herbivores and pathogens, *International Journal of*
762 *Molecular Sciences*, 14(9), 17781-17811, doi: 10.3390/ijms140917781, 2013.

763 Schade, G. W. and Goldstein, A. H.: Fluxes of oxygenated volatile organic compounds from a
764 ponderosa pine plantation, *Journal of Geophysical Research*, 106, 3111-3123, doi:
765 10.1029/2000JD900592, 2001.

- 766 Schaeffer, V.H., Bhoosan, B., Chen, S., Sonenthal, J.S. and Hodgson, A.J.: Characterisation of
767 volatile organic chemical emissions from carpet cushions, *Journal of the Air and Waste*
768 *Management Association*, 46(9), 813-820, doi: 10.1080/10473289.1996.10467516, 1996.
- 769 Schwarz, K., Filipiak, W., and Amann, A.: Determining concentration patterns of volatile
770 compounds in exhaled breath by PTR-MS, *Journal of Breath Research*, 3, 027002, doi:
771 10.1088/1752-7155/3/2/027002, 2009.
- 772 Springer, A., Kang, C., Rustgi, S., von Wettstein, D., Reinbothe, C., Pollmann, S. and
773 Reinbothe, S.: Programmed chloroplast destruction during leaf senescence involves 13-
774 lipoxygenase (13-LOX), *Proceedings of the National Academy of Sciences*, vol. 113(12),
775 3383-3388, 2016.
- 776 Steckel, V., Knöpfle A. and Ohlmeyer, M.: Effects of climatic test parameters on acetic acid
777 emission from beech, *Holzforschung*, 67, 47-51, doi: 10.1515/hf-2011-0237, 2013.
- 778 Su, T.: Parametrization of kinetic energy dependences of ion-polar molecule collision rate
779 constants by trajectory calculation. *Journal of Chemical Physics*, 100, 4703, doi:
780 10.1063/1.466255, 1994.
- 781 Takahashi, Y., Tsubaki, S., Sakamoto, M., Watanabe, S., and Azuma, J.: Growth-dependent
782 chemical and mechanical properties of cuticular membranes from leaves of *Sonneratia*
783 *alba*, *Plant, Cell and Environment*, 35, 1201-1210, doi: 10.1111/j.1365-3040.2012.02482.x,
784 2012.
- 785 Taiz, L., Zeiger, E., Moller, I.M., and Murphy, A.: *Plant physiology and development*, 6th
786 edition, Sinauer Associates, Sunderland, CT, 2015.
- 787 Tani, A., Hayward, S., and Hewitt, C.N.: Measurement of monoterpenes and related compounds
788 by proton transfer reaction-mass spectrometry, *International Journal of Mass Spectrometry*,
789 223/22, 561-578, doi: 10.1016/S1387-3806(02)00880-1, 2003.
- 790 Warneke, C., Karl, T., Judmaier, H., Hansel, A., Jordan, A., Lindinger, W. and Crutzen, P. J.:
791 Acetone+propanal, methanol, and other partially oxidized volatile organic emissions from
792 dead plant matter by abiological processes: Significance for atmospheric HO_x chemistry,
793 *Global Biogeochemical Cycles*, 13(1), 9-17, 1999.

794 Warneke, C., Luxembourg, S. L., de Gouw, J. A., Rinne, H. J. I., Guenther, A. B. and Fall, R.:
795 Disjunct eddy covariance measurements of oxygenated volatile organic compounds fluxes
796 from an alfalfa field before and after cutting. *Journal of Geophysical Research*, 107(D8),
797 ACH 6-1–ACH 6-10, doi: 10.1029/2001JD000594, 2002.

798 Wohlfahrt, G., Amelynck, C., Ammann, C., Arneth, A., Bamberger, I., Goldstein, A.H., Gu, L.,
799 Guenther, A., Hansel, A., Heinesch, B., Holst, T., Hörtnagl, L., Karl, T., Laffineur, Q.,
800 Neftel, A., McKinney, K., Munger, J.W., Pallardy, S.G., Schade, G.W., Seco, R. and
801 Schoon, N.: An ecosystem-scale perspective of the net land methanol flux: synthesis of
802 micrometeorological flux measurements, *Atmospheric Chemistry and Physics*, 15, 7413-
803 7427, doi: 10.5194/acp-15-7413-2015, 2015.

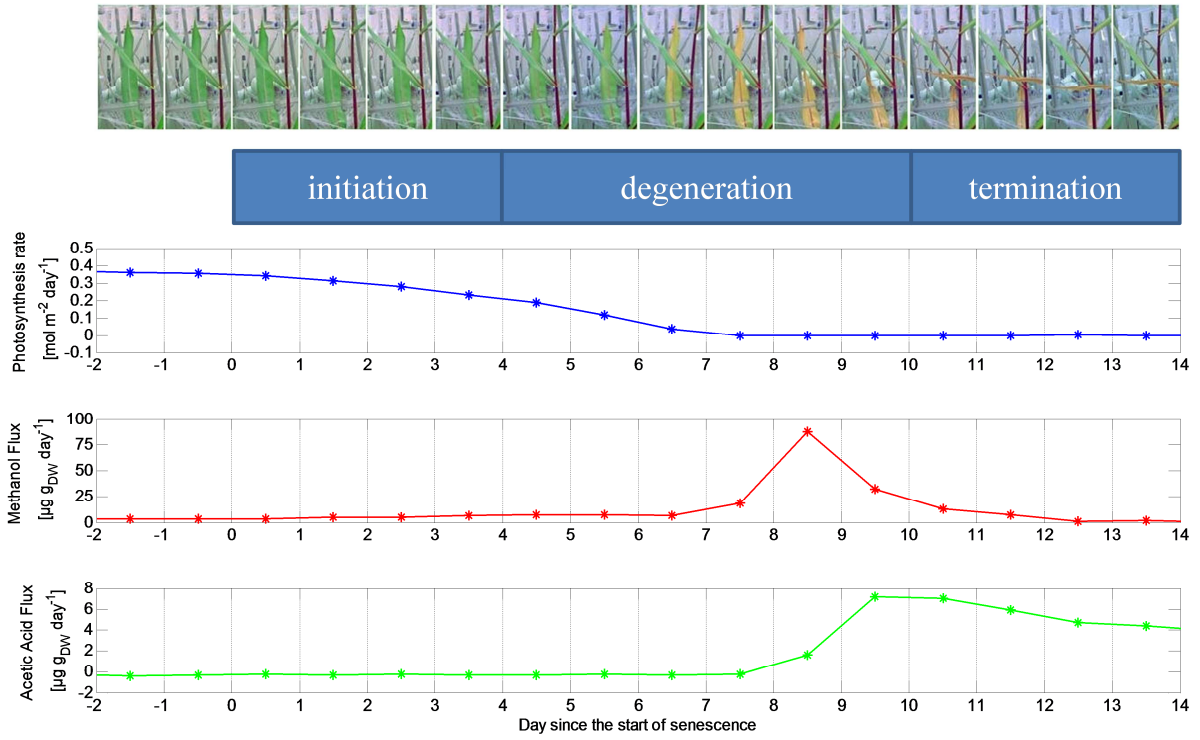
804 Woo, H. R., Kim, H. J., Nam, H. G. and Lim, P. O.: Plant leaf senescence and death - regulation
805 by multiple layers of control and implications for aging in general, *Journal of Cell Science*,
806 126(21), 4823-4833, doi: 10.1242/jcs.109116, 2013.

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Figure 1

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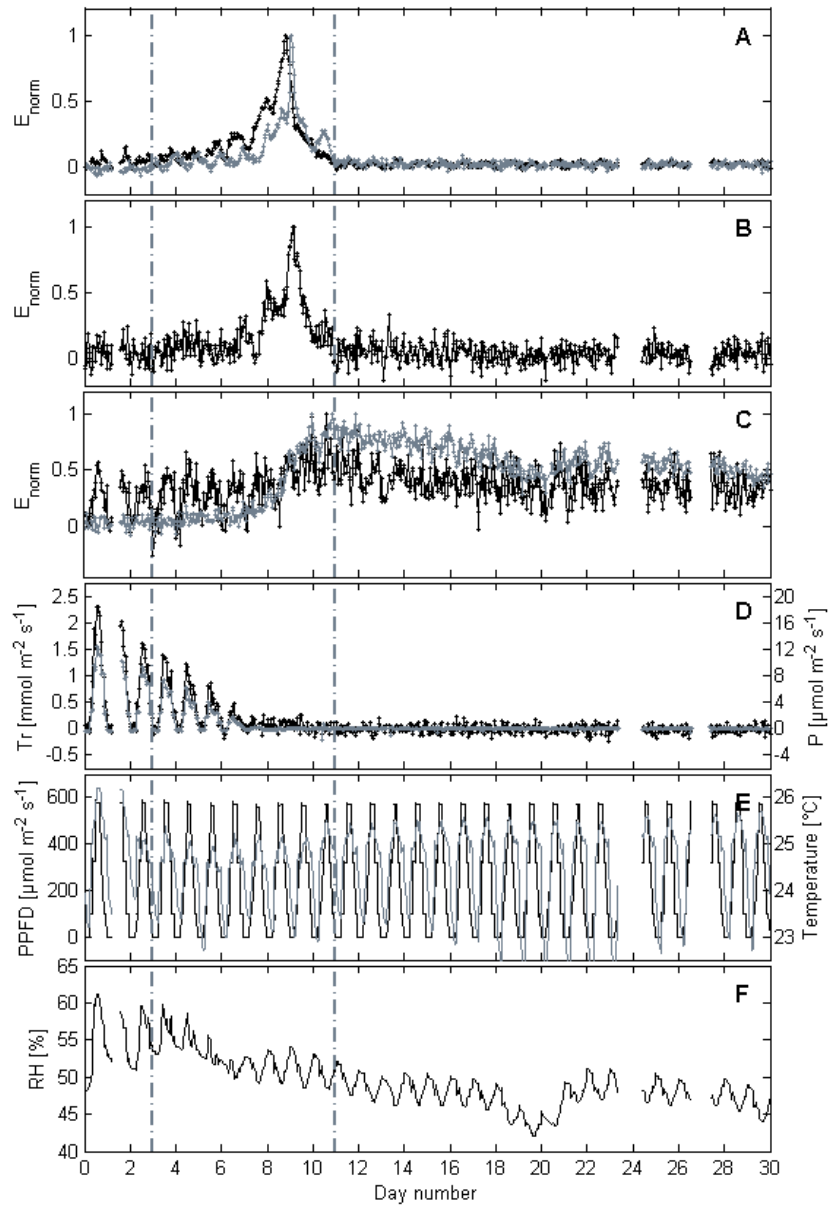
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Figure 2



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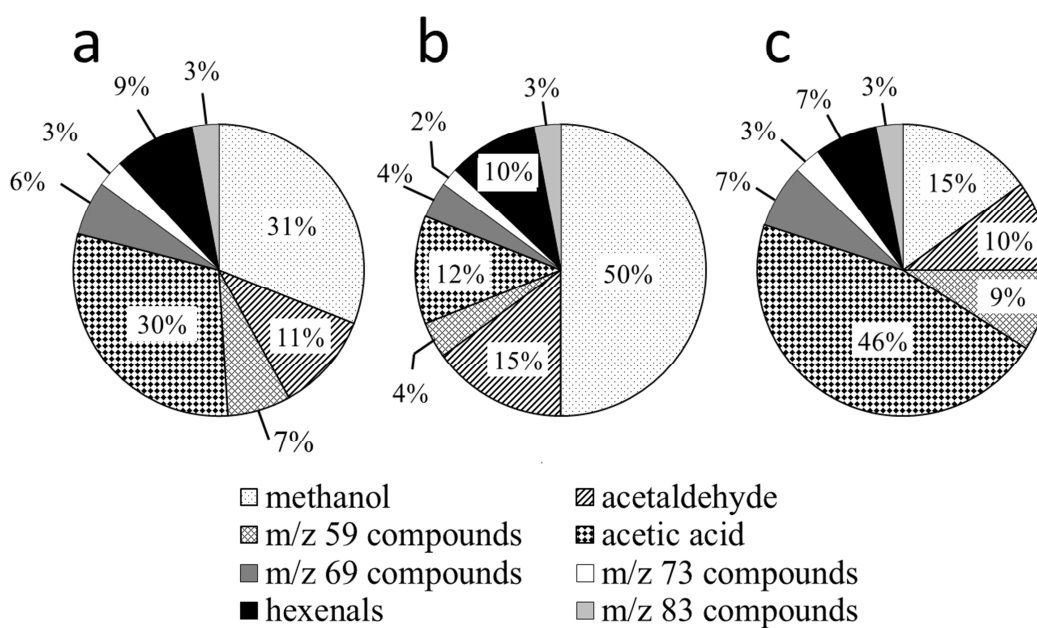
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Figure 3

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Figure captions

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844 **Fig. 1:** Temporal evolution of the daily photosynthesis rate and the daily emission rates of
845 methanol and acetic acid for a specific senescent leaf and corresponding pictures of the leaf taken
846 daily at 10 AM. The different phases of developmental leaf senescence are indicated.

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848 **Fig. 2:** Normalised instantaneous emission rates of methanol (A, black dots), acetaldehyde (A,
849 grey dots), hexenals (B), m/z 59 compounds (C, black dots), and acetic acid (C, grey dots) from a
850 senescent maize leaf since the onset of senescence, together with the transpiration rate (Tr) (D,
851 black dots) , the photosynthesis rate (P) (D, grey dots), light (PPFD) (E, black line), temperature
852 (T) (E, grey line) and relative humidity (RH) (F) conditions in the enclosure. The BVOC
853 emission rates have been normalised with respect to their maximum value to facilitate the
854 comparison of temporal emission profiles for the different compounds. The dot-dashed vertical
855 grey lines indicate the start of the degradation (day 3) and termination (day 11) phase of
856 senescence, respectively. Gaps in the data are due to instrument failure.

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858 **Fig. 3:** Proportion of cumulative BVOC emissions from the senescent leaves for different
859 periods: (a) from the onset of leaf senescence to the end of the investigations, (b) during the leaf
860 chlorosis period (degeneration phase), (c) from the end of the chlorosis period to the end of the
861 measurements (termination phase). Cumulative emissions were averaged over 10 replicates and
862 only compounds which individually made up over 0.5% of the total BVOC emission were
863 considered. Total cumulative BVOC emissions for a, b and c were 0.67 ± 0.11 , 0.33 ± 0.04 and
864 $0.34 \pm 0.08 \text{ mg/g}_{\text{DW}}^{-1}$, respectively.

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