

# 2 • Field measurements of soil respiration: principles and constraints, potentials and limitations of different methods

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## 2.1 INTRODUCTION

Soil respiration is a major component in the carbon balance of terrestrial ecosystems and has been measured in the field for more than eight decades. In this chapter, we will describe the measurement of soil CO<sub>2</sub> efflux at the soil surface that can be considered as equivalent to soil CO<sub>2</sub> production when integrated over long time periods (week, month or season). At shorter time scales the transport of CO<sub>2</sub> may uncouple the soil CO<sub>2</sub> efflux from its production inside the soil. Different methods have been developed to measure this efflux. These methods can affect the object being measured by disturbing the biochemical processes involved in CO<sub>2</sub> production, the physical properties influencing CO<sub>2</sub> movement towards the soil surface, or by changing the environmental conditions in the soil. Therefore, soil respiration measurements in the field are one of the most difficult among the ecosystem flux measurements. So far, no single method has been established as the standard but comparisons, which give important indications on their accuracy, have been performed. The choice of the measurement methodology is not limited to that of a measurement system. The experimenter has to elaborate a protocol depending on the temporal and spatial scales studied. In this chapter, we will describe the most commonly used methodologies for measuring soil CO<sub>2</sub> efflux and present their history, principles and constraints (Section 2.2). In addition, we will present a number of major error sources associated with the different methods and the ways to avoid them (Section 2.3), describe a comparison between different systems (Section 2.4) and give recommendations for the measurement protocol (Section 2.5).

## 2.2 MEASUREMENT PRINCIPLES AND HISTORY OF TECHNICAL DEVELOPMENTS

Soil respiration chambers can be grouped in three categories based on their working principle. A schematic presentation of the working principle and flux calculation of different chamber types is presented in Fig. 2.1. In *closed chambers*, the CO<sub>2</sub> flux is determined from the concentration increase within the chamber's headspace during a known period of time (Fig. 2.1, upper part). Closed chambers can further be divided into two major categories: *closed dynamic chambers* (also known as non-steady-state flow-through chambers) and *closed static chambers* also known as non-steady-state non-flow-through chambers). In *open chambers* (also known as steady-state flow-through chambers) the CO<sub>2</sub> efflux is determined from the difference between CO<sub>2</sub> concentration at the inlet and outlet of the chamber (Fig. 2.1, lower panel).

The Swedish scientist Henrik Lundegårdh (1922; 1924; 1927) was the first to start with measurements of soil respiration in the field. Lundegårdh used a chamber ('respiration bell', Fig. 2.2) with a collar that was driven into the soil. Since there was no exchange with the surrounding air the system was coined a 'closed static' in the later systematic of the methods. The CO<sub>2</sub> concentration inside the respiration bell increased proportionally to the soil respiration. After 10 to 20 minutes Lundegårdh took small samples of air from the chamber, which were analyzed in a self-constructed 'apparatus for accurate analysis of the CO<sub>2</sub> concentration in the air'. This apparatus was based on the absorption of CO<sub>2</sub> by a mixture of KOH and Ba(OH)<sub>2</sub> and the

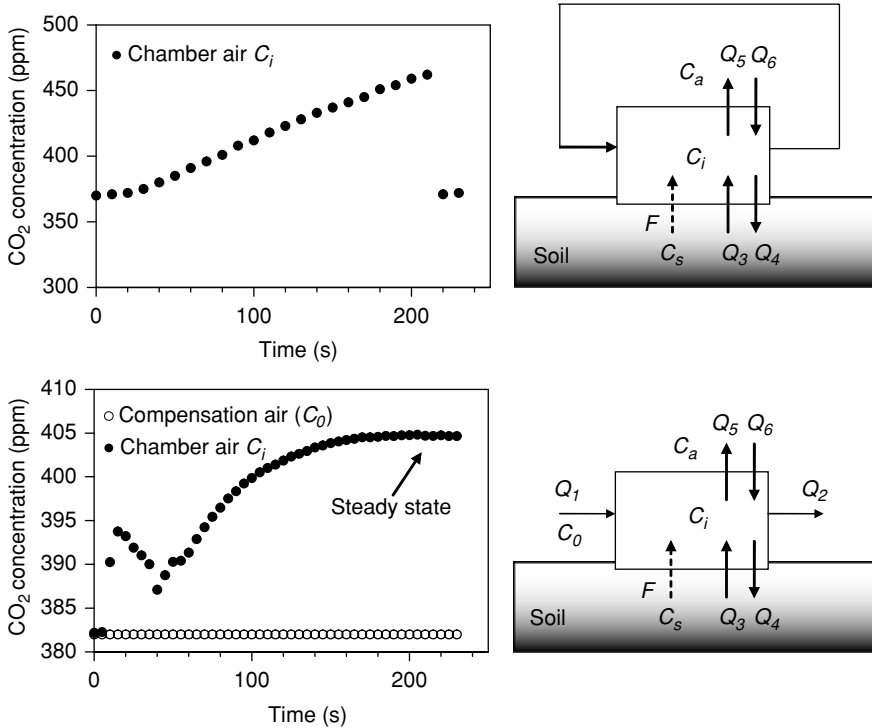


Figure 2.1 (Upper panel) In a closed chamber (dynamic or static) the  $\text{CO}_2$  efflux rate can be calculated from the slope of the  $\text{CO}_2$  concentration increase within the chamber ( $C_i$ ). Similarly, possible air flows between the soil air space and the chamber ( $Q_3$  and  $Q_4$ ) as well as between the ambient air and the chamber ( $Q_5$  and  $Q_6$ ) can generate additional mass flow of  $\text{CO}_2$  in and out of the chamber. When designing both chamber types, air flows of type  $Q_3$ ,  $Q_5$  or  $Q_6$  should be avoided. The  $\text{CO}_2$  efflux from soil ( $F$ ) can be determined using Eq. (2.1).

(Lower panel) The  $\text{CO}_2$  concentration in an open chamber depends on  $\text{CO}_2$  concentrations ( $C_0$  and  $C_i$ ) and flow rates of the incoming and outgoing air flows ( $Q_1$  and  $Q_2$ , which should be equal to avoid pressure differences). In addition, possible air flows ( $Q_3$  and  $Q_4$ ) between the soil air space and the chamber as well as between the ambient air and the chamber ( $Q_5$  and  $Q_6$ ) can generate additional mass flow of  $\text{CO}_2$  in and out of the chamber and should also be avoided. When a steady-state concentration in the chamber has been reached, the  $\text{CO}_2$  efflux from soil ( $F$ ) can be determined from the mass balance shown in Eq. (2.2).

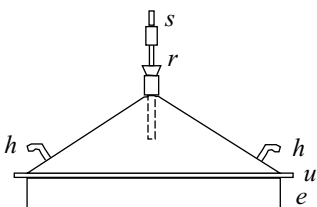


Figure 2.2 Lundegårdh's respiration bell.

measurement of the induced volume change. Between 1921 and 1925 Lundegårdh made a large number of soil respiration measurements combined with measurements of soil temperature, moisture, soil properties, manure and even  $\text{CO}_2$  profiles in the soil as well as in

the lowest parts of the atmosphere. Thus, he was able to make a first analysis of the regulation of soil respiration by climatic and edaphic factors.

The introduction of the first 'open dynamic chamber' by Porkka (1931) was motivated by the observation that the increasing  $\text{CO}_2$  concentration in the closed chamber could reduce the gradient that is driving the diffusive transport. He used a round chamber with an inlet at its top and analyzed the air sucked through this chamber by absorbing the  $\text{CO}_2$  in  $\text{Ba}(\text{OH})_2$  and titration thereafter. In parallel, he analyzed the air at 5 cm above the soil surface and calculated the soil respiration from the difference in the measured  $\text{CO}_2$  concentration.

With Porkka's publication the never ending competition between open and closed systems was opened, and it is interesting to read that most of the important arguments were exchanged even during this early stage of the discussion and eternally repeated thereafter.

Closed chambers seal a certain area of the soil surface completely from the ambient air. They can be further distinguished after the way the CO<sub>2</sub> efflux is determined. In *closed static accumulation chambers*, the CO<sub>2</sub> coming from the soil accumulates inside. The CO<sub>2</sub> concentration increase is determined from air sampled with a syringe and analyzed separately with a CO<sub>2</sub> analyzer (in former times, as Lundegårdh (1927) did, by titration, or as nowadays with either an infrared gas analyzer or a gas chromatograph). Then, the soil efflux ( $F$ ) in  $\mu\text{mol m}^{-2}$  per unit of time ( $t$ ) can be calculated from the increase of the CO<sub>2</sub> concentration within the chamber  $\frac{\Delta C_i}{\Delta t}$ , the chamber volume ( $V_c$ ), the molar volume ( $V_{mol}$ ) and the covered soil area ( $A$ ):

$$F = \frac{\Delta C_i}{\Delta t} \cdot \frac{V_c}{V_{mol} \cdot A} \quad \text{Eq. (2.1)}$$

The  $V_{mol}$  corresponds to the molar volume  $\text{m}^3 \text{mol}^{-1}$  (approx  $22.4 \cdot 10^{-3} \text{ m}^3 \text{mol}^{-1}$  for ideal gas) and is thus equal to  $8.314 \text{ J K}^{-1} \text{mol}^{-1} \cdot T_{\text{air}}/P_{\text{atm}}$ .

Another application of the closed static system is the closed static *absorption method* that was introduced by Lundegårdh (1922). The CO<sub>2</sub> coming from the soil is usually trapped with chemicals such as NaOH or soda lime placed within the chamber. They can be used either as solutions, which are titrated thereafter, or as grains that are dried and weighed. Thus, the CO<sub>2</sub> concentration within the chamber remains relatively stable. The CO<sub>2</sub> efflux can be calculated from the amount of CO<sub>2</sub> bound in the trapping chemical. When soda lime is used for the trapping, the soil efflux measured has also to be corrected to account for the water release during the chemical reaction between CO<sub>2</sub> and soda lime (Edwards, 1982; Grogan, 1998; Janssens *et al.*, 2000). The advantage of the adsorption method over the accumulation method is that it can integrate over periods up to 24 hours and can easily be applied simultaneously at tens or hundreds of soil chambers (Haber, 1958; Janssens and Ceulemans, 1998). The last kind of closed system is the *dynamic* one, where the air circulates between the CO<sub>2</sub> analyzer and the chamber. At each passage through the chamber the soil efflux adds some CO<sub>2</sub> to the air and the CO<sub>2</sub> concentration increases

regularly. The soil efflux is then calculated using Eq. (2.1).

In *open* systems, ambient air is continuously sucked through the chamber. The CO<sub>2</sub> concentration within the chamber (Fig. 2.1) reaches steady-state (equilibrium) after a period of time. Then, the CO<sub>2</sub> flux from the soil into the chamber is balanced by the transport of CO<sub>2</sub> by the air stream through the chamber, and the soil respiration rate ( $F$ ) is determined from the difference between the CO<sub>2</sub> concentration at the inlet and the outlet of the chamber ( $\Delta\text{CO}_2 = C_i - C_o$  in Fig. 2.1, lower panel), the through-flow rate ( $\Phi$ ,  $Q_f = Q_o$  in Fig. 2.1, lower panel) and the soil covered area ( $A$ ):

$$F = \Delta\text{CO}_2 \cdot \frac{\Phi}{A} \quad \text{Eq. (2.2)}$$

However, open systems were not used very often, because they required a higher amount of samples to analyze and a facility to pump the air through the chamber (Porkka used a 12.6 L bottle of water that he slowly emptied). Therefore, the time of open systems did not dawn until about 20 years later when infrared gas analyzers (IRGA) and electrical membrane pumps became available. The first soil respiration system using an IRGA was an open system developed by H. Koepf (1953a). This first open dynamic system gave promising results (Koepf, 1953b, 1954), but was only useful for stationary measurements close to the laboratory, because the IRGA alone was 115 cm high and weighed about 50 kg. In addition, the system needed grid power.

Consequently, the closed 'Lundegårdh bell' remained the standard method for another 20 years, either in the original way with the extraction of an air sample after 10 to 20 minutes or with an amount of absorbent placed inside the chamber.

With the further development of electronics smaller IRGAs became available and more open systems were developed for field measurements (Witkamp, 1969; Witkamp and Frank, 1969; Edwards and Sollins, 1973; Edwards, 1974; Kanemasu *et al.*, 1974; Schwartzkopf, 1978). They enabled much more accurate and continuous measurements as close as possible to the natural conditions. A large variety of systems developed at different research institutions are based on the open chamber technique (Fang and Moncrieff, 1996; Kutsch, 1996; Iritz *et al.*, 1997; Kutsch and Kappen, 1997; Rayment and Jarvis, 1997; Fang and Moncrieff,



Figure 2.3 Open dynamic chamber systems from different manufacturers. PP-Systems CFX-1 at the back and ADC SRS-1000 in the small figure. Automatic closed dynamic system Li-Cor Li-8100 chamber at the front is based on closed dynamic technique.

1998; Longdoz *et al.*, 2000; Rayment, 2000; Kutsch *et al.*, 2001; Pumpanen *et al.*, 2001, 2004) Recently, commercial open systems have also become available (Fig. 2.3).

Closed systems were also further developed with the technical development in electronics that provided smaller and much better portable IRGAs. Modern ‘closed dynamic’ systems recycle the air from the chamber to the analyzer and back, and can monitor the increase in concentration continuously (Fig. 2.1, upper panel, right). In addition, they may scrub the increased CO<sub>2</sub> at the end of a measurement cycle by means of a soda lime column to start a new measurement automatically. Portability and short measuring times in closed dynamic chambers allow the measurement of a high number of frames or collars within a big area, and therefore the estimation of the heterogeneity of soil respiration. Most of the commercial systems are based on the closed dynamic chamber technique (LiCor, PP-systems, ADC, Figs. 2.3, 2.4 and 2.5) but also a lot of researchers went this way (Rochette *et al.*, 1992;

Goulden and Crill, 1997; Rochette *et al.*, 1997; Davidson *et al.*, 2002; Savage and Davidson, 2003).

Two more techniques are commonly used to determine soil respiration in the field: ground level eddy covariance (Norman *et al.*, 1992) and concentration gradients (Albertsen, 1979). The eddy covariance (EC) system is composed of a sonic anemometer and an infrared gas analyzer measuring respectively at the same point the three components of the wind speed and the CO<sub>2</sub> concentration at high frequency (10 Hz or more). With these measurements, it is possible to deduce the CO<sub>2</sub> vertical turbulent flux equal to the mean product of the fluctuations of the vertical wind speed and the CO<sub>2</sub> concentration. It can be demonstrated with fluid mechanic equations (Aubinet *et al.*, 2000) that, according to few assumptions, this vertical turbulent flux at any point above an ecosystem can be equal to the net quantity of CO<sub>2</sub> produced or absorbed by this ecosystem. The principal assumptions are the horizontal homogeneity of the ecosystem and a relatively high level of turbulence (high enough to neglect the transport of CO<sub>2</sub> by



Figure 2.4 Closed chamber systems from different manufacturers (a) PP-Systems SRC-1 closed dynamic chamber, (b) Vaisala GMP-343 CO<sub>2</sub> probe connected to a closed chamber made by Tunkua Oy, Finland. In the figure, a transparent chamber is being used for measuring photosynthesis of ground vegetation.

advection and the CO<sub>2</sub> storage between the measurement point and the soil). Eddy covariance systems are now frequently installed on towers above forests, grasslands and crops (Law *et al.*, 1999) to measure their net ecosystem exchange.

In several studies EC systems were mounted within the trunk space of forests to quantify the soil CO<sub>2</sub> efflux (Norman *et al.*, 1997; Kelliher *et al.*, 1999; Janssens *et al.*, 2001; Shibistova *et al.*, 2002). This is problematic because the presence of tree crowns prevents eddy penetration below the canopy to the measurement level. Thus, the turbulence characteristics close to the forest floor may lead to a bias in the measurements or can induce large time periods with turbulence being too low for a good functioning of the EC system. However, using data coming from forests with a relatively open canopy and, more particularly, selecting measurements performed during a windy period (data quality control routines exist), it might be possible to obtain interesting quantitative information on forest soil CO<sub>2</sub> efflux. The other difficulties introduced by the EC method are the determination of the footprint area and the impossibility to separate vegetation and soil signals, which implies that EC is not really applicable for soil CO<sub>2</sub> efflux measurements on vegetated surfaces. The most important

problem in forests with relatively open canopies, in crop lands and in grasslands is the presence of understorey and ground vegetation. An EC system installed at one or few metres height is then unable to separate the soil and ground vegetation contribution to the net flux. These restrictions reduce the possibilities of using ground level EC systems to measure soil CO<sub>2</sub> efflux, even if it minimizes disturbance at the soil surface and covers a larger area than chambers do.

Soil CO<sub>2</sub> effluxes can also be derived from CO<sub>2</sub> concentration gradients between soil and the atmosphere, and between different soil layers. Carbon dioxide concentration in the soil air space between soil particles is often an order of magnitude higher than in the atmosphere (Fernandez and Kosian, 1987; Suarez and Šimunek, 1993), resulting in a large concentration gradient between the soil and the atmosphere. The primary mechanism for transporting CO<sub>2</sub> from the soil to the atmosphere is molecular diffusion (Freijer and Leffelaar, 1996). According to Fick's first law, the gas flux is dependent on the concentration gradient and the diffusivity of the soil. Thus the CO<sub>2</sub> flux in the soil is usually upwards, resulting in a CO<sub>2</sub> efflux out of the soil (Fig. 2.6). The diffusion of CO<sub>2</sub> depends essentially on the total porosity of soil layers, soil water

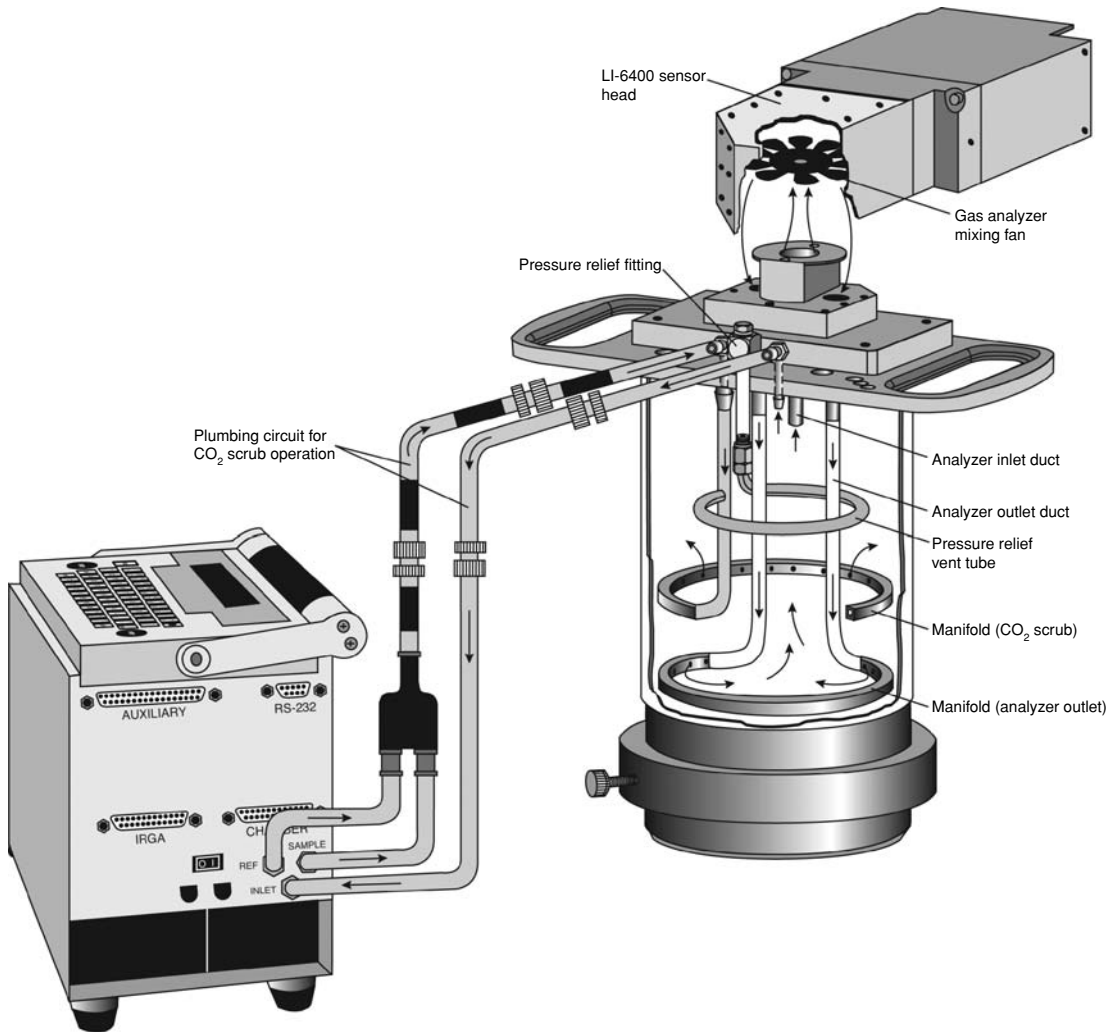


Figure 2.5 An example of the air circulation in a closed dynamic chamber (Li-Cor 6400–09). The air is pumped from the chamber to the IRGA and returned back to the chamber through a perforated manifold. The pressure equilibrium tube is located on top of the chamber (extract from Li-Cor 6400–09 Soil CO<sub>2</sub> flux chamber instruction manual, with the kind permission of Li-Cor Inc., Nebraska USA). In the PP-Systems chamber, the air is mixed by a fan installed in the upper part of the chamber.

content, layer thickness and the concentration gradient between the layers. Gradients of CO<sub>2</sub> concentration are difficult to convert into fluxes because the diffusivity of the soil is heterogeneous and also changes with soil moisture. Despite these difficulties, some studies have been conducted with this technique during recent years (Tang *et al.*, 2003; Pumpanen *et al.*, 2003a, 2003b; Jassal *et al.*, 2004; Tang *et al.*, 2005; Pihlatie *et al.*, 2007, Mykeebust *et al.*, 2008).

The next sections will provide more detailed information on the possible disturbances involved in the measurement systems and the ways to avoid them.

### 2.3 DISTURBANCES INTRODUCED BY THE MEASUREMENT SYSTEM

Systematic errors can be introduced into the flux measurement by disturbing the physical processes involved

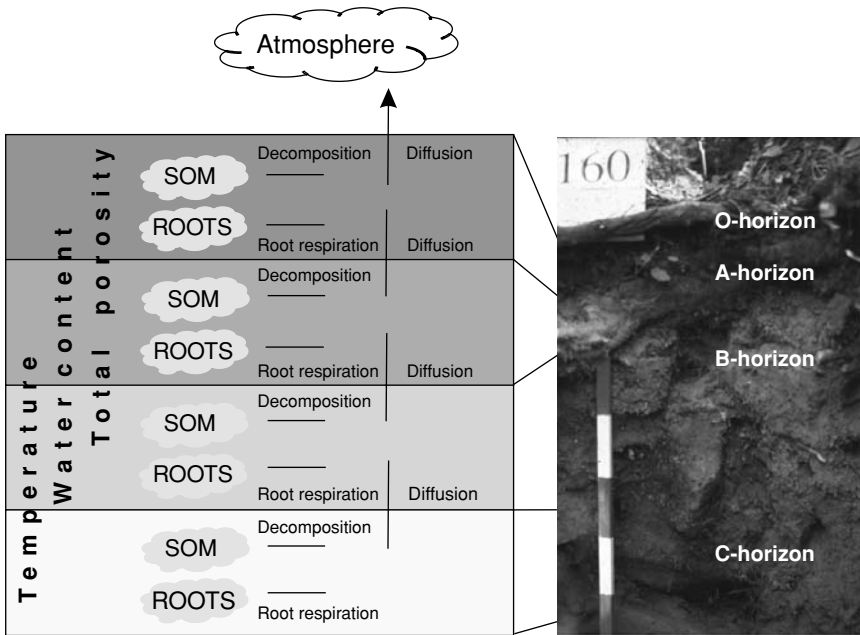


Figure 2.6 Schematic presentation of the soil profile and processes underlying soil CO<sub>2</sub> efflux.

in CO<sub>2</sub> movement within the soil or by modifying the biochemical processes involved in soil CO<sub>2</sub> production. This last disturbance can be induced directly or indirectly by changes in the soil environmental conditions. This section presents these disturbances and the ways to overcome them.

A chamber measurement system does not determine directly the soil CO<sub>2</sub> production but the soil CO<sub>2</sub> efflux. For long-term integration (month or season) these two fluxes can be considered as equivalent. The CO<sub>2</sub> is transported through the soil surface mainly by molecular diffusion but can also be driven by wind (advection or turbulence). Fang and Moncrieff (1999) as well as Lund *et al.* (1999) provide theoretical derivations and model formulations for these two ways of transport.

### 2.3.1 Vertical pressure gradient

In open and closed dynamic chamber systems the artificial air circulation generated by the pump can modify the air pressure just above the soil and thus perturb the vertical pressure gradient. Even a small pressure difference between the inside of the chamber and the atmosphere (PDC), as low as 1 Pa, has been shown to cause significant errors to the measured CO<sub>2</sub> efflux

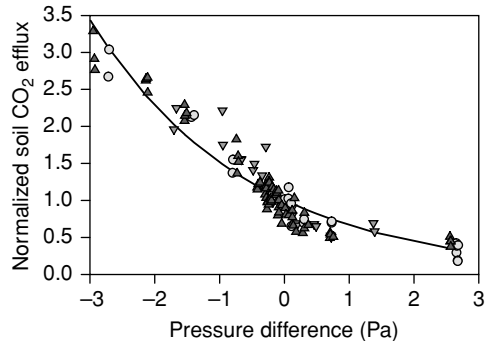


Figure 2.7 Effect of a constant pressure difference between the inside and the outside of the chamber (PDC) on soil CO<sub>2</sub> efflux in the forest of Vielsalm (Belgium). The flux is normalized by its value at PDC = 0 and negative PDC corresponds to a pressure in the chamber being lower than the atmospheric pressure.

(Kanemasu *et al.*, 1974; De Jong *et al.*, 1979; Fang and Moncrieff, 1996; Kutsch, 1996; Fang and Moncrieff, 1998; Lund *et al.*, 1999; Longdoz *et al.*, 2000; Kutsch *et al.*, 2001). Figure 2.7 shows an example of the dependence of soil CO<sub>2</sub> efflux on the PDC for the forest soil of Vielsalm in Belgium.

In a closed dynamic chamber, the reproduction of the natural vertical pressure gradient near the soil

surface is the main problem (Kanemasu *et al.*, 1974; De Jong *et al.*, 1979; Fang and Moncrieff, 1998). In these systems, pumps may create an overpressure or underpressure at its two faces. The air circulates from the overpressure face to the underpressure face along a pressure gradient. The pressure at any point is constant and depends only on its position in the circuit. Thus, the pressure in the chamber can be lower or higher than in the atmosphere. Underpressure generates an artificial mass flow of CO<sub>2</sub> from the soil into the chamber and leads to an overestimation of the efflux, whereas overpressure can block the natural CO<sub>2</sub> efflux from the soil leading to an underestimation (Norman *et al.*, 1992; Striegl *et al.*, 1992).

The CO<sub>2</sub> efflux is more sensitive to the suction of air than to a slight overpressure especially in a porous soil (Longdoz *et al.*, 2000; Pumpanen *et al.*, 2001). The mass flow of CO<sub>2</sub> from soil pores also increases with increasing pore space CO<sub>2</sub> concentration. Thus the impact of the PDC increases with the permeability and CO<sub>2</sub> productivity of the soil. To reduce the PDC, some chambers have a small hole with a tube connecting the inside and outside of the chamber and maintaining the pressure equilibrium. Unfortunately, this tube is not always sufficient to transmit the high-frequency atmospheric pressure fluctuations into the chamber, which can have a significant impact on the soil efflux measurement (Longdoz *et al.*, 2000; Pumpanen *et al.*, 2001; Takle *et al.*, 2004).

Another possible origin of PDC in closed dynamic chambers is the effect of the air mixing inside the chamber. In closed chambers, the determination of CO<sub>2</sub> efflux is based on the assumption that the concentration is homogeneous within the chamber's headspace (Fig. 2.1). In some commercial systems (e.g. Li-Cor), the air mixing is assured by extracting the sample air from the upper part of the chamber and by pushing air from the analyzer back into the chamber through a perforated manifold circulating around the chamber. In PP-systems SRC-1 chamber, the air is mixed by a fan installed on top of the chamber and inducing a vertical air flow in the chamber's headspace. Inevitably, the air mixing changes the natural turbulent conditions in the chamber. The perturbation provoked can be significantly reduced when small airflow is used or when the turbulence generated by the fan is reduced by placing a metal mesh between the fan and the soil surface. This latter option is used in the latest version of the

PP-systems soil respiration chamber. For all the closed chambers a measurement of the PDC value is recommended before starting measurements.

In open systems, pressure problems arise if the inlet size and through-flow are not balanced such that pressure differences occur, or if an open system does not only transmit but also modifies the natural pressure fluctuations due to an inappropriate design of the chamber inlet. For example, Kutsch *et al.* (2001) placed the inlet at one side of the chamber and thus caused an overpressure inside the chamber when the wind blew directly towards the inlet and an underpressure with all other wind directions. Therefore, the open chamber system has to be designed very carefully to avoid artificial underpressure or overpressure during the measurements.

Pump(s) can also induce PDC in open chamber systems (Rayment and Jarvis, 1997; Fang and Moncrieff, 1998) if the inlet and outlet airflow are not the same. Flow rate differences can be induced by the resistance at the inlet or outlet of the chamber. It may not be enough just to let the air be drawn into or out of the chamber. Even if the inlet and the outlet are large in diameter, there is always a small PDC as long as the air is only sucked into or out of the chamber. Preferably, an equal amount of compensating air should be fed into the chamber as is drawn out of the chamber. Separate pumps and mass flow controllers can be used to control the flow rates of air in and out of the chamber. Basically, this is the only way to avoid the pressure effect. However, in practice mass flow controllers can not reproduce the natural pressure fluctuation induced by the wind just above the soil surface. Nevertheless, at least systematic PDC caused by suction of air can be avoided by using separate pumps and mass flow controllers on both sides of the chamber. If it is not possible to organize separate pumps and mass flow controllers, it is of utmost importance to monitor the PDC to detect the possible pressure differences.

### 2.3.2 Vertical CO<sub>2</sub> concentration gradient

By contrast to the EC and gradient systems, all of the chamber systems can disturb the vertical CO<sub>2</sub> concentration gradient in different ways. They may modify the air CO<sub>2</sub> concentration in the chamber or disturb the soil CO<sub>2</sub> concentration by perturbing soil production directly or by changing the soil temperature or water



content (Healy *et al.*, 1996; Lund *et al.*, 1999). In absorption-based closed static chamber techniques, the CO<sub>2</sub> efflux is affected by two different mechanisms: by altered CO<sub>2</sub> concentration inside the chamber and by lack of turbulence. When soil respiration is high the CO<sub>2</sub> concentration inside the chamber and in the soil is high. At the same time, the lack of turbulence will have a large effect on the CO<sub>2</sub> efflux, because the only mechanism of CO<sub>2</sub> transport is then diffusion. This will create a severe underestimation. When the soil respiration is low, soda lime reduces CO<sub>2</sub> concentration inside chamber headspace, typically below ambient, and the resulting larger concentration gradient causes overestimation of fluxes. This overestimation is partly offset by the reduced turbulence, but this is much less important when soil respiration is low because soil CO<sub>2</sub> concentration is lower. These mechanisms can cause overestimation of low fluxes and underestimation of high fluxes (Janssens and Ceulemans, 1998).

In the closed static accumulation chamber, the CO<sub>2</sub> concentration can exceed that of the ambient air. This saturation effect changes the natural concentration gradient within the soil surface and may reduce significantly the CO<sub>2</sub> efflux (Nay *et al.*, 1994; Livingston and Hutchinson, 1995; Davidson *et al.*, 2002). So, this technique is known to underestimate the soil efflux (Janssens and Ceulemans, 1998).

In the closed dynamic system, the CO<sub>2</sub> concentration in the chamber rises by a few tens of ppm above atmospheric value during the measurement (over a few minutes). The saturation effect generated is smaller than that in the static accumulation system and can be further reduced by passing the chamber air through a CO<sub>2</sub> scrubber before the measurement record period. The perturbations of the concentration gradient can be minimized by starting the measurement just below and finishing just above the ambient CO<sub>2</sub> concentration. This CO<sub>2</sub> scrubbing technique is used, for example, in some Li-Cor chambers. The CO<sub>2</sub> concentration in the chamber's headspace is scrubbed down by a few tens of ppm below the ambient target concentration by pumping the air in the chamber through an absorber column. Then the CO<sub>2</sub> concentration is monitored as it rises above the target, and the CO<sub>2</sub> efflux is calculated from the rate of increase of CO<sub>2</sub> concentration at around the ambient concentration. With this technique, it is possible to avoid the effects of saturation on CO<sub>2</sub> diffusion from the soil to the chamber. A good

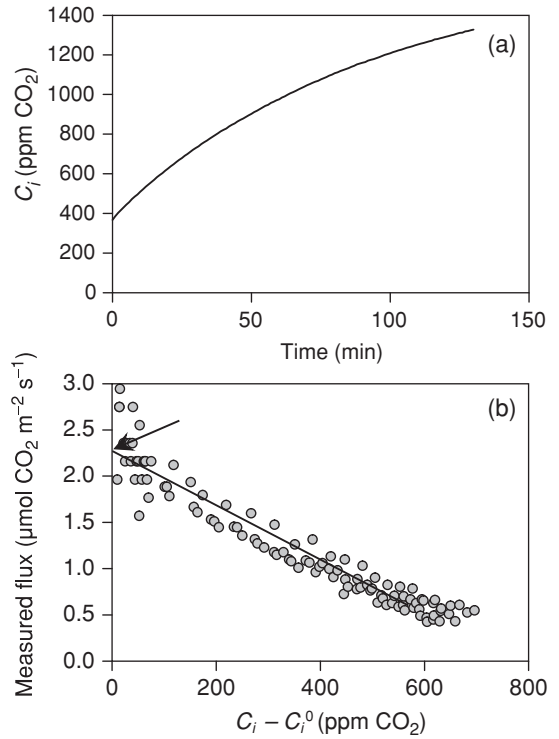


Figure 2.8 An example of the concentration development in a closed dynamic chamber (a) and the respective CO<sub>2</sub> efflux plotted against the CO<sub>2</sub> build-up within the chamber's headspace (b). The measured CO<sub>2</sub> efflux decreases along with increasing CO<sub>2</sub> concentration in the chamber due to the slow down in diffusion rate. The CO<sub>2</sub> efflux decreases rapidly at the beginning of the measurement period. A peak in the CO<sub>2</sub> efflux during the first minute or two of the measurement period may be due to pressure perturbations induced by the chamber or by the fan. After 1–2 minutes the CO<sub>2</sub> efflux decreases linearly in relation to the concentration change in the chamber. In this case, the disturbance at the beginning of the measurement period could be corrected by interpolating linearly to zero in Fig. (b). The actual CO<sub>2</sub> efflux from soil driven by diffusion is marked with an arrow.

indicator to see if saturation has a significant impact on the efflux is to detect the decrease in the slope of the time evolution of CO<sub>2</sub> concentration or a low  $r^2$ -value in the linear fit on  $\Delta C/\Delta t$  (Fig. 2.1). The decrease in time evolution of CO<sub>2</sub> concentration and soil CO<sub>2</sub> efflux is illustrated in Fig. 2.8. The saturation effect should be taken into account when designing the dimensions of a closed chamber and the length of the measurement period. The volume vs. surface-area ratio of the chamber determines its sensitivity. For example, if the

volume vs. surface-area ratio is low the system will detect even a small concentration increase and is thus applicable to measure small fluxes. However, in this case, the saturation of CO<sub>2</sub> concentration will take place faster, which is a problem at high CO<sub>2</sub> effluxes. Thus the volume vs. surface-area ratio should be designed bearing in mind the efflux range to be measured.

In the open dynamic system, the CO<sub>2</sub> concentration in the chamber is constant but slightly above the ambient one. To make sure that the impact of this on the efflux is negligible, a closed dynamic chamber can be used for comparison. The slope of the time evolution of CO<sub>2</sub> at the concentration recorded in the open system should be equal (very close) to the slope obtained at the ambient concentration (Longdoz *et al.*, 2000).

### 2.3.3 Horizontal wind

The EC and gradient methods are the only ones that do not disturb the natural horizontal wind. Chambers are unable to reproduce natural wind conditions. In closed-static chambers (Edwards, 1982; Crill, 1991) wind velocity is zero. The only mechanism of CO<sub>2</sub> transport is then diffusion. In dynamic-closed chambers and in open chambers wind velocity is determined by the airflow rate and by the chamber's design. Thus the boundary layer thickness and consequently the soil CO<sub>2</sub> efflux will be altered. To our knowledge, the separate effect of this perturbation has not been studied systematically to date (this effect is difficult to distinguish from the PDC perturbation). Recently, Bain *et al.* (2005) studied the effect of horizontal wind speed on the PDC and found that horizontal wind induced substantial error in the CO<sub>2</sub> efflux measurements also in closed dynamic chambers through the so-called 'Venturi effect', which was first reported by Conen and Smith (1998). In the Venturi effect, the wind movement around the vent of a closed chamber depressurizes the chamber by pulling air out of the chamber headspace resulting in a mass flow of soil gases from the porous soil into the chamber interior (Bain *et al.*, 2005). This effect was shown to be very significant. According to Conen and Smith (1998), a steady wind speed of 2 m s<sup>-1</sup> resulted in a 233% increase in measured soil emissions, and even in very calm conditions, with wind speed less than 1 m s<sup>-1</sup>, systematic errors of 10 to 50% were discovered. Similar results were later confirmed by Bain *et al.* (2005). The wind direction and speed on the

soil surface usually fluctuates leading to unpredictable pressure variations inside the chamber. The Venturi effect and other anomalous pressure effects resulting from wind turbulence can be studied by testing the chamber vent and possible PDC by fast-response differential pressure sensor on a non-permeable plate and on different soil types and wind conditions. The pressure differences resulting from the Venturi effect may not be seen if the tests are conducted on a porous soil only, because the air flow through the porous soil into the chamber may compensate for the pressure loss in the chamber induced by the Venturi effect. Recently, Xu *et al.* (2006) suggested a new vent type for decreasing the Venturi effect in the chambers. The vent allows pressure inside the chamber to track pressure at the soil surface outside the chamber. The vent is designed so that it slows down the wind velocity within the vent such that the dynamic pressure changes induced by the wind are converted to static pressure, which the chamber equilibrates. They have tested this new vent design in field conditions with promising results.

### 2.3.4 Other effects

Usually chambers seal the soil surface either by pushing the chamber on or into the bare ground ('insertion method') or by placing the chamber on a collar penetrating the soil surface ('collar method'). This latter mode of chamber-soil contact assures airtight connection between the chamber and the soil, which is of advantage especially if the soil is porous and the measurement place is subjected to winds. Compared to the insertion of the chamber on the soil only, the collar has the advantage of reducing the risk of CO<sub>2</sub> leakage out of the chamber and facilitates repeated measurements on the same spot. However, the drawback of collars is that roots are cut during insertion into the soil (the trenching effect). Consequently, because roots contribute significantly to soil respiration (Hanson *et al.*, 2000), the soil CO<sub>2</sub> concentration and the efflux will be affected. To reduce the trenching effect, the collars should not be pushed down into the rooting zone. If this cannot be avoided, the collars should be left in place until the roots have re-grown inside the collar. The recovery of fine roots may take several years. Makkonen and Helmisaari (1999) studied fine-root biomass growth with root ingrowth core method in a boreal forest. The ingrowth cores were initially without root biomass and

no levelling off in the biomass growth was found during the first three growing seasons. If installed very deeply in the soil, the collars may produce a similar effect. However, most of the damage may be mitigated by flat collars that have spikes to keep the collar grounded but do not trench the whole root system.

Recent studies confirmed that the chamber–soil contact mode has a significant impact on the apparent soil respiration rates (Ngao *et al.*, 2006) with generally higher values obtained for the ‘insertion’ mode. Even if both chamber–soil contact modes have advantages and disadvantages, it can be concluded that the advantages of using a collar prevail. Therefore, the use of collars is recommended.

The heating of the chambers in the sun when the measurement time is long and the chamber is transparent can affect the temperature and water status of the ground vegetation and soil surface inside the chamber. The temperature can increase by up to 10 to 15 °C during chamber closure on a warm, sunny day. This may disturb the respiration rate and results in changes in the CO<sub>2</sub> concentration gradient. Therefore, a thermostatically controlled cooling system would be recommended, especially if the chamber is transparent and closed. Heating of the chambers in the sun should also be taken into consideration as a possible cause of pressure–gradient perturbation, especially when the measurement time is long and the chamber is transparent. The temperature increase results in a physical expansion of air inside the chamber, which may result in an overpressure blocking soil efflux and creating a mass flow of CO<sub>2</sub> out of the chamber through the equilibrium tube or through the soil.

## 2.4 COMPARISON OF THE EXISTING SYSTEMS AND RECOMMENDATIONS

Several recent studies compared the accuracy of different systems. Janssens *et al.* (2000) as well as Shibistova *et al.* (2002) found systematically lower values of soil respiration measured by ground level EC compared to chamber measurements. In the study by Janssens *et al.* (2000) this underestimation was correlated with photosynthetically active radiation (PAR), suggesting a confounding effect of the ground vegetation. Chamber comparisons were either conducted in the field by comparing the apparent fluxes at the same collars (Janssens *et al.*, 2000), by artificial systems with known fluxes

(Pumpanen *et al.*, 2004) or by a combination of both (Butnor *et al.*, 2005). The comparisons have indicated relative differences between chamber types (Raich *et al.*, 1990; Norman *et al.*, 1997; Janssens *et al.*, 2000; Pumpanen *et al.*, 2004) or showed chamber-specific limitations (Nay *et al.*, 1994; Fang and Moncrieff, 1998; Gao and Yates, 1998).

In most cases, closed dynamic and static chambers have been shown to give systematically lower values than open dynamic chambers – the difference ranging from 10% (Rayment, 2000) to 40–50% (Norman *et al.*, 1997; Pumpanen *et al.*, 2003a). However, in a recent study (Pumpanen *et al.*, 2004), the differences between chambers using different measurement principles seemed not to be consistent. When most of the currently available chamber systems were compared against known CO<sub>2</sub> effluxes generated by a calibration chamber in laboratory conditions, their reliability appeared to be independent of the measurement principle as such (Pumpanen *et al.*, 2004). Instead, the geometrical design of the chamber, the mixing of air inside the chamber and the collar model seemed to affect the measured CO<sub>2</sub> efflux more than the measurement principle. Even identical chambers with different collar designs showed highly variable results. However, the general trend seemed to be that closed static chambers underestimated CO<sub>2</sub> effluxes by 4 to 14%. No systematic differences were found between open dynamic chambers and closed dynamic chambers (Pumpanen *et al.*, 2004). An extract from the recent paper by Pumpanen *et al.* (2004) is presented in Table 2.1. In Table 2.2, we have listed the advantages and disadvantages of the major chamber systems. A standard chamber will hardly be available, because different ecosystems require different chamber designs. One technical solution may not be the best for all purposes. For example, a small chamber suitable for forest with abundance of stones and small shrub vegetation is probably not suitable for measuring grasslands. The only reasonable way to standardize various chamber systems is to compare them against known CO<sub>2</sub> effluxes.

In conclusion, the method used for measuring soil CO<sub>2</sub> efflux should be chosen based on the research conducted and the type of ecosystem (see Section 2.5). The chamber methods affect the flux being measured, but this error can be detected and corrected if the chambers are tested against known CO<sub>2</sub> effluxes. Reliable CO<sub>2</sub> efflux measurements can be carried out with open as

Table 2.1 (Extract from Pumpanen et al. (2004)). Correction factors for different chambers. Each chamber can be scaled to the reference flux obtained from the calibration tank by dividing the measured flux by the correction factor of a specific soil type.

Chamber type <sup>a</sup>	Coarse sand	95% confidence interval	Dry fine sand	95% confidence interval	Wet fine sand	95% confidence interval
NSF-1 (Licor 6400-09)	1.01	0.99–1.03	1.01	0.98–1.04	1.05	1.01–1.09
NSF-1b (Licor 6400-09)	1.13	1.07–1.18	1.09	0.98–1.19	1.09	1.04–1.14
NSF-2 (EGM-3 + SRC-1)	1.21	1.17–1.26	1.27	1.15–1.39	1.05	0.97–1.13
NSF-3 (EGM-3 + SRC-1 widened collar)	0.86	0.82–0.89	1.00	0.94–1.05	–	–
NSF-4 (EGM-1 + SRC-1 no collar)	1.03	1.01–1.06	1.19	1.14–1.24	0.94	0.86–1.03
NSF-5 (EGM-4 + SRC-1 mesh)	1.16	1.12–1.19	1.19	1.11–1.27	1.33	1.20–1.47
NSF-6 (University of Bayreuth)	0.96	0.91–1.02	0.89	0.86–0.92	0.96	0.87–1.06
NSF-7 (Finnish Meteorological Institute)	1.03	1.01–1.05	1.07	0.99–1.15	1.00	0.92–1.08
NSF-8 (Woodshole Research Center)	0.83	0.79–0.86	0.91	0.86–0.96	0.83	0.80–0.85
NSF-9 (Max Planck Institute)	0.81	0.79–0.83	0.80	0.79–0.82	0.79	0.77–0.80
NSF-10 (University of Helsinki)	1.01	0.96–1.05	1.19	1.14–1.23	1.04	0.96–1.13
NSF-11 (University of Helsinki)	1.00	0.96–1.03	0.85	0.81–0.87	0.87	0.84–0.89
NSF-12 (University of Helsinki)	–	–	1.13	1.08–1.18	0.93	0.87–0.99
NSF-Average	1.00	–	1.04	–	0.99	–
NSNF-1 (University of Joensuu)	0.98	0.95–1.01	0.94	0.89–0.98	0.85	0.81–0.88
NSNF-1 (University of Joensuu with extension)	0.95	0.86–1.05	0.98	0.92–1.03	0.85	0.75–0.94
NSNF-2 (Agrifood Research Finland, 10 min.)	0.96	0.91–1.01	0.96	0.76–1.15	0.95	0.84–1.06
NSNF-2 (Agrifood Research Finland, 30 min.)	0.85	0.79–0.90	0.85	0.71–0.98	0.90	0.80–1.00
NSNF-3 (University of Helsinki)	1.06	0.96–1.17	0.82	0.63–1.01	0.85	0.78–0.93
NSNF-4 (University of Helsinki)	–	–	0.65	0.56–0.74	0.84	0.81–0.87
NSNF-Average	0.96	–	0.86	–	0.87	–
SSFL-1 (University of Bayreuth)	1.03	1.01–1.05	0.96	0.92–1.01	1.09	1.02–1.15
SSFL-2 (University of Kiel)	1.05	0.99–1.11	1.08	1.01–1.15	0.95	0.80–1.09
SSFL-Average	1.04	–	1.02	–	1.02	–

<sup>a</sup> NSF = non-steady-state flow-through chamber (closed dynamic chamber); NSNF = non-steady-state non-flow-through chamber (closed static accumulation chamber); SSFL = steady-state flow-through chamber (open dynamic chamber).

Table 2.2 *Benefits and drawbacks of different techniques.*

	Open dynamic	Closed dynamic	Closed static	Gradient	Eddy covariance
+	No build up of CO <sub>2</sub> in the chamber	Easily commercially available	Easy to construct	Does not disturb the CO <sub>2</sub> efflux	Does not disturb the CO <sub>2</sub> efflux
+	Can be used for measuring photosynthesis	Small and practical to carry in the field	Low costs	Suits well for long-term monitoring	Suits well for long-term monitoring
+	No need for tall collars and therefore no trenching effect	Fast measurements	Can be used for spatial sampling	Maintenance free	
+		Can be used for measuring photosynthesis		Monitoring of below-ground dynamics of CO <sub>2</sub>	
-	Sensitive to pressure differences between the inside and outside of the chamber	Build up of CO <sub>2</sub> in the chamber	Build up of CO <sub>2</sub> in the chamber	Expensive, because it requires several CO <sub>2</sub> sensors and water content measurements	Problems associated with the footprint area
-	Low mobility	Sensitive to turbulence inside the chamber	Long measurement time	Turbulent transport close to the soil surface is difficult to measure	Low accuracy in non-turbulent conditions under the tree canopy
-	Requires mass flow controllers (expensive)	Temperature warms up in the chamber if transparent	Low time resolution if used with CO <sub>2</sub> trapping chemical	Sensitive to the concentration measurements	Advection problems on hill slope sites
-	Not easily available commercially	Trenching effect associated with the collars	Temperature dependency of trapping chemicals	Needs additional information on soil for determining the soil CO <sub>2</sub>	Cannot separate respiration and photosynthesis
-	May change the moisture conditions if the chamber persistently in the same location		Trenching effect associated with the collars	Diffusion coefficient in the soil is difficult to determine	Expensive and complicated technology
-				Disturbs the soil when installing the equipment	Low mobility

well as with closed dynamic chambers if the chamber is designed taking into account the following issues.

- Special attention should be paid to the mixing of air within the chamber, because it can be a major source of error. Excessive turbulence inside the chamber can cause over- or underpressure compared to natural ambient conditions, which can lead to artificial mass flow of CO<sub>2</sub> between soil and the chamber. This is particularly important when using closed dynamic chambers where the CO<sub>2</sub> concentration has to be evenly distributed within the chamber in order to calculate the flux correctly. The air mixing should be efficient enough to provide homogeneous CO<sub>2</sub> concentration within the chamber's headspace, but weak enough not to cause pressure anomalies.
- The headspace concentration inside the chamber affects the flux by altering the concentration gradient between the soil and the chamber, and therefore the chamber should be designed so that the increase in CO<sub>2</sub> concentration of the chamber headspace is as small as possible.
- Collars, which are the recommended chamber–soil contact method, should be designed for minimum disturbance of the root system. This can be solved in different ways either by using spikes or by a sufficient delay between collar insertion and measurement.

## 2.5 EXPERIMENTAL DESIGN

The experimental design of a measurement campaign depends on the space and time scales studied. The measurement system and protocol have to be adapted to the type of ecosystems in which the soil CO<sub>2</sub> efflux measurements are performed. In studies requiring continuous monitoring of soil CO<sub>2</sub> efflux over a long time period, the measurement technique should cause minimum disturbance on the soil. Unfortunately, most of the chamber techniques cannot fulfil this requirement. The eddy covariance technique would be an ideal method for long-term monitoring, because, unlike the chamber methods, it does not affect the processes underlying soil CO<sub>2</sub> effluxes. It is efficient for analyzing the time evolution with small time steps (half-hour) even over a long time period but only in open forests (Baldocchi and Vogel, 1996; Black *et al.*, 1996; Law *et al.*, 1999) and for an area of several square metres.

Another method causing relatively small disturbance of the measurement object is the gradient method. However, the use of this method is rather uncommon because the flux data are very sensitive to the soil diffusivity, a parameter that is difficult to determine, and to the accuracy of the CO<sub>2</sub> concentration measurements. The accuracy of this method is still debatable. However, on grasslands and in forests with dense ground vegetation it could be a good alternative, because EC and chambers do not dissociate ground vegetation exchange and soil CO<sub>2</sub> efflux. The gradient method is better suited for long-term monitoring of CO<sub>2</sub> effluxes from a relatively small surface area rather than for determining statistically representative flux estimates over a large area. This is because the soil diffusivity and soil air CO<sub>2</sub> concentration often vary at the small spatial scale and the equipment needed to capture this spatial variability would be rather expensive, especially if the area to be measured were heterogeneous. The gradient method is at its best when studying the CO<sub>2</sub> effluxes on a process level. The gradient measurement itself does not disturb the CO<sub>2</sub> efflux after the initial disturbance from the installation of the measurement devices. The devices should be installed in an undisturbed soil, by soil core removing, and should avoid as much as possible root cutting and modifications in the diffusion properties of the soil. However, at the same time, continuous vertical pathways from the CO<sub>2</sub> sensors to the soil surface should be avoided, because they change the hydraulic properties of the soil by providing a water passage from the soil surface to the sensor. In addition, the CO<sub>2</sub> diffusion along these vertical pathways may be faster than in the natural soil. This can result in erroneous CO<sub>2</sub> concentration measurements and consequently lead to biased CO<sub>2</sub> efflux estimates.

In view of the difficulties met by the gradient method (see above), the dynamic chamber systems (open or closed) appear to be the most adequate on vegetated soil. The open chamber measures with a high temporal resolution but has to fulfil some requirements before being applicable for long-term measurements. After a few days the soil conditions (temperature and water content) can be significantly affected by the perpetual presence of the chamber if the chamber top is not opened in between the measurements. If opened between the measurements, the open dynamic chamber needs some time to reach steady-state conditions. This equilibrium time depends on the flow rate and on

the CO<sub>2</sub> efflux. In addition, the surface area that can be sampled by the open chamber is small because one chamber covers only a few tens to hundreds of square centimetres and the scaling-up of the system is often expensive. Therefore, open chambers may not be the best in studies requiring good spatial coverage.

Spatial sampling is often carried out with portable chambers. The closed static chamber systems have largely been abandoned because of the systematic errors in this technique (the tendency to underestimate the efflux). For spatial sampling, manual or automated closed dynamic chambers may be a better option. Manual chambers can be moved from one collar to another by the experimenter. Because the measurement time is relatively short, he or she is able to collect data representing the efflux of up to one or more hectares in one day. In this condition, it is possible to integrate the spatial variability of an entire plot. However, this integration is done to the detriment of the temporal resolution of the measurement on the same spot. Temporal changes in soil respiration have to be taken into consideration when studying the spatial variation with manual chambers. Ideally, the temperature fluctuation during the spatial sampling should be as small as possible. The measurements carried out at different temperature conditions on different measurement collars cannot be used for averaging the efflux over the whole measurement plot, or at least the CO<sub>2</sub> efflux values measured from individual collars should be corrected for the difference in temperature. This in turn requires information on the temperature response of the individual collars.

Best results can probably be obtained by combining continuous monitoring of CO<sub>2</sub> efflux with automated chambers and spatial measurements with manual chambers. This way, it is possible to obtain both good temporal resolution and spatial representativeness. There are already automatic closed chambers available, which are attached to one collar at a time and measured at a chosen frequency. Then the short-term CO<sub>2</sub> evolution (half-hour) can be obtained over a long time period (season or even year) from a number of points depending on the financial resources. However, the representative area sampled cannot be as large as with a manual system.

The experimental design and the measurement protocol applied in individual studies is usually a trade-off between the technical and human resources available. It is the experimenter's task to plan the

measurements so that they, on the one hand, provide enough information on spatial variation for sound statistical analysis and, on the other hand, provide data with high enough temporal resolution for studying the processes underlying soil CO<sub>2</sub> efflux and all this with feasible costs. We hope that the issues discussed in this chapter will help the reader to solve this equation.

## REFERENCES

- Albertsen, M. (1979) Carbon dioxide balance in the gas-filled part of the unsaturated zone, demonstrated at a Podzol. *Z Pflanz Bodenk*, **142**, 39–56.
- Aubinet, M., Grelle, A., Ibrom, A. *et al.* (2000) Estimates of the annual net carbon and water exchange of forests: the EUROFLUX methodology. *Advances in Ecological Research*, **30**, 113–75.
- Bain, W. G., Hutrya, L., Patterson, D. C. *et al.* (2005) Wind-induced error in the measurement of soil respiration using closed dynamic chambers. *Agricultural and Forest Meteorology*, **131**, 225–32.
- Baldocchi, D. D. and Vogel, C. A. (1996) Energy and CO<sub>2</sub> flux densities above and below a temperate broad-leaved forest and boreal pine forest. *Tree Physiology*, **16**, 5–16.
- Black, T. A., den Hartog, G., Neumann, H. H. *et al.* (1996) Annual cycles of water vapour and carbon dioxide fluxes in and above a boreal aspen forest. *Global Change Biology*, **2**, 219–29.
- Butnor, J. R., Johnsen, K. H. and Maier, C. A. (2005) Soil properties differently influence estimates of soil CO<sub>2</sub> efflux from three chamber-based measurement systems. *Biogeochemistry*, **73**, 283–301.
- Conen, F. and Smith, K. A. (1998) A re-examination of closed flux chamber methods for the measurement of trace gas emissions from soils to the atmosphere. *European Journal of Soil Science*, **49**, 701–7.
- Crill, P. M. (1991) Seasonal pattern cycles of methane uptake and carbon dioxide release by a temperate woodland soil. *Global Biogeochemistry*, **5**, 319–34.
- Davidson, E. A., Savage, K., Verchot, L. V. and Navarro, R. (2002) Minimizing artifacts and biases in chamber-based measurements of soil respiration. *Agricultural and Forest Meteorology*, **113**, 21–37.
- De Jong, E., Redmann, R. E. and Ripley, E. A. (1979) A comparison of methods to measure soil respiration. *Soil Science*, **127**, 300–6.
- Edwards, N. T. (1974) A moving chamber design for measuring soil respiration rates. *Oikos*, **25**, 97–101.

- Edwards, N. T. (1982) The use of soda-lime for measuring respiration rates in terrestrial ecosystems. *Pedobiologia*, **23**, 321–30.
- Edwards, N. T. and Sollins, P. (1973) Continuous measurements of carbon dioxide evolution from partitioned forest floor components. *Ecology*, **54**, 406–12.
- Fang, C. and Moncrieff, J. B. (1996) An improved dynamic chamber technique for measuring CO<sub>2</sub> efflux from the surface of soil. *Functional Ecology*, **10**, 297–305.
- Fang, C. and Moncrieff, J. B. (1998) An open-top chamber for measuring soil respiration and the influence of pressure difference on CO<sub>2</sub> efflux measurement. *Functional Ecology*, **12**, 319–25.
- Fang, C. and Moncrieff, J. B. (1999) A model for soil CO<sub>2</sub> production and transport I: Model development. *Agricultural and Forest Meteorology*, **95**, 225–36.
- Fernandez, I. J. and Kosian, P. A. (1987) Soil air carbon dioxide concentrations in a New England spruce-fir forest. *Soil Science Society of America Journal*, **51**, 261–3.
- Freijer, J. I. and Leffelaar, P. A. (1996) Adapted Fick's law applied to soil respiration. *Water Resources Research*, **32**, 791–800.
- Gao, F. and Yates, S. R. (1998) Laboratory study of closed and dynamic flux chambers: experimental results and implications for field application. *Journal of Geophysical Research*, **103**, 26115–25.
- Goulden, M. L. and Crill, P. M. (1997) Automated measurements of CO<sub>2</sub> exchange at the moss surface of a black spruce forest. *Tree Physiology*, **17**, 537–42.
- Grogan, P. (1998) CO<sub>2</sub> flux measurement using soda lime: correction for water formed during CO<sub>2</sub> adsorption. *Ecology*, **79**, 1467–8.
- Haber, W. (1958) Ökologische Untersuchung der Bodenatmung. Mit einer Übersicht über frühere Bearbeitungen, insbesondere deren Methoden. *Flora*, **146**, 109–57.
- Hanson, P. J., Edwards, N. T., Garten, C. T. and Andrews, J. A. (2000) Separating root and soil microbial contributions to soil respiration: a review of methods and observations. *Biogeochemistry*, **48**, 115–46.
- Healy, R. W., Striegl, R. G., Russell, T. F., Hutchinson, G. L. and Livingston, G. P. (1996) Numerical evaluation of static-chamber measurements of soil-atmosphere gas exchange: identification of physical processes. *Soil Science Society of America Journal*, **60**, 740–7.
- Iritz, Z., Lindroth, A. and Gärdenäs, A. (1997) Open ventilated chamber system for measurements of H<sub>2</sub>O and CO<sub>2</sub> fluxes from the soil surface. *Soil Technology*, **10**, 169–84.
- Janssens, I. A. and Ceulemans, R. (1998) Spatial variability in forest soil CO<sub>2</sub> efflux assessed with a calibrated soda lime technique. *Ecology Letters*, **1**, 95–8.
- Janssens, I. A., Kowalski, A. S., Longdoz B. and Ceulemans, R. (2000) Assessing forest soil CO<sub>2</sub> efflux: an in situ comparison of four techniques. *Tree Physiology*, **20**, 23–32.
- Janssens, I. A., Kowalski, A. S. and Ceulemans, R. (2001) Forest floor CO<sub>2</sub> fluxes estimated by eddy covariance and chamber-based model. *Agricultural and Forest Meteorology*, **106**, 61–9.
- Jassal, R. S., Black, T. A., Drewitt, G. B. *et al.* (2004) A model of the production and transport of CO<sub>2</sub> in soil: predicting soil CO<sub>2</sub> concentrations and CO<sub>2</sub> efflux from a forest floor. *Agricultural and Forest Meteorology*, **124**, 219–36.
- Kanemasu, E. T., Powers, W. L. and Sij, J. W. (1974) Field chamber measurements of CO<sub>2</sub> efflux from soil surface. *Soil Science*, **118**, 233–7.
- Kelliher, F. M., Lloyd, J., Arneth, A. *et al.* (1999) Carbon dioxide efflux density from the floor of a central Siberian pine forest. *Agricultural and Forest Meteorology*, **94**, 217–32.
- Koepf, H. (1953a) Die Verwendung eines Ultrarotabsorptionsschreibers (URAS) für die kontinuierliche Registrierung der Bodenatmung im Freiland. *Landwirtschaftliche Forschung*, **5**, 54–62.
- Koepf, H. (1953b) Die Temperatur/Zeit – Abhängigkeit der Bodenatmung. Experimentelle Untersuchung unter Verwendung des Ultrarotabsorptionsschreibers (URAS). *Zeitung für Pflanzenernährung, Düngung, Bodenkunde*, **61**, 29–47.
- Koepf, H. (1954) Die biologische Aktivität des Bodens und ihre experimentelle Kennzeichnung. *Zeitung für Pflanzenernährung, Düngung, Bodenkunde*, **64**, 138–46.
- Kutsch, W. L. (1996) *Untersuchung zur Bodenatmung zweier Ackerstandorte im Bereich der Bornhöveder Seenkette*. EcoSys. Beiträge zur Ökosystemforschung, Vol. Suppl. 16. Kiel: Verein zur Förderung der Ökosystemforschung.
- Kutsch, W. L. and Kappen, L. (1997) Aspects of carbon and nitrogen cycling in soils of the Bornhöved lake district. 2. Modelling the influence of temperature increase on soil respiration and organic carbon content in arable soils under different managements. *Biogeochemistry*, **39**, 207–24.



- Kutsch, W. L., Staack, A., Wötzel, J., Middelhoff, U. and Kapfen, L. (2001) Field measurements of root respiration and total soil respiration in an alder forest. *New Phytologist*, **150**, 157–68.
- Law, B. E., Baldocchi, D. D. and Anthoni, P. M. (1999) Below-canopy and soil CO<sub>2</sub> fluxes in a ponderosa pine forest. *Agricultural and Forest Meteorology*, **94**, 171–88.
- Livingston, G. P. and Hutchinson, G. L. (1995) Enclosure-based measurement of trace gas exchange: applications and sources of error. In *Methods in Ecology, Biogenic Trace Gases: Measuring Emissions from Soil and Water*, ed. P. A. Matson and R. C. Harris. Cambridge: Blackwell Science, pp. 14–50.
- Longdoz, B., Yernaux, M. and Aubinet, M. (2000) Soil CO<sub>2</sub> efflux measurements in a mixed forest: impact of chamber disturbances, spatial variability and seasonal evolution. *Global Change Biology*, **6**, 907–17.
- Lund, C. P., Riley, W. J., Pierce, L. L. and Field, C. B. (1999) The effects of chamber pressurization on soil-surface CO<sub>2</sub> flux and the implications for NEE measurements under elevated CO<sub>2</sub>. *Global Change Biology*, **5**, 269–81.
- Lundegårdh, H. (1922) Neue Apparate zur Analyse des Kohlensäuregehalts der Luft. *Biochem Z*, **131**, 109.
- Lundegårdh, H. (1924) *Der Kreislauf der Kohlensäure in der Natur. Ein Beitrag zur Pflanzenökologie und zur Landwirtschaftlichen Düngungslehre*. Jena: Verlag von Gustav Fischer.
- Lundegårdh, H. (1927) Carbon dioxide evolution of soil and crop growth. *Soil Science*, **23**, 417–53.
- Makkonen, K. and Helmisaari, H. S. (1999) Assessing fine-root biomass and production in a Scots pine stand: comparison of soil core and root ingrowth core methods. *Plant Soil*, **210**, 43–50.
- Mykeebust, M. C., Hippias, L. E. and Ryel, R. J. (2008) Comparison of eddy covariance, chamber, and gradient methods of measuring soil CO<sub>2</sub> efflux in an annual semi-arid grass, *Bromus tectorum*. *Agricultural and Forest Meteorology*, **148**, 1894–907.
- Nay, M. S., Mattson, K. G. and Bormann, B. T. (1994) Biases of chamber methods for measuring soil CO<sub>2</sub> efflux demonstrated with a laboratory apparatus. *Ecology*, **75**, 2460–3.
- Ngao, J., Longdoz, B., Perrin, D. *et al.* (2006) Cross-calibration functions for soil CO<sub>2</sub> efflux measurement systems. *Annals of Forest Science*, **63**, 477–84.
- Norman, J. M., Garcia, R. and Verma, S. B. (1992) Soil surface CO<sub>2</sub> fluxes and the carbon budget of a grassland. *Journal of Geophysical Research*, **97**, 18845–53.
- Norman, J. M., Kucharik, C. J., Gower, S. T. *et al.* (1997) A comparison of six methods for measuring soil-surface carbon dioxide fluxes. *Journal of Geophysical Research – Atmospheres*, **102**, 28771–7.
- Pihlatie, M., Pumpanen, J., Rinne, J. *et al.* (2007) Gas concentration driven fluxes of nitrous oxide and carbon dioxide in boreal forest soil. *Tellus Series B – Chemical and Physical Meteorology*, **59**, 458–69.
- Porkka, O. H. (1931) Über eine neue Methode zur Bestimmung der Bodenatmung. *Ann Soc Zool-Bot Fenn Yanamo*, **15**, 101–18.
- Pumpanen, J., Ilvesniemi, H., Keronen, P. *et al.* (2001) An open chamber system for measuring soil surface CO<sub>2</sub> efflux: analysis of error sources related to the chamber system. *Journal of Geophysical Research*, **106**, 7985–92.
- Pumpanen, J., Ilvesniemi, H., Peramaki, M. and Hari, P. (2003a) Seasonal patterns of soil CO<sub>2</sub> efflux and soil air CO<sub>2</sub> concentration in a Scots pine forest: comparison of two chamber techniques. *Global Change Biology*, **9**, 371–82.
- Pumpanen, J., Ilvesniemi, H. and Hari, P. (2003b) A process-based model for predicting soil carbon dioxide efflux and concentration. *Soil Science Society of America Journal*, **67**, 402–13.
- Pumpanen, J., Kolari, P., Ilvesniemi, H. *et al.* (2004) Comparison of different chamber techniques for measuring soil CO<sub>2</sub> efflux. *Agricultural and Forest Meteorology*, **123**, 159–76.
- Raich, J. W., Bowden, R. D. and Steudler, P. A. (1990) Comparison of two static chamber techniques for determining carbon dioxide efflux from forest soils. *Soil Science Society of America Journal*, **54**, 1754–7.
- Rayment, M. B. (2000) Closed chamber systems underestimate soil CO<sub>2</sub> efflux. *European Journal of Soil Science*, **51**, 107–10.
- Rayment, M. B. and Jarvis, P. G. (1997) An improved open chamber system for measuring soil CO<sub>2</sub> effluxes in the field. *Journal of Geophysical Research*, **102**, 28779–84.
- Rochette, P., Gregorich, E. G. and Desjardins, R. L. (1992) Comparison of static and dynamic closed chambers for measurement of soil respiration under field conditions. *Canadian Journal of Soil Science*, **72**, 605–9.

- Rochette, P., Ellert, B., Gregorich, E. G. *et al.* (1997) Description of a dynamic closed chamber for measuring soil respiration and its comparison with other techniques. *Canadian Journal of Soil Science*, **77**, 195–203.
- Savage, K. E. and Davidson, E. A. (2003) A comparison of manual and automated systems for soil CO<sub>2</sub> flux measurements: trade-offs between spatial and temporal resolution. *Journal of Experimental Botany*, **54**, 891–9.
- Schwartzkopf, S. H. (1978) An open chamber technique for the measurement of carbon dioxide evolution from soils. *Ecology*, **59**, 1062–8.
- Shibistova, O., Lloyd, J., Zrazhevskaya, G. *et al.* (2002) Annual ecosystem respiration budget for a *Pinus sylvestris* stand in central Siberia. *Tellus Series B*, **54**, 568–89.
- Striegl, R. G., McConnaughey, T. A., Thorstenson, D. C., Weeks, E. P. and Woodward, J. C. (1992) Consumption of atmospheric methane by desert soils. *Nature*, **337**, 145–7.
- Suarez, D. L. and Šimunek, J. (1993) Modeling of carbon dioxide transport and production in soil 2. Parameter selection, sensitivity analysis and comparison of model predictions to field data. *Water Resources Research*, **29**, 499–513.
- Take, E. S., Massman, W. J., Brandle, J. R. *et al.* (2004) Influence of high-frequency ambient pressure pumping on carbon dioxide efflux from soil. *Agricultural and Forest Meteorology*, **124**, 193–206.
- Tang, J., Misson, L., Gershenson, A., Cheng, W. and Goldstein, A. H. (2005) Continuous measurements of soil respiration with and without roots in a ponderosa pine plantation in the Sierra Nevada Mountains. *Agricultural and Forest Meteorology*, **132**, 212–27.
- Tang, J. W., Baldocchi, D. D., Qi, Y. and Xu, L. K. (2003) Assessing soil CO<sub>2</sub> efflux using continuous measurements of CO<sub>2</sub> profiles in soils with small solid-state sensors. *Agricultural and Forest Meteorology*, **118**, 207–20.
- Witkamp, M. (1969) Cycles of temperature and carbon dioxide evolution from litter and soil. *Ecology*, **50**, 922–4.
- Witkamp, M. and Frank, M. L. (1969) Evolution of CO<sub>2</sub> from litter, humus, and subsoil of a pine stand. *Pedobiologia*, **9**, 358–65.
- Xu, L., Furtaw, M. D., Madsen, R. A. *et al.* (2006) On maintaining pressure equilibrium between a soil CO<sub>2</sub> flux chamber and the ambient air. *Journal of Geophysical Research*, **111**, D08S10, doi:10.1029/2005JD006435.