

1 **Comparative analysis of Cd and Zn impacts on root distribution and**
2 **morphology of *Lolium perenne* and *Trifolium repens*: implications for**
3 **phytostabilization**

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26 Comments on the figures :

27 For figure 1 and 2, I would split the profiles for *Lolium* and *Trifolium* for better
28 readability

29

30

31 **Abstract**

32

33 *Backgrounds and Aims* Plant phytostabilization potential is a direct function of root
34 systems. An experimental design was developed to investigate the impact of Cd and
35 Zn on the root distribution and morphology of *Lolium perenne* and *Trifolium repens*.

36 *Methods* Seedlings were transplanted into columns filled with washed quartz and
37 irrigated daily with Cd- or Zn-containing nutrient solutions during one month. Roots
38 were then harvested at different depth and scanned to determine root length density
39 (RLD) and diameter. Pot experiments were also performed to quantify metal, lignin
40 and structural polysaccharides concentrations as well as cell viability.

41 *Results* *Lolium perenne* accumulated Cd and Zn in the roots while *T. repens* was
42 unable to restrict heavy metal translocation. Cadmium and Zn reduced rooting depth
43 and RLD but induced thick adventitious roots in *L. perenne*. This root swelling was
44 related to Cd-induced lignification occurring in exodermis and parenchyma of central
45 cylinder. Hemicellulose and lignin did not play a key role in root metal retention.
46 Cadmium slightly reduced mean root cell viability while Zn increased this parameter
47 in comparison to Cd.

48 *Conclusions* Even though plant species like *Lolium perenne* and *Trifolium repens*
49 may seem suitable for a phytostabilization strategy based on their shoot metal
50 tolerance, exposure to toxic heavy metals drastically impairs their root distribution

51 and could jeopardize the setting up of phytostabilization trials. The metal-induced
52 alterations of root properties differ clearly according to the considered pollutant and
53 plant species.

54

55 **Keywords**

56

57 Heavy metal contamination; root distribution; root diameter; lignin and structural
58 polysaccharides; *Lolium perenne*; *Trifolium repens*.

59

60 **1. Introduction**

61

62 Urbanization, industrialization and the use of heavy-metal containing
63 amendments in agriculture have resulted in soil contamination with heavy metals in
64 many areas of the world. Among the various remediation technologies applicable to
65 metal-polluted soils (Mulligan et al. 2001), phytostabilization is a low-cost strategy
66 aiming at limiting pollutant dispersion out of the contaminated area (Kidd et al. 2009;
67 Pilon-Smits 2005; Vangronsveld et al. 2009). This management strategy is suitable
68 for sites where the land value is low in comparison to the cost of remediation
69 (Robinson et al., 2009).

70 Various processes are involved in the control of heavy metal dispersion by
71 means of phytostabilization. Plants may decrease the metal bioavailability in the
72 polluted soil, which can be enhanced through the use of adapted soil amendments
73 (Houben et al. 2012, Kumpiene et al. 2008; Lambrechts et al. 2011; Mench et al.
74 2003). Plant cover may also reduce contaminant leaching (Houben et al. 2012), as
75 well as stabilize soils and control water and wind erosion (Reubens et al. 2007,

76 **Lambrechts et al., submitted**). Plants may help stabilize contaminants by adsorption
77 or accumulation in roots (Vangronsveld et al. 2009), but the translocation to the
78 shoots should remain limited to avoid contaminant transfer to the food chain (Kidd et
79 al. 2009). The selection of adequate plant species is therefore a fundamental aspect
80 for phytostabilization. The selected plants must also be perennial plant-species well-
81 adapted to the local environmental conditions, with a rapid installation, high biomass
82 production, and resistance to pollution (Pilon-Smits 2005). Root biomass and
83 distribution within the soil profile is especially crucial for the long-term maintenance of
84 the plant cover. Moreover, the entire root system architecture will contribute greatly to
85 mechanical soil stabilization and erosion control, thanks to the tensile strength and
86 friction or adhesion properties of single roots and to the morphological characteristics
87 of the root system (Mattia et al. 2005; Reubens et al. 2007).

88 Several pilot studies and landscape applications of phytostabilization have been
89 performed (e.g., Boisson et al. 2009; Dominguez et al. 2008), but more fundamental
90 research is still needed to better understand the interactions between heavy metals,
91 soil, plant roots and microorganisms in the rhizosphere (Vangronsveld et al. 2009). In
92 particular, the establishment of a plant cover could be jeopardized because of heavy
93 metal phytotoxicity. Chlorosis, necrosis and growth inhibition are common visible
94 symptoms of toxicity (Foy et al. 1978; Kabata-Pendias 2001). Metal impacts on
95 plants are generally ascribed to (1) blocking of the essential biological functional
96 groups of enzymes and/or modification of the active conformation of biomolecules
97 due to metal affinities for thioyl-, histidyl- and carboxyl-groups; (2) displacement of
98 essential cations from specific binding sites, leading to inhibition of enzyme activity;
99 (3) induced oxidative stress by increased reactive oxygen species (ROS) due to
100 perturbation of the mitochondrial and photosynthetic electron transfer chain and

101 inhibition of the antioxidant defense system; (4) competition with essential nutrients
102 during root uptake; and (5) mutagenic effect. This leads, for instance, to perturbation
103 of the water and nutrient status, and impairment of photosynthesis and selective
104 permeability of cell walls (Broadley et al. 2007; Clemens 2006; Kabata-Pendias 2001;
105 Sharma and Dietz 2009; Verbruggen et al. 2009).

106 Most studies dealing with heavy metal toxicity and phytostabilization have
107 focused on shoots, although studying roots is of crucial importance, as it is the
108 gateway for heavy metals uptake and the main accumulation compartment for most
109 plant species (Lux et al. 2011). Studying the effects of heavy metals on roots is
110 fraught with difficulties, however. Indeed, roots grow in an opaque medium from
111 which they cannot be extracted or observed without introducing artifacts (Lynch
112 1995). Moreover, it is almost impossible to avoid root contamination with polluted soil
113 particles when working with real soils. To overcome these difficulties, many scientists
114 have worked with hydroponic cultures or with agar medium (Zhu et al. 2011). These
115 methods are useful for analyzing the impact of heavy metals on root morphology and
116 physiological mechanisms. Heavy metals affect the root system by (1) growth
117 inhibition (Fusconi et al. 2007; Larbi et al. 2002); (2) alteration of the nutrient status
118 (Sandalio et al., 2001); (3) alteration of root anatomy and increased root diameter
119 (Lux et al. 2011); (4) ultrastructural modifications (Sresty and Rao 1999); (5)
120 modification of root architecture by induction of lateral roots (Ďurčeková et al. 2007);
121 and (6) accelerated maturation and lignification (Ederli et al. 2004; Schützendübel et
122 al. 2001). Cell walls, mainly composed of cellulose, hemicellulose and pectins may
123 constitute the main metal accumulation compartment in roots (Deiana et al. 2000;
124 Nishizono et al. 1987). Therefore, the metal-induced modification of their composition
125 may have an impact on both root metal retention and root morphological properties

126 (Lux et al., 2011; Zhu et al., 2012). However, when working with hydroponic or agar
127 medium conditions, nothing is known about the impact of heavy metals on tri-
128 dimensional root architecture (Zhu et al. 2011). Yet, this knowledge is crucial to
129 assess the phytostabilization potential of a plant species and a possible metal-
130 induced modification of the root distribution, which would affect the soil
131 phytostabilization potential.

132 In this paper, an experimental design was developed to assess together the
133 impact of heavy metals on root morphology and root architecture at different depths,
134 in order to assess the possible reduction in plant potential to stabilize soil because of
135 metal pollution. The two selected plant species, *Lolium perenne* (perennial ryegrass)
136 and *Trifolium repens* (white clover), exhibit drastically different root systems and were
137 successfully tested during phytostabilization pot experiments (Arienzo et al. 2004;
138 Lopareva-Pohu et al. 2011; Pichtel and Salt 1998; Santibáñez et al. 2008). The study
139 focused on the impact of Cd and Zn, two common metal pollutants in industrial
140 contaminated soils, on root distribution (biomass and length) and morphology (root
141 diameter) with depth. This analysis was carried out for different heavy metal
142 concentrations. The observed alterations of the root system were investigated by
143 physiological measurements regarding lignin and structural polysaccharide contents
144 and cell viability.

145

146 **2. Materials and methods**

147

148 2.1. Column experiment - root distribution with depth

149

150 The experiment was performed in a growth chamber under fully controlled
151 environmental conditions (16 h photoperiod, 24°C day, 22°C night, relative moisture
152 80%, light intensity 100 $\mu\text{mol.m}^{-2}.\text{s}^{-1}$). The two selected plant species were *Lolium*
153 *perenne* cv. Mondial and *Trifolium repens* cv. Alice. Two weeks after sowing in a pot
154 filled with loamy soil, plant seedlings, selected for uniformity, were transferred into
155 PVC columns (1 plant per column) of 55 cm height and 25 cm diameter filled with
156 washed quartz and were allowed to grow for 2 weeks with a daily irrigation of
157 Yoshida's nutrient solution. These columns were previously tested in order to
158 develop an experimental device which allowed study root system with depth without
159 disturbing root architecture setting up. The use of washed quartz enabled to easily
160 recover roots from the substrate and avoid root contamination by particles. Moreover,
161 washed quartz (particle size: 0.3-0.6 mm) exhibited a low compaction hazard in
162 comparison to others substrates. The washed quartz was not added directly into the
163 columns, but poured into a plastic bag inserted into the columns. The bottom of the
164 plastic bag was perforated to allow free drainage of excess water. Moreover, the
165 bottom of the columns was filled up with a 5-cm layer of coarse sand to facilitate
166 drainage and put on a perforated dish. Columns were irrigated each day with 2 l of
167 appropriate nutrient solution (see below). The irrigation volume was adapted in order
168 to add enough nutrients for plant growth and to avoid heterogeneity of the water
169 content with depth, which could strongly affect the root distribution. To avoid splash
170 effects during irrigation, a thin layer of gravel was added at the top of the washed
171 quartz. During preliminary tests, a gradual accumulation of nutrients and heavy
172 metals was detected in the columns. Therefore, pots were flushed weekly with
173 deionized water according to the recommendations of Zobel et al. (2007).

174

175 Five different treatments were performed for each plant species, with 3 replicates
176 per treatment. The control treatment consisted of Yoshida's nutrient solution
177 (Yoshida et al. 1976). All the columns received the Yoshida's solution for 2 weeks
178 after seedlings transplantation, and then the treatments were applied for 4 weeks.
179 For the other treatments, Cd or Zn was added to this nutrient solution to achieve the
180 following concentrations: for *Lolium perenne*, Zn 500 μ M, Zn 1000 μ M, Cd 25 μ M, Cd
181 50 μ M; for *Trifolium repens*, Zn 50 μ M, Zn 100 μ M, Cd 5 μ M, Cd 10 μ M. These
182 concentrations were selected based on previous hydroponic and pot tests, and
183 showed the lower tolerance of *Trifolium repens* to heavy metals in comparison to
184 *Lolium perenne* (Lambrechts et al. 2013).

185

186 After 4 weeks of plant growth, shoots were harvested, weighed and dried at 70°C
187 in an oven for 3 days. Shoot Cd and Zn concentrations were obtained by dissolving
188 50 mg of shoot dry matter in an open crucible with HNO₃ (AnalaR NORMAPUR
189 65%). The mixture was gently heated on a hot plate until complete dryness. The
190 residue was dissolved again with aqua regia (HCl AnalaR NORMAPUR 37% and
191 HNO₃ 65%) and filtered (Whatman 41). Finally, the solution was diluted to 10 ml with
192 deionized water and analyzed with an AA Spectrometer (Thermo Scientific S Series).
193 After shoot harvest, dishes at the bottom of the columns were replaced by a thick
194 PVC circle. The substrate was then gradually pushed up into the column by means of
195 a hydraulic jack, and 5-cm thick slices were cut with a blade-cutter. Roots were
196 separated from the washed quartz and rinsed with deionized water, then dried with
197 absorbent paper and fresh weighted. Roots were then stored into FAA solution
198 (ethanol 70% - acetic acid 100% - formaldehyde 35%, 18:1:1 by volume) for further
199 analyzes.

200

201 Root scanning was performed with a custom flat bed scanner (Medion 3600 DPI)
202 whose scanning window can be filled with water to enable an easier positioning of
203 the roots (Lobet and Draye 2013). Roots were spread out in the water and scanned
204 in positive transparency mode and 8-bit grey scale at 600 dpi. Image analysis was
205 performed with the open source software ImageJ (Abràmoff et al. 2004). A macro
206 was written to routinely determine root length density and root diameter. Briefly, for
207 every images the following steps were performed: (1) root object were retrieved using
208 a thresholding algorithm (Kapur et al. 1985), (2) total length was computed based on
209 the skeletonized image (Arganda-Carreras et al. 2010) and (3) root diameter
210 histogram was obtained by creating and euclidian distance map of the root object
211 and combining it with the skeleton (this manipulation yields a skeleton image in which
212 every pixel is equal to the root diameter at its position).

213

214 2.2. Pot experiments – physiological measurements

215

216 A pot experiment was performed in order to assess heavy metal concentrations in
217 shoots and roots. Pots were smaller than columns (13 cm diameter, 16 cm height).
218 All the experimental conditions were the same as for the columns, but the irrigation
219 volumes were adapted in order to preserve the same solution/pot volumetric ratio as
220 for the column experiment. All the treatments were replicated 5 times. After 4 weeks
221 of growth, the number of tillers was counted for *Lolium perenne*, as well as the
222 numbers of leaves and stolons for *Trifolium repens*. Shoots were harvested and fresh
223 weighed. Roots were separated from quartz, rinsed with deionized water, dried with
224 absorbent paper and fresh weighted. The number of nodules in *Trifolium repens* was

225 counted for each root system. Shoots and roots were oven dried at 70°C for 3 days,
226 dry-weighed, and the Cd and Zn concentrations were measured as previously
227 described. Translocation factors were calculated as the ratio between the metal
228 concentration in the shoots and concentration in roots.

229

230 A second pot experiment was performed next to deepen the analysis of metal
231 impact on roots. All the experimental conditions were identical as for the first pot
232 experiment, with 10 replications per treatment. After 4 weeks of treatment, roots of
233 half of the pots were harvested and put together per treatment, oven dried then
234 crushed with mortar and liquid nitrogen to get fine powder. Concentrations of lignin
235 and structural polysaccharides (cellulose and hemicellulose) were assessed
236 according to Van Soest et al. (1991). Briefly, the crushed plant sample was exposed
237 successively to a neutral detergent solution during 1h at 100°C to obtain by filtration
238 the NDF fraction (neutral detergent fibers), then to an acid detergent solution during
239 1h at 100°C to get the ADF fraction (acid detergent fibers), and then to sulfuric acid
240 72% during 3h to obtain the ADL fraction (acid detergent lignin). The ADL fraction
241 was incinerated at 550°C during 3h, and the mass loss allowed us to calculate the
242 percentage of lignin. The difference between ADF and ADL gave the percentage of
243 cellulose, and the difference between NDF and ADF the percentage of
244 hemicelluloses.

245

246 For *Lolium perenne*, some roots of the five remaining root systems were randomly
247 harvested for direct measurement of cell viability. The remaining parts of the root
248 systems of *Lolium perenne* and *Trifolium repens* were fixed into FAA solution for
249 further histochemical analyses. Specific measurements of cell viability were

250 performed on freshly harvested *Lolium perenne* apex roots, through the reduction of
251 2,3,5-triphenyltetrazolium chloride (TTC) into formazan (Lutts et al. 2004), with eight
252 replications per treatment. Approximately 50 mg of fresh tissue were excised, quickly
253 rinsed in deionized water containing 0.05% Tween 20 and incubated in darkness in
254 tubes containing 5 ml of 0.5% TTC in 50 mM K₂HPO₄ pH 7.0 for 15 h. Samples were
255 then filtered (Whatman n°4), rinsed with deionized water and incubated in 3 ml
256 ethanol 94% at 80°C during 5 min under gentle agitation (80 rpm). After
257 centrifugation at 5000 g during 1 min, extracted formazan was quantified
258 spectrophotometrically at 487 nm (Shimadzu UV Spectrophotometer UV-1800). The
259 viability index is defined as the absorbance measured per gram of fresh tissue.

260

261 For the root systems stored into FAA solution, cross-sections were performed on
262 random roots at 1 cm below the crown. Selected roots were dehydrated in a graded
263 ethanol series, embedded in paraffin and sectioned at 5 µm with a rotary microtome.
264 Serial cross-sections were stained with safranine-fast green and observed with a light
265 microscope Leica DM500 equipped with a camera Leica ICC50. Analysis of root
266 diameter and lignification percentage was assessed thanks to image analysis
267 software ImageJ. For the lignification percentage, a macro was developed to convert
268 the color image into a binary one based on a fixed threshold, which only selects
269 lignin-colored pixel, and to compare the binary area to the total root area.

270

271 2.3. Statistical analysis

272

273 Data of plant growth and development related parameters, metal concentrations,
274 lignin and structural polysaccharides relative contents, parameters related to cross-

275 sections and cell viability were statistically analysed through ANOVA 1, considering
276 five metal treatments for each plant species (Student-Newman-Keuls test; SAS
277 System for Windows, version 9.2). For relative contents of lignin, hemicelluloses and
278 cellulose, an arcsinus transformation was applied to obtain normal distribution of the
279 data. Statistical significance was assessed at the 5 % level, with a Welch correction
280 when there was no equality of the variance (Levene's test).

281

282 **3. Results**

283

284 3.1. Column experiment – modification of root distribution

285

286 The addition of heavy metals into the nutrient solution led to a decrease in the
287 shoot biomass of Cd 50 μ M and Zn 1000 μ M in comparison to control plants for *Lolium*
288 *perenne* (Table 1). Regarding *Trifolium repens*, heavy metal addition did not impact
289 the shoot dry weights in comparison to control plants. The concentrations of Cd and
290 Zn (Table 1) increased in shoots of *Lolium perenne* and *Trifolium repens* as a
291 function of the concentration in the nutrient solution. The differences of shoot
292 concentrations between the low pollutant dose and the highest one were statistically
293 significant for Cd, but not for Zn.

294

295 Fig. 1 presents the Root Weight Density (RWD), Root Length Density (RLD) and
296 root diameter depth profiles for *Lolium perenne*. The mean RWDs estimated for the
297 whole column were the following, in descending order: C 0.63^A g.dm⁻³, Cd25 0.38^B
298 g.dm⁻³, Zn500 0.20^{BC} g.dm⁻³, Cd50 0.09^C g.dm⁻³, Zn1000 0.06^C g.dm⁻³. Hence, Cd
299 and Zn significantly decreased the root biomass production by *Lolium perenne*. The

300 analysis of the RWD profile (Fig. 1A) showed that the decrease in root fresh weights
301 was dependent on root depth. Roots were detected until 40 cm depth for control
302 plants, but the maximum root depths decreased for the metal treatments: Cd25 35
303 cm, Cd50 30 cm, Zn500 25 cm, Zn1000 20 cm. These observations suggest that
304 heavy metals drastically impaired root elongation. Nevertheless, RWDs were also
305 severely affected in the top layer. At this level, the different treatments are ranked as
306 follows: C 0.36^A g.dm⁻³, Cd25 0.17^B , Zn500 0.16^B , Cd50 0.07^C , Zn1000 0.05^C .
307 Although Cd25 and Zn500 exhibited similar values in the top layer, the RWD
308 between the two treatments differed in the deeper layers: RWD for Zn500 decreased
309 strongly from the first to the second layer, whereas RWD for Cd25 only slightly
310 decreased and remained similar to control values as off 10 cm depth.

311 Mean RLD values for the whole column were the following, in descending order:
312 C 8.21^A m.dm⁻³, Cd25 4.05^B , Zn500 2.3^{BC} , Cd50 1.77^C , Zn1000 1.13^C . These values,
313 as well as Fig. 1B, show that RLD followed the same trend as RWD, except for Cd25.
314 Indeed, for this treatment, the RLD were lower than those of control plants until 20
315 cm depth, instead of 5 cm for the RWD. The analysis of root diameter (Fig. 1C)
316 showed that the presence of heavy metals induced an increase in the mean root
317 diameter in the top 5-cm layer in comparison to the control plants, which exhibited
318 relatively constant mean root diameters until 35 cm depth. This increase was
319 significant for Cd25, Cd50 and Zn500, but not for Zn1000. However, a higher mean
320 root diameter was only significant for the first 5-cm soil layer in the case of Cd50 and
321 Zn500 treatments, and the root diameters decreased strongly in the deeper layers in
322 comparison to control. This was not the case for Cd25: the mean root diameters were
323 higher than those of control plants until 35 cm depth.

324

325 Fig. 2 presents the RWD, RLD and root diameter depth profiles for *Trifolium*
326 *repens*. Mean RWD values for the whole column were, in descending order: C 0.37^A
327 g.dm⁻³, Zn100 0.26^{AB}, Zn50 0.18^B, Cd5 0.18^B, Cd10 0.18^B. Analysis of Fig. 2A
328 revealed that the trends observed for the whole column could also be detected in the
329 top soil layer, but not deeper. Rooting depth was not affected by the presence of Cd
330 and Zn in the nutrient solutions. The mean RLDs for the whole root systems were the
331 following, in descending order: C 3.23^A m.dm⁻³, Zn100 2.40^A, Zn50 2.40^A, Cd5 2.29^A,
332 Cd10 2.26^A. Therefore, Cd and Zn did not significantly affect the total root lengths. As
333 for RWD, differences between treatments were mainly observed in the top substrate
334 layer for Cd but not for Zn (Fig. 2B). Indeed, for this depth, there was a significant
335 difference between Cd treatments and the control. Some peaks of mean root
336 diameter (Fig. 2C) could be detected for the control and Cd treatments between 30
337 and 40cm depth, but this was not the case for the two Zn treatments. Roots of Zn100
338 exhibited a higher mean diameter in the first 15 cm, but below this depth the values
339 decreased faster than the other treatments.

340

341 3.2. Pot experiment – growth and metal concentrations

342

343 For the two considered plant species, shoot biomass was higher than root
344 biomass whatever the considered treatment (Fig. 3A & 3B). This biomass production
345 generally decreased when heavy metals were introduced into the device in
346 comparison with control plants. As far as *Lolium perenne* is concerned (Fig. 3A), root
347 biomass production was generally more affected than shoot ones by the addition of
348 Cd and Zn. Cadmium 50 µM induced a decrease in the shoot and root dry weights.
349 The addition of Zn into the nutrient solution did not significantly affect shoot and root

350 biomasses. For *Trifolium repens* (Fig. 3B), Cd 10 μ M significantly decreased both the
351 shoot and the root dry weights in comparison to the control. On the contrary, Zn did
352 not affect plant biomass, as for *Lolium perenne*. Heavy metals also affected plant
353 development, especially for the Cd treatments in *Trifolium repens* (Fig. 3D & 3E).
354 Indeed, addition of Cd significantly decreased both the number of leaves and stolons
355 in comparison to controls. Zn 50 μ M had no impact on plant development, and Zn
356 100 μ M decreased the number of stolons. Cd and Zn treatments had no impact on
357 the number of tillers for *Lolium perenne* (Fig. 3C). Heavy metals affected also the
358 occurrence of root nodules for *Trifolium repens* (Fig. 3F), especially for Cd 10 μ M.

359

360 The increase in Cd concentration in contaminated nutrient solution led to
361 significantly higher concentrations in both shoot and root tissues of the two plant
362 species (Table 2). A similar trend was observed for Zn, but the increase between Zn
363 500 μ M and Zn 1000 μ M was not significant for *Lolium perenne* roots. Calculation of
364 translocation factors indicated that zinc treatments led to lower values of Zn
365 translocation in comparison to control for *Lolium perenne*, but not for *Trifolium*
366 *repens*. Moreover, translocation factors were lower for *Lolium perenne* than for
367 *Trifolium repens*.

368

369 3.3. Pot experiment – lignin and structural polysaccharides

370

371 The analysis of the relative content of lignin, hemicellulose and cellulose were
372 performed for the whole root systems of the two plant species (Fig. 4). Structural
373 polysaccharides and lignin content differed strongly between the controls of the two
374 plant species, especially for cellulose and hemicelluloses. Indeed, their relative

375 amounts were clearly higher in *Lolium perenne* in comparison to *Trifolium repens*.
376 Zinc excess decreased cellulose, hemicelluloses and lignin content in *Lolium*
377 *perenne* roots. These modifications were proportional to the exogenous Zn
378 concentration. Cadmium also decreased the root cellulose and hemicellulose
379 content, but lignin remained unaffected. Zn50 and Cd5 treatments did not lead to
380 detectable changes in the lignin and structural polysaccharides contents in *Trifolium*
381 *repens* in comparison to controls. However, all fractions markedly increased in
382 response to Zn100 and Cd10.

383

384 3.4. Pot experiment – cross-sections and cell viability

385

386 This complementary experiment focused mainly on *Lolium perenne* roots, which
387 exhibited clearer response to heavy metals than *Trifolium repens*. Fig. 5 shows the
388 results of the analysis of cross-sections performed on randomly selected roots 1 cm
389 below the crown. A significant increase in the root diameter for the metal treatments
390 in comparison to control was observed for *Lolium perenne* (Fig. 5A), confirming the
391 increase in mean root diameter previously recorded in the top layer of the column
392 experiment for the contaminated treatments (Fig. 1C). A visual comparison of the
393 root systems for the different treatments (Fig. 5) allowed complementing the previous
394 observations. The plant crown of Zn500, Zn1000 and Cd50 showed a proliferation of
395 short and thick adventitious roots, which was not observed for Cd25 and control
396 plants. The assessment of root lignification for these roots thanks to safranin
397 staining and image analysis with ImageJ revealed two different trends for Cd and Zn
398 (Fig. 5B). Contamination with Cd led to a local increase in root lignification compared
399 to control plants, whereas Zn decreased the values of this parameter. This Cd-

400 induced increase in lignification was observed mainly for the exodermis as well as for
401 the parenchyma cells in the central cylinder.

402

403 The Cd and Zn induced short adventitious roots were compared to apex roots of
404 control plants in terms of cell viability for *Lolium perenne* (Fig. 6). Zn1000-
405 contaminated root cells led to higher cell viability than the control and Cd-treated
406 roots.

407

408 **4. Discussion**

409

410 4.1. Contrasted plant species response to heavy metal exposure

411

412 *Lolium perenne* and *Trifolium repens* exhibited drastically different responses to
413 Cd and Zn (Table 2). Despite lower Cd and Zn concentrations in contaminated
414 nutrient solutions for *Trifolium repens* in comparison to *Lolium perenne* (5 and 10
415 times less for Cd and Zn, respectively), the accumulation of those two elements in
416 the shoots were within the same range for the two species. Given the higher metal
417 concentrations in root tissues of *Lolium perenne* (3 times more than for *Trifolium*
418 *repens*), it appears that perennial ryegrass is able to accumulate Cd and Zn mainly in
419 the roots and limit their translocation to the shoots. Therefore, shoot growth and
420 development were not significantly affected by metal contamination in this species
421 (Fig. 3). On the contrary, *Trifolium repens* translocated more metals to the shoots,
422 which impaired shoot development. Yang et al. (1995) observed the same patterns
423 for white clover and perennial ryegrass with Cd hydroponic experiments. Moreover,
424 for the same endogenous metal concentration as used here for *Lolium perenne*,

425 shoot growth and development of white clover were more affected than ryegrass by
426 metal contamination, especially for Cd. This may be explained by more efficient
427 mechanisms allowing *Lolium perenne* to cope with metal accumulation in shoots
428 tissues. Macnicol and Beckett (1985) collected data from many hydroponic and soil
429 experiments to determine critical tissue concentrations of heavy metals for different
430 plant species, beyond which toxic effects are detected. They indicated critical Cd
431 shoot concentrations of 10 mg.kg⁻¹ for white clover and 30-35 mg.kg⁻¹ for ryegrass,
432 and for Zn 250-300 mg.kg⁻¹ for clover and 370-560 mg.kg⁻¹ for ryegrass. These
433 values confirm the lower level of metal tolerance of *Trifolium repens* comparatively to
434 *Lolium perenne*. The lower biomass production of white clover could also be
435 explained by the Cd-induced inhibition of symbiotic nitrogen fixation (Broos et al.
436 2004), as suggested in the present study with a reduction of the occurrence of root
437 nodules. Bidar et al. (2007, 2009) reported a higher sensitivity to metal-induced
438 oxidative stress in *Lolium perenne* than in *Trifolium repens*. This suggests that for a
439 given plant species, plant behavior could vary depending on the considered cultivar
440 (Grant et al. 2008) or that oxidative stress is not the major cause of growth inhibition,
441 at least in *Lolium perenne*.

442

443 4.2. Impact of heavy metals on the root system distribution and morphology

444

445 Pot experiment suggested that *Lolium perenne* was suitable for phytostabilization
446 purpose due to its low metal translocation from roots to shoots and its shoot
447 tolerance to heavy metals (Table 2; Fig. 3). Indeed, ryegrass shoots were only
448 significantly affected by exposure to Cd 50µM in comparison to control and could
449 provide an important plant cover to limit pollutant dispersion by water and wind

450 erosion. However, Cd and Zn decreased both the rooting depth and the RLD for this
451 plant species (Fig. 1). Therefore, soil exploration by roots may progressively be
452 limited while increasing heavy metal concentrations, which in turn could lead to
453 nutrient deficiencies and thus affect the durability of the plant cover. Moreover, this
454 induced decrease in the total root length, especially in the top layer of soils, could
455 reduce plant anchorage and the roots' potential to stabilize soils (Reubens et al.
456 2007). Concerning *Trifolium repens*, Cd treatments led to a significant decrease of
457 biomass and RLD in the top soil layers (Fig. 2). Therefore, as for ryegrass, metal
458 exposure also restricts the potential of white clover for phytostabilization. Some metal
459 hyperaccumulator plant species, such as *Thlaspi caerulescens* or *Sedum alfredii*, are
460 able to maintain or increase their root distribution under metal contaminated
461 conditions (Li et al. 2009; Schwartz et al. 1999). However, these plant species always
462 display a low biomass and a shallow root system and are therefore not
463 recommended for phytostabilization (Mench et al. 2010). These observations
464 suggested the requirement, during the development of a phytostabilization strategy,
465 to combine plant installation with the use of adapted amendments to overcome this
466 metal-induced plant growth inhibition (Kumpiene et al. 2008; Lambrechts et al. 2011;
467 Mench et al. 2010).

468

469 The reduction of root depth and RLD in *Lolium perenne* (Fig. 1) was linked to a
470 Cd- and Zn-induced proliferation of short and thick adventitious roots (Fig. 5) which
471 contributed greatly to the observed increase of mean root diameter in the top soil
472 layer and to the drastic decrease of root depth. The reduction of root growth may
473 result from cell death, inhibition of cell division, a decrease in the rate of cell
474 elongation or different processes occurring simultaneously (Delpérée and Lutts 2008;

475 Sresty and Rao 1999). In our work, Zn treatments induced an increase in cell viability
476 in comparison to the control (Fig. 6). Moreover, cell viability was not affected in
477 response to Cd treatments, as observed also by Delpérée and Lutts (2008) with
478 *Solanum lycopersicum* after contamination with Cd 250µM during 14 days. Fusconi
479 et al. (2007) observed with *Pisum sativum* a decrease of cell viability only for Cd
480 250µM, but not for the lower Cd concentrations. They attributed the reduction of plant
481 growth to a decrease of apex length in relation to the inhibition of mitotic activity.

482

483 However, some authors suggested that root growth inhibition could be rather
484 attributed to metal-induced inhibition of cell elongation occurring through several
485 mechanisms. Our work demonstrates that such inhibition was linked with root
486 swelling (Fig. 5). Cytoskeleton, and especially microtubules, constitutes a major
487 target of heavy metals. Přibyl et al. (2005) showed that heavy metals could induce a
488 disintegration of microtubules followed by a steep increase in cell width and a
489 decrease in growth rate for green alga *Spirogyra decimina* exposed to Cd. Heavy
490 metals could also damage cell membrane, as shown by Bidar et al. (2008) for *Lolium*
491 *perenne* and *Trifolium repens*. This loss of membrane integrity would result in turgor
492 loss and inhibition of cell expansion (Poschenrieder and Barceló 2004). However,
493 heavy metals at lower concentrations could also act on cell elongation by decreasing
494 cell wall extensibility without damaging cell membrane. Exposure of *Pinus sylvestris*
495 roots to Cd concentrations exceeding cell detoxification capacity led to accumulation
496 of H₂O₂ because of an imbalance of the redox systems (Schützendübel et al. 2001).
497 H₂O₂ is a signaling intermediate which could increase the mechanical strength and
498 lower the extensibility of plant cell walls, following an abnormal lignification
499 (Ďurčková et al. 2007; Ederli et al. 2004; Schützendübel et al. 2001). This

500 premature xylogenesis shortened the root elongation zone and therefore reduced
501 root growth (Ďurčková et al. 2007, Lunáčková et al. 2003).

502

503 Cd-induced lignification was detected in the present study based on the cross-
504 sections of ryegrass roots (Fig. 5). We demonstrate that lignification, linked to root
505 swelling, was localized in *Lolium perenne* in the exodermis and in the parenchyma
506 cells in the central cylinder. Lux et al. (2011) showed that root tissues which
507 increased their size after metal exposure varied according to plant species and Cd
508 concentration. However, the lignin contents were not significantly different between
509 control and Cd-treated whole root systems (Fig. 4). This suggests that the short and
510 thick adventitious roots represented only a small part of the whole root biomass,
511 which included also longer and finer roots formed before the exposure to metal
512 toxicity, for which only their apices were exposed to Cd-induced lignification.

513

514 Several cell wall compounds may adsorb heavy metals (Chen et al. 2013) and
515 Cd-induced lignification is accordingly sometimes reported as a plant adaptation
516 mechanism to tolerate heavy metals (Ederli et al. 2004, Maksimović et al. 2007)..
517 Metals can sorb to carboxylic (hemicelluloses, pectin and lignin), phenolic (lignin,
518 waxes, fat), hydroxylic (structural polysaccharides, lignin, pectin) and carbonyl
519 groups (lignin) (Pejic et al. 2009). In contrast, cellulose could not efficiently bind metal
520 ions efficiently due to its unbranched structure (Zhu et al. 2012). Although
521 hemicelluloses, pectins and lignin may bind heavy metal compounds, their sorbing
522 efficiency depends strongly of the considered plant species and metal types. Indeed,
523 Hu et al. (2010) obtained the highest metal binding potential for hemicelluloses from
524 rice bran, as well as Nawirska (2005) for pectin and hemicelluloses in pomace fibers.

525 However, Chen et al. (2013) showed that hemicelluloses negatively impact Cd
526 adsorption in willow roots, while pectins enhanced Cd adsorption. In the present
527 study, hemicellulose and lignin relative contents increased in *Trifolium repens* after
528 exposure to heavy metals (Fig. 4), although metal translocation factors remained
529 unaffected (Table 2). An opposite trend was recorded for *Lolium perenne*: metal
530 translocation to the shoots decreased while relative contents of hemicelluloses and
531 lignin also decreased after Cd and Zn contamination. This suggests that
532 hemicelluloses and lignin did not play a key role in metal root accumulation in the
533 tested species. However, pectins, which were not quantified in the present work, may
534 contribute to heavy metal retention in roots, as well as other mechanisms also taken
535 into account in metal homeostasis, such as chelation, vacuole sequestration, etc.
536 (Clemens 2006; Verbruggen et al. 2009).

537

538 Zn 1000 μM induced interesting and drastically different results than Cd
539 treatments for ryegrass roots. Indeed, root lignification decreased in comparison to
540 other treatments (Fig. 5), and such a trend was also observed for the lignin content at
541 the whole root system level (Fig. 4). Therefore, root swelling could not be linked to
542 root lignification in this case. The higher cell viability detected for Zn 1000 μM
543 indicated also that other mechanisms act on root swelling (Fig. 6). However, to our
544 knowledge, scientific data are missing about the Zn-induced root swelling
545 mechanisms. Cell viability measurements showed that root elongation inhibition and
546 swelling were not linked to cell death. Sresty and Rao (1999) proposed that Zn
547 affected mainly cell elongation instead of cell division. The mechanism may be a
548 disturbance of cytoskeleton, leading to an increase in cell width (Přibyl et al. 2005).

549 However, the data in the present study were not sufficient to support such
550 hypothesis.

551

552 4.3. Comparison of the experimental devices

553

554 When designing a pot experiment, experimental factors such as pot size, rooting
555 medium, nutrient content, *etc.* should be cautiously selected (Poorter et al. 2012).
556 Therefore the pots size was adapted in the present study based on the scientific
557 objectives: column size allowed not to limit root system setting up and was suitable
558 for root distribution study. However, such experimental devices are costly in terms of
559 space occupation, maintenance and harvest times, thus limiting the feasible number
560 of treatments and replicates. Therefore, smaller pots containing the same rooting
561 medium and metal concentration are often used to investigate metal-induced
562 modifications of root properties. Although the same ratio volume pot / volume
563 irrigation was considered for pots and columns for daily irrigation, biomass production
564 was higher in pots than in columns, especially for *Trifolium repens* (Table 1; Fig. 3).
565 During preliminary experiments, an experimental duration of 8 weeks instead of 4
566 weeks was tested, among other things, and higher biomass production was observed
567 for plants into columns in comparison to pots (results not shown), because of a
568 limitation of root development after this duration. The difference in biomass
569 production after 4 weeks of treatment may be linked to different water content
570 gradient between the two containers, which is a typical and often-unrecognized
571 artifact with pot experiments (Passioura 2006). Because of the rapid drainage of
572 nutrient solution irrigated each morning, water content was low at the top of the
573 column device, and increased gradually with depth. This water content gradient, as

574 well as nutrient content gradient, were detected during preliminary tests. The addition
575 of coarse material at the bottom of the columns to facilitate drainage and the weekly
576 flushes with deionized water should partly have alleviated this artifact. However, the
577 residual water content gradient may have affected the root distribution with depth of
578 *Trifolium repens*, as observed with small peaks for RWD, RLD and root diameter
579 (Figure 4) at the bottom of the device. On the contrary, for the pot experiment, root
580 medium was more saturated with water despite the presence of coarse particles at
581 the bottom of the containers. Therefore, the higher duration of contact between the
582 roots and the nutrient solutions for the pot experiment may lead to an increase in
583 nutrient and metal uptake. Indeed, Cd and Zn mineralomasses were higher in shoots
584 grown in pots. It may be assumed that the metal exposure of root systems led to the
585 same impacts for the two experimental devices, but that these effects may be
586 exacerbated with the pot experiment, despite the same concentrations in the
587 contaminated nutrient solutions and the same irrigation ratios.

588

589 **Conclusions**

590

591 Even though plant species like *Lolium perenne* and *Trifolium repens* may
592 seem suitable for a phytostabilization strategy based on their shoot metal tolerance,
593 exposure to toxic heavy metals drastically impairs their root distribution within the
594 substrate profile, leading to a reduction of their soil stabilization potential.
595 Assessment of plant potential for phytostabilization purpose should therefore be
596 based on both shoot and root behaviors against toxic metal exposure. This metal-
597 induced reduction of root distribution was linked for *Lolium perenne* with a
598 proliferation of thick and short adventitious roots. However, the mechanisms leading

599 to such a root swelling differed between Cd and Zn. The increase in root diameter
600 was linked to Cd-induced root lignification, localized in exodermis and parenchyma of
601 central cylinder. Exposure to Zn increased the root diameter without increasing root
602 lignification, and higher cell viability was detected. Moreover, the composition of root
603 cell walls was differently affected by metal exposure according to the considered
604 plant species, with an increase in lignin, hemicelluloses and cellulose contents for
605 *Trifolium repens*. These cell wall modifications did not affect the root metal retention.
606 This study demonstrates that impacts of heavy metal toxicity on root properties differ
607 according to the considered pollutant and plant species.

608

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610

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614

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