Atmospheric CO₂ flux from mangrove surrounding waters

A. V. Borges, S. Djenidi, G. Lacroix, J. Théate, B. Delille, and M. Frankignoulle Université de Liège, MARE, Liège, Belgium

Received 17 February 2003; accepted 6 May 2003; published 4 June 2003.

[1] The partial pressure of CO_2 (p CO_2) was measured at daily and weekly time scales in the waters surrounding mangrove forests in Papua New Guinea, the Bahamas and India. The pCO₂ values range from 380 to 4800 µatm. These data, together with previously published data, suggest that overall oversaturation of CO₂ with respect to atmospheric equilibrium in surface waters is a general feature of mangrove forests, though the entire ecosystems (sediment, water and vegetation) are probably sinks for atmospheric CO₂. The computed CO₂ fluxes converge to about +50 mmolC m⁻² day^{-1} . If this conservative value is extrapolated for worldwide mangrove ecosystems, the global emission of CO_2 to the atmosphere is about 50 10⁶ tC year⁻¹. Based on this tentative estimate, mangrove waters appear to be regionally a significant source of CO₂ to the atmosphere and should be more thoroughly investigated, especially at seasonal time scale. INDEX TERMS: 4806 Oceanography: Biological and Chemical: Carbon cycling; 4805 Oceanography: Biological and Chemical: Biogeochemical cycles (1615); 4820 Oceanography: Biological and Chemical: Gases; 4231 Oceanography: General: Equatorial oceanography; 4235 Oceanography: General: Estuarine processes. Citation: Borges, A. V., S. Djenidi, G. Lacroix, J. Théate, B. Delille, and M. Frankignoulle, Atmospheric CO₂ flux from mangrove surrounding waters, *Geophys.* Res. Lett., 30(11), 1558, doi:10.1029/2003GL017143, 2003.

1. Introduction

[2] Mangroves are among the most productive coastal intertidal ecosystems in the world, confined to the tropics and subtropics, that dominate approximately 75% of the world's coastline between 25°N and 25°S, and are estimated to cover between 0.17 and 0.20 10⁶ km². The mangrove ecosystem as a whole (aquatic, below- and above-ground compartments) is net autotrophic, but the water column and the sediment are largely net heterotrophic, due to three processes: (1) aquatic primary production is limited by high turbidity, canopy shadow and large changes in salinity; (2) water column and sediments receive important quantities of leaf and wood litter from the overlying canopy; (3) export of labile organic carbon from mangroves to adjacent aquatic systems, although variable from one site to another, can be low [see Jennerjahn and Ittekkot, 2002]. Consequently, the mangrove surrounding waters should act as a net source of CO_2 to the atmosphere, although no attempt has been made so far to estimate its magnitude.

2. Sites and Methods

[3] Nagada Creek (5°09'S 145°48'E) is a small tidal creek (3 km long, 400 m wide) in the northern Papua New Guinea

Copyright 2003 by the American Geophysical Union. 0094-8276/03/2003GL017143\$05.00

coast, that opens to Nagada Lagoon. The latter is a wide lagoon (maximum width 5 km) separated from the Pacific Ocean by a continuously submerged barrier reef. Nagada Creek is fringed by well developed mangroves while seagrass beds and submerged reefs are mainly present at the mouth of the creek. Gaderu Creek $(16^{\circ}50'N 82^{\circ}19'E)$ is the main tidal creek connecting the Godavari estuary (Anhra Pradesh, east coast of India) to the semi-enclosed shallow Kakinada Bay. It has a total length of 11 km (about 30 m wide and 4 m deep) and consists of intertidal mudflats, most of which are covered by mangrove forest and no seagrass beds or coral reefs are found nearby. Norman's Pond Cay (23°46'N 76°07'W) is one of the many Exuma Cays islands (Bahamas archipelago) which form the boundary between the shallow Great Bahama Bank and the deep Exuma Sound. Norman's Pond (about 500 m long and wide, 2 m deep), is a brackish to hypersaline pond where a dense mangrove is present. It is located in the southern part of the Norman's Pond Cay and connected to the leeward shore of the island by a tidal channel.

[4] In Papua New Guinea (July-August 2000), pH, Total Alkalinity (TAlk), salinity and water temperature were sampled every second day around 10:00 (local time), during 2 weeks, in a small (400 m long, 100 m wide) and shallow (3 m deep) cove surrounded by dense mangrove in the upper part of Nagada Creek (5°9'N 145°47'E). About 1 h later, a second station (5°9'N 145°48'E) located in Nagada Lagoon, near the mouth of Nagada Creek, was sampled. In the Bahamas (December 2000), we used a floating equilibrator system (FES) to carry out three 24 h cycles of pCO₂, water temperature and wind speed measurements, with an one minute sampling interval. Two daily cycles were carried out in Norman's Pond (23°46'N 76°07'W) and one at Block Rock (23°49'N 76°10'W) in Bock Cay. The FES consists of an equilibrator mounted on a buoy, including batteries, a solar panel, air and water temperature probes, an anemometer and a data logger. A non-dispersive infrared gas analyser (Li-Cor, Li-6262) was used to measure pCO₂ by equilibration, with an estimated accuracy of ± 5 µatm. In India (June 2001), we carried out a 24 h cycle in Gaderu Creek (16°47'N 82°18'E) with measurements of pH, TAlk, salinity, temperature and dissolved oxygen sampled every hour. The same variables were also sampled during a survey of Kakinada Bay (28 May 2001) and along a transect in the Godavari estuary (1 June 2001). In Papua New Guinea and India, wind speed measurements were made with a handheld anemometer and all dissolved variables, except TAlk, were measured on site with portable instruments. For a detailed description of the pH and TAlk measurement methods, the computations of pCO_2 and DIC (Dissolved Inorganic Carbon) from pH and TAlk and the calibration procedure of the Li-6262 refer to Frankignoulle and Borges [2001].



Figure 1. Variations of pCO₂ (full circles), *n*TAlk (open squares) and *n*DIC (full squares) in July–August 2000, in Nagada Creek (Papua New Guinea). Horizontal dotted line corresponds to atmospheric equilibrium.

3. Results and Discussion

3.1. pCO₂, DIC and TAlk Dynamics

[5] In Nagada Creek (Papua New Guinea), pCO₂ values vary from 540 to 1680 µatm, during the 12 days sampling period (Figure 1A). Normalized TAlk (nTAlk) and normalized DIC (nDIC) follow closely the pCO₂ evolution, except for the sample from August 1 (Figure 1B). This strongly suggests that the same process increases the three variables. Furthermore, this process is specific to the mangrove because the average pCO₂, nTAlk and nDIC values are significantly higher than those in the adjacent Nagada Lagoon (Table 1). In Gaderu Creek (India), the pCO₂ variations are striking during the 24 h cycle, with values ranging between 1380 and 4770 µatm (Figure 2A). As in Nagada Creek, *n*TAlk and *n*DIC follow the trend of pCO₂ (Figure 2B). Also, the average pCO₂, nTAlk and nDIC values are much higher than those in the adjacent Godavari estuary and those in Kakinada Bay (Table 1). On first analysis, pCO₂, *n*DIC and Oxygen saturation level (%O₂) exhibit the patterns expected from the daily cycle of primary production and respiration (decrease of pCO_2 and nDICuntil sunset, increase of pCO_2 and *n*DIC during the night until dawn, and %O₂ showing the inverse trend). However, there are two inconsistencies with this interpretation. First, pCO₂ and *n*DIC strongly decrease (and %O₂ increases) one hour before sunrise. Second, pCO₂ and nDIC values remain almost constant from 22:00 to 05:00, instead of increasing regularly throughout the night as expected from respiration (% O_2 evolution shows the same trend). A closer look at Figure 2 reveals that the lowest pCO₂, nTAlk and nDIC values correspond to high tide and the highest values are observed at low tide (%O₂ shows the inverse trend), and, coincidently high tide occurred at dusk.

[6] To explain the daily variations of various chemical parameters during three 24 h cycles in the creek waters of a

Table 1. Average \pm Standard Deviation of pCO₂ (µatm), *n*DIC (mmol kg⁻¹) and *n*TAlk (mmol kg⁻¹) at the Different Sites in Papua New Guinea (top half of table) and India (bottom half of table)

	pCO ₂	nDIC	nTAlk	
Nagada Creek (6)	1180 ± 360	2.6 ± 0.79	2.9 ± 0.7	
Nagada Lagoon (6)	585 ± 200	1.8 ± 0.1	2.1 ± 0.1	
Gaderu Creek (25)	2565 ± 868	2.5 ± 0.1	2.6 ± 0.1	
Godavari estuary $(2)^{a}$	465 ± 20	1.9 ± 0.2	2.2 ± 0.2	
Kakinada Bay (9)	500 ± 125	1.7 ± 0.1	2.0 ± 0.1	

DIC and TAlk were normalized to a constant salinity close to the mean of the sampled values: 34 in Papua New Guinea and 30 in Gaderu Creek. Numbers in brackets are the number of samples.

 a For the same salinity range (28.4–31.6) as the one sampled at Gaderu Creek.



Figure 2. Variations of pCO₂ (full circles), $%O_2$ (open circles), *n*TAlk (open squares) and *n*DIC (full squares) during a 24 h cycle in Gaderu Creek (India, 4–5 June 2001). High tide (HT) and low tide (LT) are given in cm, relative to mean sea level, based on tide tables for Kakinada harbour. The shaded area corresponds to night-time. $%O_2$ is computed from observed O₂ concentration (O_{2observed}) and O₂ concentration at saturation (O_{2saturation}) according to: $%O_2 = O_{2observed}/O_{2saturation}$.

Brazilian mangrove, Ovalle et al. [1990] suggested the following mechanism: during the ebb and at low tide, there is a strong influx of porewater that mixes with the creek water, substantially affecting the chemical properties of the latter. During the flow, the migration of porewater towards the creek strongly decreases until it stops when the sediment surface is inundated at high tide. The time course of *n*TAlk and *n*DIC in Gaderu Creek fits well with this hypothesis, as it is a well established fact that mangrove porewaters have a high TAlk and DIC content [e.g. Ovalle et al., 1990; Middelburg et al., 1996; Alongi et al., 2001; Kristensen et al., 2000]. Ovalle et al. [1990] explain the high porewater TAlk by sulfate reduction that along with aerobic respiration, account for almost all the diagenetic carbon degradation in mangroves. Generally, sulfate reduction is the major diagenetic pathway in mangroves [e.g. Alongi et al., 1998], but in some cases aerobic degradation predominates [Alongi et al., 2001], and in one Thai mangrove, iron reduction was reported as the dominant process [Kristensen et al., 2000].

[7] The relative variation of TAlk and DIC follows a well established stochiometry that is specific to the biogeochemical process controlling these variables. In Nagada Creek, *n*TAlk and *n*DIC are well correlated and the slope of the



Figure 3. *n*TAlk *versus n*DIC in Nagada Creek (Papua New Guinea) and Gaderu Creek (India). Solid lines correspond to the theoretical evolution to attain the highest *n*TAlk and *n*DIC values for the potential biogeochemical processes that can control these variables: areobic respiration (areobic resp.), sulfate reduction (SO₄ red.), calcium carbonate dissolution (CaCO₃ dissol.), manganese reduction (Mn red.) and iron reduction (Fe red.). Numbers in brackets correspond to the theoretical slope of the relative variation of *n*TAlk and *n*DIC for each of the biogeochemical processes.



Figure 4. pCO_2 variations during two 24 h cycles in Norman's Pond (Bahamas) carried out on 10–11 December 2000 (A) and 14–15 December 2000 (B). Horizontal dotted line corresponds to atmospheric equilibrium. Measurements of high tide (HT) and low tide (LT) are given in cm, relative to mean sea level and were obtained nearby but outside the pond. The shaded area corresponds to night-time.

regression line $(0.99 \pm 0.04, r^2 = 0.994, N = 6)$ is very close to the one predicted if sulfate reduction was the main biogeochemical process controlling these variables, and assuming a net accumulation of FeS in the sediments (Figure 3A). In Gaderu Creek, the slope of the regression line of *n*TAlk and *n*DIC is also highly significant $(0.61 \pm 0.03, r^2 = 0.945, N = 25)$ but lower than the one predicted assuming that sulfate reduction is the only biogeochemical process controlling these variables (Figure 3B). This suggests that in this system areobic respiration also strongly contributes to the TAlk and DIC dynamics. Another mechanism that could explain such a pattern is the oxidation in the water column of H₂S released for the sediments. This would decrease *n*TAlk but leave *n*DIC unaffected.

[8] The input to the creek of porewater with a high DIC and TAlk content can explain the weekly variability of pCO₂, *n*DIC and *n*TAlk in Nagada Creek (Figure 1). Indeed, the highest values were observed either at low tide or during the ebb, and the lowest values (August 5) were obtained at high tide. The corresponding values of pCO₂, *n*TAlk and *n*DIC, respectively, 538 μ atm, 2.16 and 1.86 mmol kg⁻¹, are close to the average values of Nagada Lagoon waters (Table 1).

[9] In Norman's Pond, the range of pCO_2 variations was 385–750 µatm during the first 24 h cycle and 395–690 µatm during the second one (Figure 4). The daily change of pCO_2 during both 24 h cycles follows the general pattern expected from the diel alternation between photosynthesis and respiration. No significant pCO_2 signature is associated to low or high tides. This could be due to the fact that, in this semi-enclosed system, the tidal amplitude is smaller than outside the pond and the input of porewater is therefore

less marked and less dependent on tidal inundation as described for Nagada Creek and Gaderu Creek. The average pCO₂ values during both 24 h cycles in Norman's Pond (516 ± 83 (sd, standard deviation, N = 1521) µatm and 531 ± 89 (sd, N = 1403) µatm) are significantly higher than the average pCO₂ value obtained during the 24 h cycle carried out at Block Rock, close to Bock Cay, (396 ± 7 (sd, N = 1461) µatm). Furthermore, the range of daily variations is one order of magnitude lower at Block Rock, as indicated by the standard deviation values. These differences are most probably related to the absence of mangrove forest on Bock Cay and the fact that this cycle was carried out in the open waters of the Bahamas Bank.

3.2. Air-Water CO₂ Fluxes

[10] The air-water CO_2 fluxes (F) were computed according to the equation $F = \alpha . k . \Delta p CO_2$, where α is the solubility coefficient of CO_2 , k is the gas transfer velocity of CO_2 and ΔpCO_2 is the air-water gradient of pCO_2 . We used the relationship of k as a function of wind speed given by Carini et al. [1996], based on a SF₆ release experiment in the Parker River estuary. This formulation was preferred to those used for open oceanic waters because mangrove systems are relatively similar to estuary ones from a physical point of view (shallow and relatively sheltered). Note that the *Carini et al.* [1996] relationship gives lower k values when compared to other relationships developed for estuaries [see Raymond and Cole, 2001], probably because of strong fetch-limitation and low tidal currents in the Parker River estuary, so the fluxes we computed are minimal estimates. The strongest atmospheric CO₂ fluxes were computed for Gaderu Creek and the lowest for Norman's Pond, in accordance with the ΔpCO_2 , in spite of relatively different wind speed values at each site (Table 2). The error on the flux computations taking into account the estimated uncertainty on ΔpCO_2 of $\pm 5 \mu atm$ yields $\pm 4\%$ for Norman's Pond, $\pm 1\%$ for Nagada Creek, $\pm 0.3\%$ for Gaderu Creek, increasing with decreasing values of ΔpCO_2 . If an uncertainty of $\pm 5\%$ on the wind speed measurements is added, then the error on the flux computations becomes, on average, $\pm 10\%$ for Norman's Pond, $\pm 5\%$ for Nagada Creek, $\pm 2\%$ for Gaderu Creek. A further bias in the CO₂ flux computations could arise from the representativeness of the immediate wind measurements for each site. However, it is not possible to evaluate this source of uncertainty because of the absence of meteorological stations at the studied sites. Finally, the fluxes computed with the Carini et al. [1996] relationship are on average, for the three sites, 25% lower than those computed with the Raymond and Cole [2001]

Table 2. Average \pm Standard Deviation of Wind Speed (m s⁻¹), CO₂ Gas Transfer Velocity (k in cm h⁻¹) as Formulated by *Carini et al.* [1996], the pCO₂ Air-Water Gradient (Δ pCO₂ in μ atm) and the Atmospheric CO₂ Flux (mmolC m⁻² day⁻¹), for the Three Study Sites Plus Published Data in Four Other Sites

Site	wind speed	k	ΔpCO_2	Atmospheric CO ₂ flux	refs
Nagada Creek (Papua New Guinea $\sim 5^{\circ}S$ 146'E)	3.0 ± 2.1	8 ± 6	799 ± 357	43.6 ± 33.2	1
Gaderu Creek (India $\sim 17^{\circ}N 82^{\circ}E$)	1.4 ± 1.9	4 ± 5	2215 ± 864	56.0 ± 100.9	1
Norman's Pond (Bahamas $\sim 23^{\circ}N^{\circ}76^{\circ}W$)	5.5 ± 1.3	13 ± 3	165 ± 86	13.8 ± 8.3	1
Saptamukhi Creek (India $\sim 22^{\circ}N 89'E$)	nd	4 ^a	2210 ± 1460	56.7 ± 37.4	2
Mooringanga Creek (India $\sim 22^{\circ}N 89'E$)	nd	4 ^a	830 ± 363	23.2 ± 10.1	2
Itacuraçá Creek (Brazil $\sim 23^{\circ}\text{S} 44'\text{W}$)	nd	4 ^a	3845 ± 3520	113.5 ± 104.4	3
Florida Bay (U.S.A. $\sim 25^{\circ}N 81'W$)	nd	4 ^a	170 ± 200	4.6 ± 5.4	4

^aTaken as constant (see text). nd. no data. 1. This study. 2. *Ghosh et al.* [1987]. 3. Based on the pH and TAlk data from *Ovalle et al.* [1990] and computed using the thermodynamic constants described by *Frankignoulle and Borges* [2001]. 4. *Millero et al.* [2001].

relationship, that is based on a compilation of *k* values from various estuaries and using different methodologies.

[11] All the pCO₂ data available, to our best knowledge, in literature for the waters surrounding mangrove forests are summarised in Table 2. As for our three study sites, CO₂ oversaturation is observed in other mangrove waters, although highly variable from one site to another. Since wind speed is not reported in these publications, the corresponding air-water CO₂ fluxes were computed using a constant *k* value of 4 cm h⁻¹, that is in the lower range of the *k* values from our study sites (Table 2). The choice of this low value is based on the fact that the study sites are, except Florida Bay, narrow creeks that seem relatively well sheltered as Gaderu Creek. The range of the CO₂ fluxes is high but in agreement with the one computed for our three study sites.

4. Synthesis

[12] Based on the above CO_2 data, we can speculate that oversaturation of CO₂ with respect to atmospheric equilibrium is the general feature in the waters surrounding mangrove forests. Although information on the seasonal variability of CO₂ fluxes in mangrove waters is missing, we feel that CO₂ flux data in 7 mangrove ecosystems, covering both subtropical and tropical latitudes, allow us to attempt a preliminary extrapolation and global integration. The computed air-water CO₂ fluxes converge to a value of about 50 mmolC m⁻² day⁻¹, although the range of variability encompasses one order of magnitude. If we extrapolate this conservative value to the surface area of worldwide mangrove ecosystems ($\sim 0.2 \ 10^6 \ \text{km}^2$) the global emission of CO_2 to the atmosphere would be about 50 10⁶ tC year⁻¹. This emission is comparable to the global estimation of leaf litter remineralized within mangroves (23 10^6 tC year⁻¹), according to a recent review of literature [Jennerjahn and Ittekkot, 2002]. It should be noted that the CO_2 flux integration was made over the surface area of the mangrove forests per se but that the surface area of the waters directly influenced by mangroves, although unknown, is probably higher. This CO₂ emission can be considered modest at a global scale since the net open oceanic CO_2 sink is globally of about 2.2 GtC year⁻¹. On a regional scale, the subtropical and tropical open oceanic waters behave as a net source of CO₂ of about 0.43 GtC year $^{-1}$ [between 32°N and 32°S, based on Takahashi et al., 1997, extracted from the internet at http://ingrid.ldeo.columbia.edu/SOURCES/ .LDEO/.Takahashi/]. Thus, mangrove surrounding waters would be an additional CO_2 source of about 12% to the one of open oceanic waters, in tropical and subtropical latitudes, with a surface area about one thousand times smaller.

[13] Further research is needed to better evaluate the role of mangrove ecosystems in the global carbon cycle. The sediments of human impacted mangroves are known to be more acidic than natural ones [*Middelburg et al.*, 1996], while, the dominant diagenetic organic carbon degradation pathway of recently reclaimed mangroves is different than

in natural (old) ones [*Alongi et al.*, 1998]. The impact of these processes on DIC dynamics of the adjacent waters and related CO_2 fluxes has not yet been investigated. Also, about 20% of the global production and about 30% of the global accumulation of CaCO₃ occur in tropical latitudes, but the role of mangrove ecosystems as a sink of CaCO₃ still remains to be determined [*Middelburg et al.*, 1996]. Finally, the open continental shelf areas adjacent to mangrove forests influenced by their organic carbon and nutrient inputs should also investigated from the point of view of atmospheric CO_2 fluxes.

[14] Acknowledgments. The authors would like to thank Nagada Diving Centre (Papua New Guinea), Caribbean Marine Research Centre (Bahamas) and Andhra University (India) for the welcome received at each site, and, numerous colleagues for help in sampling and the elaboration of the manuscript. Comments from two anonymous reviewers improved the manuscript. This work was funded by the F.N.R.S. (contracts numbers 2.4521.96, 2.4594.01, 2.4596.01, 2.4545.02), with which AVB and MF are, respectively, a postdoctoral researcher and a research associate.

References

- Alongi, D. M., A. Sasekumar, F. Tirendi, and P. Dixon, The influence of stand age on benthic decomposition and recycling of organic matter in managed mangrove forests of Malaysia, *J. Exp. Mar. Biol. Ecol.*, 225, 197–218, 1998.
- Alongi, D. M., G. Wattayakorn, J. Pfitzner, F. Tirendi, I. Zagorskis, G. J. Brunskill, A. Davidson, and B. F. Clough, Organic carbon accumulation and metabolic pathways in sediments of mangrove forests in southern Thailand, *Mar. Geol.*, 179, 85–103, 2001.
- Carini, S., N. Weston, C. Hopkinson, J. Tucker, A. Giblin, and J. Vallino, Gas exchange rates in the Parker River estuary, Massachusetts, *Biol. Bull.*, 191, 333–334, 1996.
- Frankignoulle, M., and A. V. Borges, Direct and indirect pCO₂ measurements in a wide range of pCO₂ and salinity values (The Scheldt Estuary), *Aquat. Geochem.*, 7, 267–273, 2001.
- Ghosh, S., T. K. Jana, B. N. Singh, and A. Choudhury, Comparative study of carbon dioxide system in virgin and reclaimed mangrove waters of Sundarbans during freshet, *Mahasagar Bull. Nat. Instit. Oceanogr.*, 20, 155–161, 1987.
- Jennerjahn, T. C., and V. Ittekkot, Relevance of mangroves for the production and deposition of organic matter along tropical continental margins, *Naturwissenschaften*, 89, 23–30, 2002.
- Kristensen, E., F. Ø. Andersen, N. Holmbe, M. Holmer, and N. Thongtham, Carbon and nitrogen mineralization in sediments of Bangrong mangrove area, Phuket, Thailand, *Aquat. Microb. Ecol.*, 22, 199–213, 2000.
- Middelburg, J. J., J. Nieuwenhuize, F. J. Slim, and B. Ohowa, Sediment biogeochemistry in an East African mangrove forest (Gazi Bay, Kenya), *Biogeochemistry*, 34, 133–155, 1996.
- Millero, F. J., W. T. Hiscock, F. Huang, M. Roche, and J. Z. Zhang, Seasonal variation of the carbonate system in Florida Bay, *Bull. Mar. Sci.*, 68, 101–123, 2001.
- Ovalle, A. R. C., C. E. Rezende, L. D. Lacerda, and C. A. R. Silva, Factors affecting the hydrochemistry of a mangrove creek, Sepetiba Bay, Brazil, *Estuar. Coast. Shelf S.*, 31, 639–650, 1990.
- Raymond, P. A., and J. J. Cole, Gas exchange in rivers and estuaries: Choosing a gas transfer velocity, *Estuaries*, 24, 312–317, 2001.
- Takahashi, T., R. A. Feely, R. F. Weiss, R. H. Wanninkhof, D. W. Chipman, S. C. Sutherland, and T. T. Takahashi, Global air-sea flux of CO₂: An estimate based on measurements of sea-air pCO₂ difference, *Proc. Nat. Acad. Sci. USA*, *94*, 8292–8299, 1997.

A. V. Borges, B. Delille, S. Djenidi, M. Frankignoulle, G. Lacroix, and J. Théate, Université de Liège, MARE, Intitut de Physique (B5), B-4000 Liège, Belgium. (Alberto.Borges@ulg.ac.be; Bruno.Delille@ulg.ac.be; sdjenidi@ulg.ac.be; Michel.Frankignoulle@ulg.ac.be; G.Lacroix@mumm. ac.be; jmtheate@ulg.ac.be)